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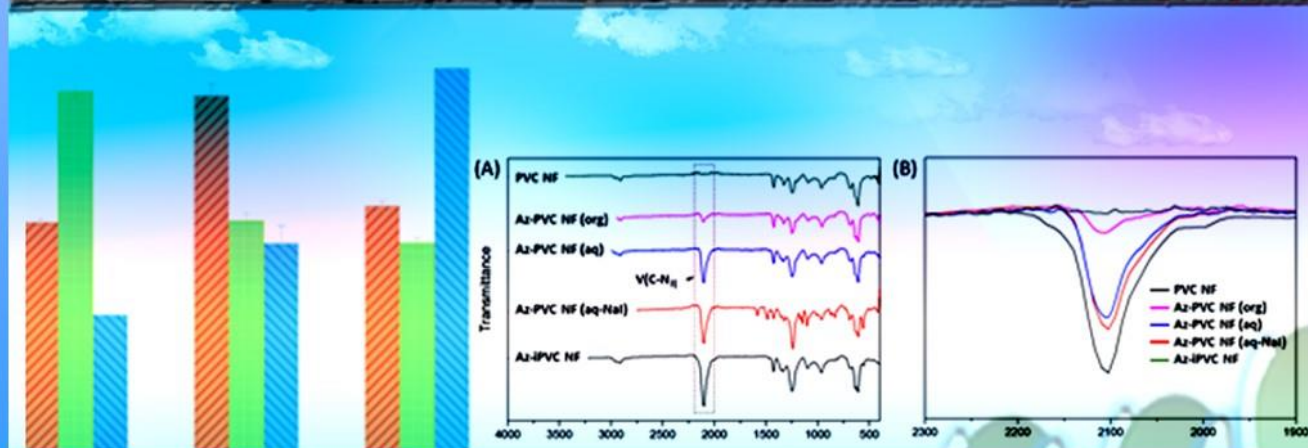
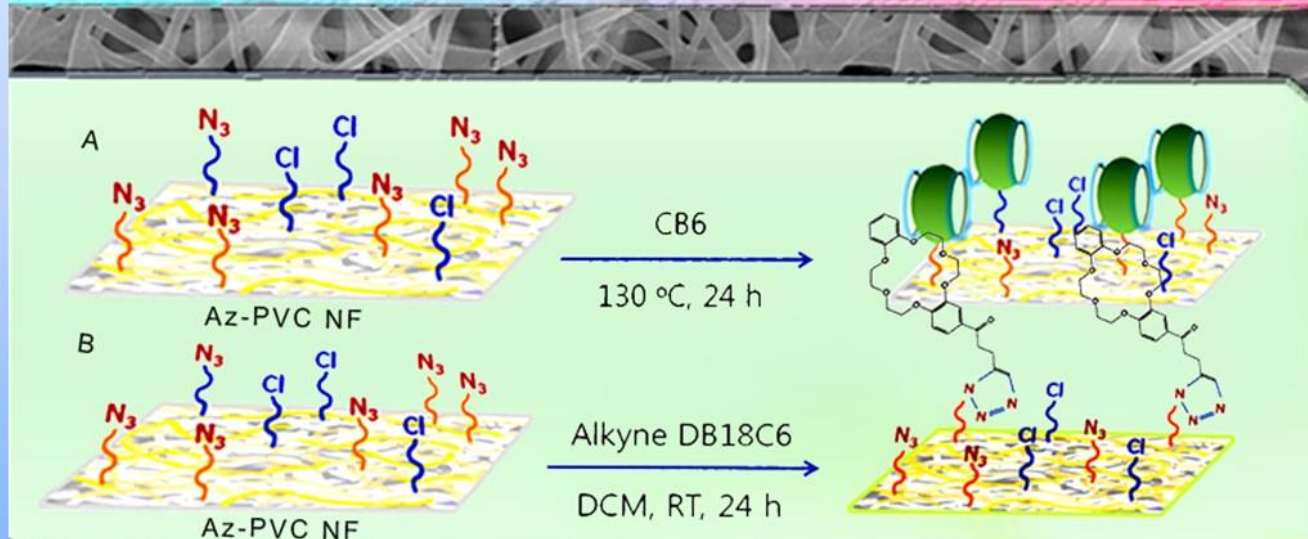
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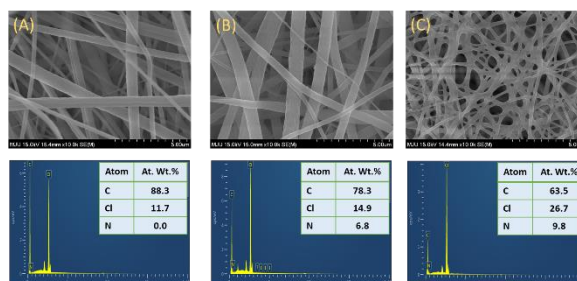
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## COVER PHOTOS

Scanning electron microscopy (SEM) images and energy dispersive X-ray (EDX) spectra of pristine PVC NF (image A), Az-PVC NF (image B) and Az-iPVC NF (image C) membranes were obtained to visualize their surface morphology and estimate their surface elemental composition, respectively. SEM images show that electrospun PVC NF features thin, long fibers with nanosized diameters. Pristine PVC NF ( $\varnothing \sim 413 \pm 195$  nm) and Az-PVC NF (aq-NaI,  $\varnothing \sim 426 \pm 201$  nm) membranes are essentially morphologically similar which means that iodination and subsequent azidation of PVC NF membrane cause no change in the physical structure of the membrane. On the other hand, having been prepared using a different set of electrospinning conditions, Az-iPVC NF ( $\varnothing \sim 187 \pm 53$  nm) features thinner, densely meshed nanofibers. Results from surface elemental analyses via EDX further confirmed the presence of azide groups in Az-PVC NF and Az-iPVC NF, which contain approximately 6.8% and 9.8% nitrogen, respectively. **Images courtesy of EC Escobar.**



**COVER DESIGN:** Jovan A. Gimarangan

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## EDITORIAL

As a multidisciplinary journal, The Palawan Scientist has been evolving across different research areas in sciences for more than a decade. As proof of its wide dissemination and comprehensive efforts in publishing relevant and timely research outputs, published papers have expanded from agriculture and fisheries to engineering technical papers. The journal gears towards partnerships seeking to relentlessly uphold and publish outputs of many experts and young researchers alike.

In this year's partnership with the University of the Philippines Los Baños (UPLB), College of Engineering and Agro-industrial Technology (CEAT), the Palawan Scientist publishes its Special Issue containing selected research outputs presented during the 2nd International Conference of Engineering and Agro-industrial Technology (iCEAT). iCEAT is an annual event organized by CEAT UPLB which aims to provide avenue for exploration, discussion, recommendations and interconnectedness among leading experts, educators, policymakers, and industries on the latest trends, issues, and solutions for the field of engineering and agro-industrial technology.

This year, the iCEAT 2023 theme "*Materials and Processes for the New Normal: Sustainability, Innovation, and Opportunities*". iCEAT envisions to work in collaboration with stakeholders for a universal call to act towards the attainment of the sustainable development goals (SDGs) towards no poverty, zero hunger, clean water and sanitation, affordable and clean energy, industry, innovation and infrastructure, sustainable cities and communities, responsible consumption and production, and climate action. Despite experiencing community disruptions during the onset of the COVID-19 pandemic, it has opened doors to technological and engineering advancements, all forward to societal future-proofing and sustainability. Innovations opened many opportunities for providing solutions to critical challenges.

The Palawan Scientist continues to encourage both experts and young researchers to experience being known through their research work published in this journal. This peer-reviewed journal serves as a platform for disseminating and sharing relevant research outputs in various fields that will contribute to developing further projects towards meeting the needs of communities especially the poor, needy and marginalized sector.

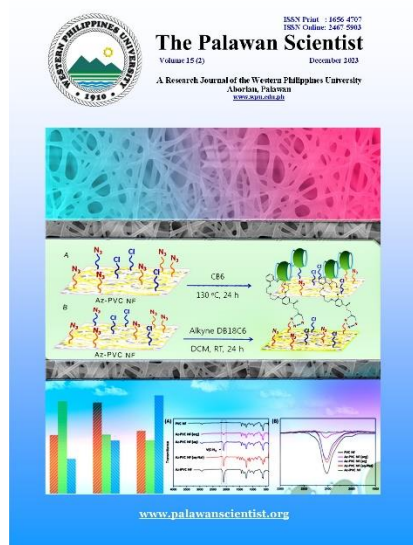
I would like to commend all authors, the iCEAT Organizing Committee members, and the Palawan Scientist Editorial Board members and staff for this year's successful collaboration. May this effective collaboration between UPLB and The Palawan Scientist continues to prosper in the years to come.

**Hector R. Lim Jr., Ph.D.**

Associate Editor-The Palawan Scientist  
President, Agape Rural Program

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# Current trends in modular wastewater treatment - a patent review

Jaren U. Tulipan and Jey-R S. Ventura\*

*Biomaterials and Environmental Engineering Laboratory, Department of Engineering Science, College of Engineering and Agro-Industrial Technology, University of the Philippines Los Baños, College, Laguna, Philippines*

\*Correspondence: [jsventura@up.edu.ph](mailto:jsventura@up.edu.ph)

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## ABSTRACT

Wastewater is a major source of water pollution, as it contains contaminants from human waste, chemicals, industrial runoff, and other sources that can be harmful to the environment and human health. Therefore, treatment is necessary to reduce its harmful effects. However, wastewater treatment (WWT) is often costly and difficult to implement in many areas. With conventional WWT designs lacking robustness and adaptability, modular designs can be a viable solution for this problem. In this review, a patent survey covering patents from 1996 to 2021 yielded 66 patents regarding modular WWT. The patent search showed increasing patents regarding modular WWT in recent years. Most of the patents were filed by companies in China. The International Patent Classification (IPC) was used to determine the current trends in modular WWT. Current patented designs suggest that there have been attempts to develop modularized electrochemical, separation, and sorption methods. Also, multistep patent designs showed fully integrated modular units from conventional WWT design and offer robustness and adaptability. This patent review showed that modular WWT technologies are capable and ready to be used in different settings.

**Keywords:** compact wastewater treatment, water pollution, wastewater treatment technologies

## INTRODUCTION

Water pollution has become a global concern due to the exponential growth of the world's population and the rapid industrialization and urbanization (Sonune and Ghate 2004). This has led to a scarcity of clean water that makes it necessary to address this issue urgently. Wastewater generated by human activities further contributes to the problem, with an estimated 80% of wastewater worldwide discharged into the environment without adequate

treatment (Dutta et al. 2021). This poses a significant challenge, especially for low-income countries where industrial wastewater contains high concentrations of metals, and domestic or municipal wastewater is laden with nutrients. In contrast, high-income countries can treat a larger proportion of their wastewater due to stricter regulations and the presence of wastewater treatment (WWT) facilities (Crini and Lichtfouse 2018; Dutta et al. 2021).

Wastewater treatment processes encompass a range of physical, biological, and chemical



techniques that have been extensively studied and applied over the years (Henze et al. 1997; Sonune and Ghate 2004; Anjaneyulu et al. 2005; Barakat 2011; Sharma et al. 2012). Current WWT processes typically involve four main stages: preliminary treatment, primary treatment, secondary treatment, and tertiary treatment. However, these conventional WWT plants suffer from notable drawbacks (Marchesseault et al. 2005; Reilly and Jelderks 2011a; Molinos-Senante et al. 2012; Jin et al. 2014). These challenges include the need for skilled operators, high operational costs, the requirement for overdesigning to accommodate future capacity increases and influent variations, the production of large amounts of waste sludge, and difficulties in adapting to varying influent volumes and contaminant concentrations.

Modular WWT systems offer a promising solution to these challenges by providing a robust, adaptable, and flexible approach to WWT, addressing many of the issues encountered in current designs (Spiller et al. 2015). These systems improve energy efficiency by optimizing the modules used for treating current influent wastewater, thereby reducing energy consumption. Additionally, their compact design minimizes land area requirements, making them suitable for small towns and industries with limited space. Modular treatment systems enable on-site WWT, extending the lifespan of existing facilities and ensuring effective treatment in specific areas. Moreover, modular designs allow for easy expansion by adding additional treatment modules to accommodate increased influent wastewater. Integrating modular WWT systems into current treatment plants can extend their operational life and ensure compliance with new effluent quality regulations.

However, it is crucial for modular systems to deliver comparable performance to conventional WWT plants. The primary objective of WWT is the removal of nutrients, with biodegradation being the preferred mechanism for breaking down these compounds (Dobson and Burgess 2007). The activated sludge process (ASP), a widely employed method, utilizes the aerobic degradation of organic compounds by supplying air to microbial communities in the reactor (Daverey et al. 2019). The ASP is considered a simple and cost-effective approach to WWT.

However, current ASP systems are energy-intensive, resulting in high operational costs (Dobson and Burgess 2007; Daverey et al. 2019). To overcome this challenge, alternative secondary treatment technologies, such as membrane bioreactors, moving bed bioreactors, and sequencing batch reactors, have been explored.

This review aims to collect and evaluate patents filed between 1996 and 2021, focusing on the current design trends and technologies used in the development of modular WWT systems. By analyzing these patents, this review serves as a foundation for future advancements in modular WWT systems that will foster sustainability and address the pressing challenges associated with WWT.

### PATENT SEARCH METHODOLOGY

Figure 1 shows the patent search methodology used for this study. A keyword search approach was used for the initial patent search using the Espacenet database of the European Patent Office. Generic keywords were used for the initial search which include “modular”, “compact”, “wastewater”, “treatment”, “plant”, “design”, “configuration”, “process”, and “method”. Only those patents relevant to modular WWT were considered, regardless of their International Patent Classification (IPC) classes. The selected patents were downloaded as a search query from the Espacenet database and compiled in an Excel spreadsheet. Two search queries were downloaded from the Espacenet database. The first search query was downloaded from the database on 02 August 2022 and contained 212 patents. The second search query, downloaded from the database on 08 August 2022 contained 54 patents. The patents were then reexamined to exclude searches that did not deal with treating wastewater and lacked modularity. From the previous search queries, only 66 patents were considered. The remaining relevant patents were then reanalyzed via the Google Patents database to find other data about the patent, i.e., patent application date, status, and countries where the patents were approved. The additional information extracted from each patent was done on 04 January 2023.

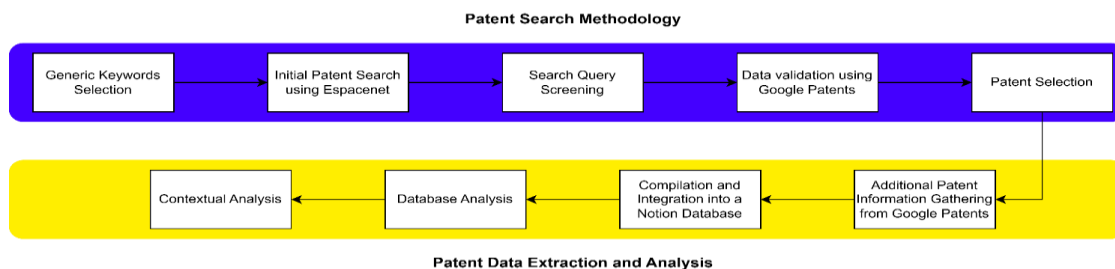


Figure 1. Patent search and data analysis methodology.

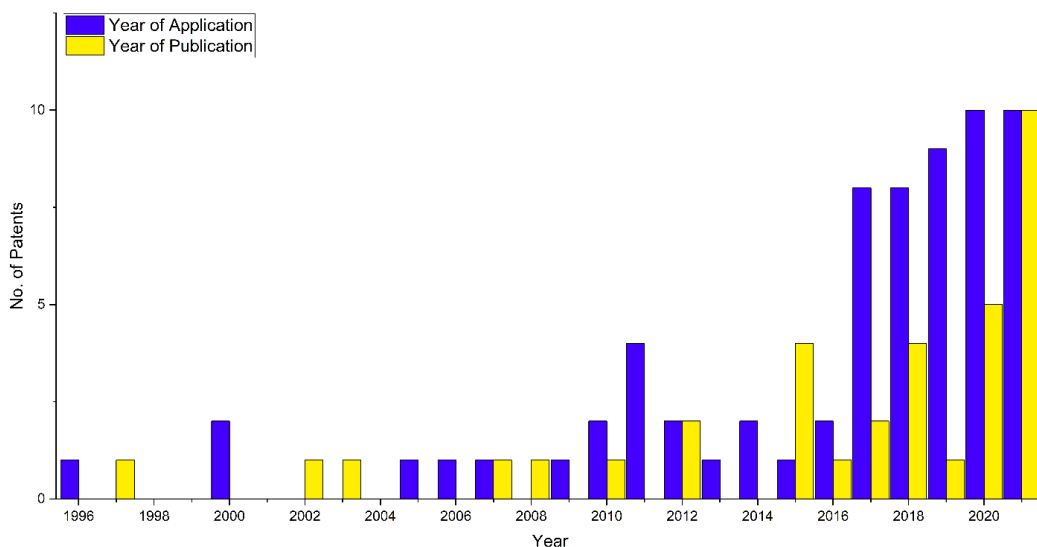


## PATENT DATA EXTRACTION AND ANALYSIS

A total of 66 patents were found from the Espacenet and Google Patents databases that describe a design or a method of implementing a modular WWT technology. These patents were integrated into a Notion database (<https://tinyurl.com/patentswwt>). The exported information from each patent includes the patent number, patent name, application date, author/inventor, claims, figures, country of application, patent status, and IPC classes. The considered patents for the database are the applied patents only. However, during data gathering, the subcategories of the filed patents were recorded. The Notion database was used to analyze the patents and filter the patents according to different types of information exported. The filtered patents are used to determine the patent distribution with regard to the date of filing, country, and IPC classes. The claims of each patent were analyzed to give context to its design, validate its IPC classes, and further generalize the current design trends and application of WWT technology.

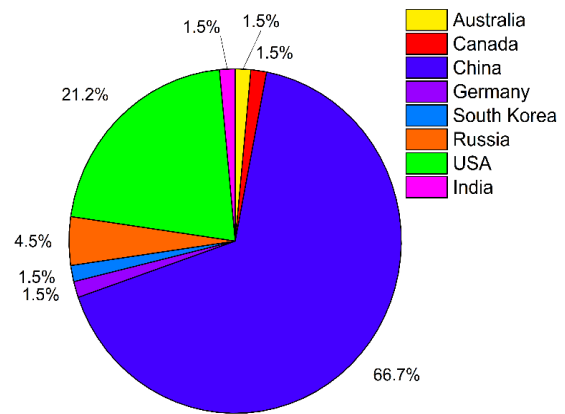
## PATENT SITUATION FOR THE MODULAR WWT PLANT

As shown in Figure 2, the majority of the patents are from China, with 44 patents, followed by the USA with 14, Russia with three, and Australia, Canada, India, Germany, and South Korea with one patent each. Figure 3 shows the distribution of patents based on the year of application and the year of publication.



**Figure 3.** Distribution of patents based on the year of application and year of publication.

publication. It shows that the number of patent applications increased significantly from 2017 to 2021, with a peak of 10 patent applications in 2020 and 2021. During this period, 66% of the relevant patents for modular WWT were filed. Most of the patents from 1996 to 2016 were filed in the USA. Meanwhile, 38 out of the 45 patents filed from 2017 to 2021 were filed in China.



**Figure 2.** Distribution of patents based on the countries where they were filed.

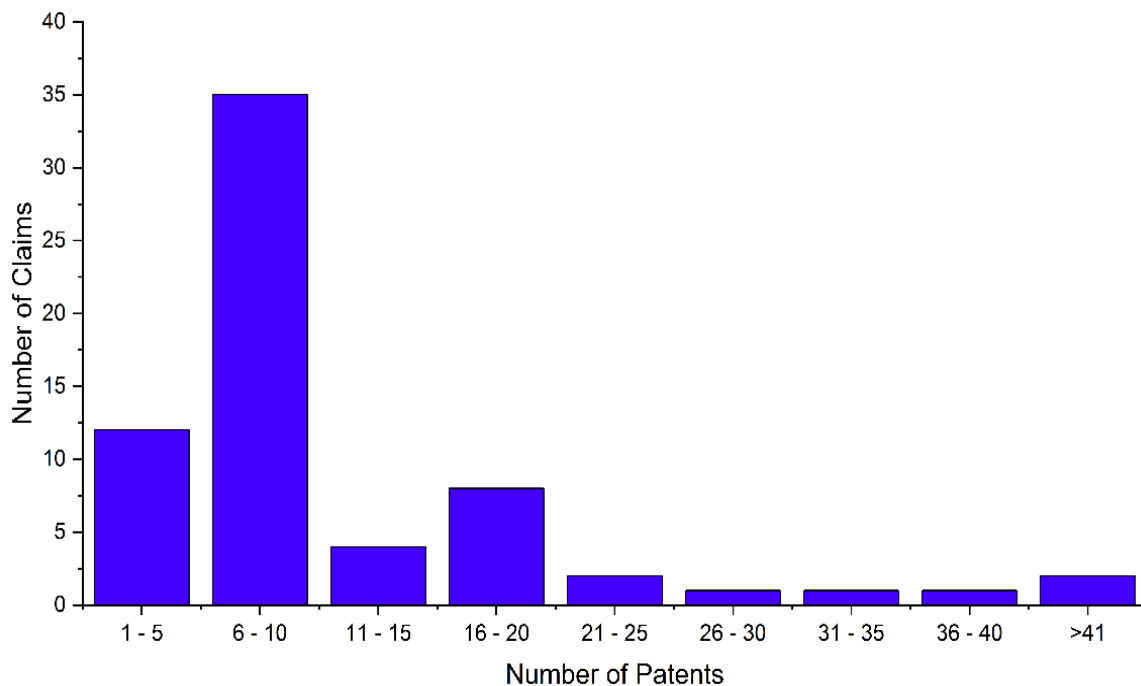
It was seen that patents filed from 1996 to 2016 experienced slower publication, with an average of 902 days between the date of application and the date of publication. However, this has not been the case recently. Patents from 2017 to 2021 are generally granted faster, with an average of 340 days between the date of application and the date of publication.

Figure 4 shows the distribution of patents based on the number of claims. Thirty-five out of the 66 patents considered in this study have between 6 and 10 claims. This shows that there has been a huge advancement in modular WWT technology. The presence of two patents with more than 41 claims suggests the uniqueness and complexity of their modular system. Furthermore, it is important to consider the implications of claim distribution in the context of intellectual property management and litigation. Patents with a large number of claims are more robust in defending against infringement as they cover a wider scope of potential violations. Conversely, patents with fewer claims could also be more concise and easier to manage.

Figure 5 shows the distribution of patents based on their application status. Currently, there are 34 active patents with modular WWT designs, of which 23 are from China. There is one withdrawn patent (CN108178461A) and two abandoned patents (US2005274669A1 and US20070289922A1). There are 10 expired patents, of which seven are classified as expired due to fee-related issues. As of this writing, there are 19 pending patent applications. Seventeen of them are from China, and two are from the USA.

Figure 6 shows the distribution of patents based on the nature of their current assignee. Most of the relevant patents have companies as their current

assignees. Meanwhile, eight patents have universities as their current assignees, and eight patents have no assignee. The presence of multiple companies as patent holders can indicate several key points related to commercialization, technological advancements, and competitiveness within specific industries. Firstly, when numerous companies hold patents, it typically signifies a strong inclination toward commercializing inventions (Mazzoleni and Nelson 1998). Companies, driven by profit motives and the need to stay competitive in the market, tend to direct their innovations towards practical applications that can generate revenue. Also, companies often possess more substantial resources compared to universities, including robust research budgets, specialized facilities, and a skilled workforce (Lockett and Wright 2005). Consequently, a higher prevalence of company-held patents can indicate a faster pace of technological advancement in particular industries (Garavito and Rueda 2021). A greater number of companies holding patents can also foster a competitive ecosystem (Encaoua et al. 2006; Arora et al. 2008). Companies may engage in competition to innovate, which, in turn, promotes continuous enhancements in their products and services. This competitive environment can be advantageous for consumers by offering them a wider array of choices and potentially driving down prices.



**Figure 4.** Distribution of patents based on the number of claims.

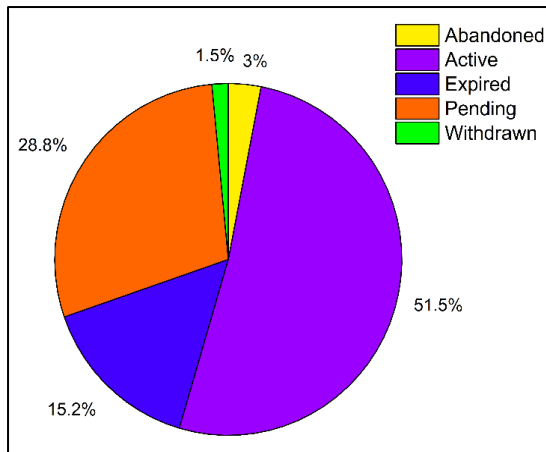


Figure 5. Distribution of patents based on application status.

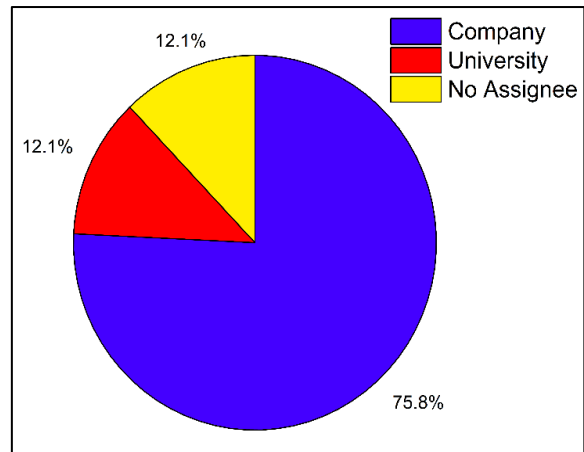


Figure 6. Distribution of patents based on current assignee.

### CURRENT DESIGN TRENDS FOR MODULARITY OF WWT PLANT BASED ON THE IPC OF THE PATENTS

For this patent review, the classification considered for the patents is the IPC. This classification features a systematic and uniform means of classifying patents in terms of the area of technology of the patents. It uses a system of letters and numbers to form IPC symbols with class level, subclass level, group level, and subgroup level integrated into a single IPC symbol. Using this classification, it is easier to evaluate the current design trends for modular WWT based on filed patents.

Table 1 summarizes the IPC of the patents considered for this study. In the table, there are seven subclasses under which the 66 relevant patents were classified. This shows that there are numerous

attempts at modularization of different technologies for WWT. The subclasses were focused on the separation process (B01D, B65D), treatment of water, wastewater, and sewage (C02F), containers and fixed constructions (E03B, E03F), and process control systems (G05B, G05D). Since the relevant patents pertain to modular WWT systems, groups under the C02F subclass were examined. Figure 7 shows the distribution of patents based on the IPC.

Besides the C02F subclass, the highest number of patents classified under a subclass is five. This subclass is B01D, which covers the separation of solids via physical processes such as filtration and sedimentation. Another subgroup involving separation methods is C02F9/02 where US2005274669A1 is classified. Also, the E03F5/14 subgroup covers the separation of solid or liquid substances from sewage.

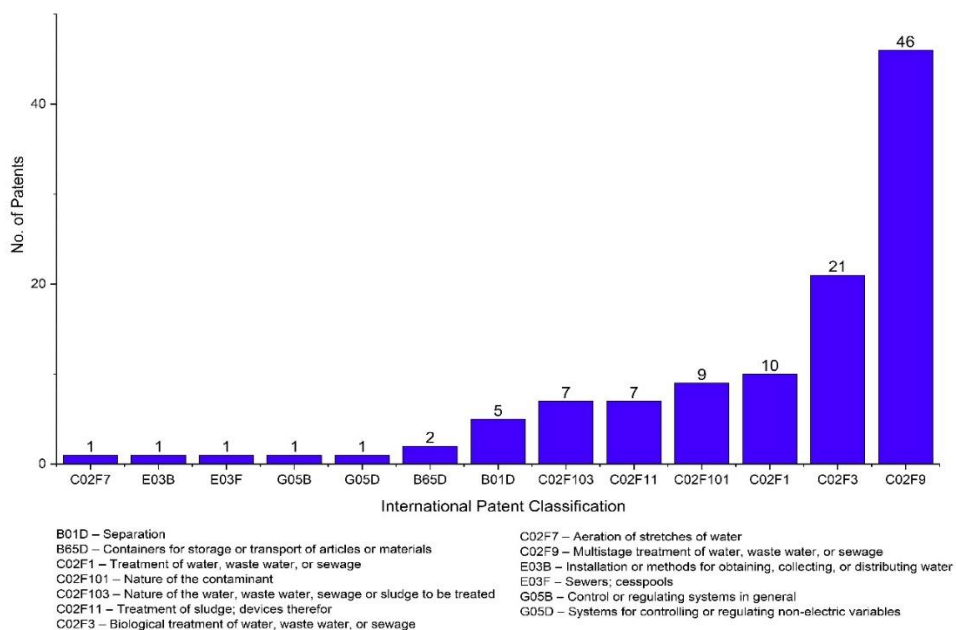


Figure 7. Distribution of patents based on IPC.

This implies that modular primary treatments are already designed and feasible for application. Further implications are the C02F9/08, C02F9/10, and C02F9/12 subgroups, where a total of five patents are classified under these classifications. These subgroups cover multistep WWT involving physical treatment. However, since physical treatment is a rather simpler treatment method than biological and chemical, the small number of patents classified under relevant classifications involving a physical treatment method shows that the modularity of physical treatments has not yet been fully explored.

The B65D and E03B subclasses imply a compact modular WWT design that is flexible to accommodate any wastewater-influent variations. B65D covers the storage and transport of materials. The IPC subgroups under this subclass under which two patents are classified are B65D21/024, B65D21/028, B65D88/010, B65D88/74, B65D90/02, and B65D90/10. The B65D21 group implies a flexible WWT design where modules are made of nestable, stackable, or joinable containers. Only WO2008006175A1 utilized such a design. The E03B subclass further aids the flexibility of WWT design because it considers the arrangement of treatment modules and methods of installation of the modules.

Another approach to having a flexible WWT system is the design disclosed in US2012055883A1. The patent disclosed a system of treatment modules that can go offline or online depending on the treatment needs of the influent wastewater. The patent is classified under the G05B subclass, which covers control systems regulating component variables. Meanwhile, US9079125B2 disclosed a design that has a flow level sensor that adjusts the influent volume dynamically so the treatment system can properly accommodate the influent wastewater. This patent is under the G05D subclass, which covers control systems for non-electrical variables.

From the relevant patents examined, there are already three approaches patented for a flexible and adaptable WWT system. The first is to have a design where modules are stackable and joinable. This signifies a modular system that can increase or decrease its treatment capacity by adding or removing treatment modules. The second approach is to have a control system that can dynamically control which treatment modules should be activated or deactivated to accommodate the varying wastewater influent on the system. The third approach, which is common to non-modular design, is to regulate the volume flow of influent wastewater to prevent overloading the system.

The C02F subclass deals with the treatment of water, wastewater, and sewage. Under this subclass, seven groups were identified by the IPCs of the relevant patents considered for this review. The C02F101 subclass covers the identity of contaminants in water, wastewater, and sewage. The patents showed that the patented modular WWT systems can treat

different contaminants, such as heavy metals, inorganic compounds, nitrogen compounds, and organic compounds. This improves the applicability of a modular WWT system to any industry, small business, or municipality.

Common WWTs identified through the C02F1 group are sorption (C02F1/28), UV-light irradiation (C02F1/32), reverse osmosis (C02F1/44), electrochemical methods (C02F1/46), and oxidation with ozone (C02F1/78). Among the 66 patents considered in this review, only 10 are classified under the C02F1 group, as shown in Figure 5. This means that although different treatment methods are explored for modular systems, these designs are not mature and can be further developed. However, this shows that the modularity of WWT methods is feasible, and that further development of modular designs could lead to a higher number of patents being classified under this group. This also shows that different methods can be used as a treatment module for WWT. This means that a modular WWT design has a variety of options to treat a target characteristic of the influent wastewater.

The relevant patents also showed that their modular WWT system is capable of biological treatment. There are 21 patents classified under the C02F3 group and 27 patents under the C02F9/14 subgroup. The biological treatment used in modular designs can be categorized into four – aerobic processes, activated sludge processes, anaerobic digestion, and membrane bioreactors.

There are four patents using the biological treatment of wastewater through aerobic processes, which are classified under C02F3/02 and C02F3/06. The aerobic process disclosed by US9890067 utilized a column design of submerged filters to treat wastewater. Meanwhile, there are eight patents using aerobic and anaerobic processes in their multistep treatment design, as classified under C02F9/14. Designs include modular tanks, ponds, and treatment modules.

Meanwhile, there are 11 patents using the activated sludge process under the C02F3 group. Most of the designs use the conventional activated sludge process. Four patents under C02F9/14 used activated sludge in various integrations in multistep systems. There are also nine patents using anaerobic digestion for biological treatment. The patents generally use anaerobic digestion to have a complete nitrification and denitrification process.

Under the C02F3 group, CN108117150A and CN208055010U utilized membrane bioreactor modules. Furthermore, seven patents under C02F9/14 used membrane technology for their biological treatment in their multistep treatment systems.

The classification of the patents under C02F3 and C02F9/14 shows that even advanced biological treatments can have a modular design, which gives a glimpse of the high treatment efficiency of modular

WWT systems. This shows that biological treatment is now well integrated into modular systems.

Patented modular WWT systems also have sludge treatment modules. There are seven patents that are classified under the C02F11 group. The treatment methods for sludge disclosed by these patents include biological treatment, anaerobic treatment, dewatering and thickening, oxidation, and adding chemical agents. The small number of patents integrating sludge treatment modules into their treatment systems suggests that sludge treatment is not the primary concern of such modular systems. However, the seven patents showed that sludge treatment modules are feasible to integrate into modular systems, making a more complete system.

Among the groups under the C02F subclass, the C02F9 group has the highest number of identified patents. This group covers the multistep treatment of wastewater. The identified subgroups that the patents are classified under involve separation steps (C02F9/02), chemical treatment steps (C02F9/04), electrochemical treatment steps (C02F9/06), physical treatment steps (C02F9/08, C02F9/10, C02F9/12), and biological treatment steps (C02F9/14). As shown in Figure 8, the number of patents being classified under

C02F9 is increasing from 2015 to 2021. The high number of patents and the annual increase in the number of patents classified under this group offer two insights. The first is that, due to more complex wastewater influents that need to be treated, a move toward designing a multistep modular treatment system is crucial to preventing untreated wastewater effluent. The second is that, due to the development of different treatment modules through the years, there are now enough designs and methods to integrate different treatment modules into a single modular treatment system. Designing a multistep modular WWT system also offers robustness in terms of treatment efficiency. Multistep treatment offers a higher possibility of removing any unexpected contaminants the influent wastewater may contain.

The current patented modular WWT systems showed the following insights on the treatment system design.

1. Primary treatment methods and secondary biological treatment methods are well integrated into modular systems. This integration ensures that the initial stages of WWT efficiently remove solid materials and pollutants.

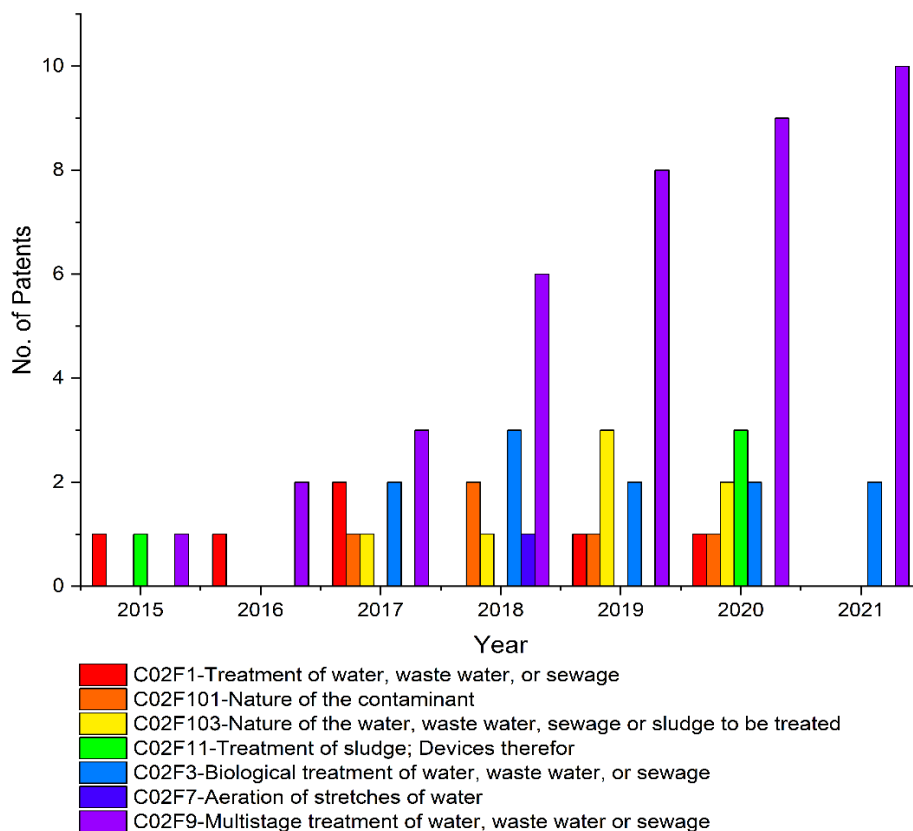


Figure 8. Distribution of patents from 2015-2021 based on the C02F subclass.

**Table 1.** Current design trends based on International Patent Classification (IPC).

IPC	IPC Title	IPC Description	Sample Patents (Source)	Modular Design Features of the Patents	Source
B01D	Separation of solids	Physical processes such as filtration, sedimentation, crystallization, precipitation, absorption, and adsorption	US9079125B2 (Reilly and Jelderks 2011b), US9890067 (Monsrreal and Rogelio 2012) KR102069161B1 (Hyun and Choi 2019) CN109381898A (Wang 2017)	Treatment basins  removable screens  settling baffles  filtration via loose material	(WIPO, 2023a)
B65D	Containers for storage and transport	Covers nestable, stackable, or joinable containers	WO2008006175A1 (Wallis 2007) KR102069161B1 (Hyun and Choi 2019)	Rearrangeable tanks with flexible applicability  Tank design for wastewater treatment made of reinforced fiber glass	(WIPO, 2023b)
C02F1	Treatment methods for water and wastewater	Covers various methods such as flotation, sorption, irradiation, dialysis, and reverse osmosis	US20160115062 (Krieger 2015) CN102050539A (Zhu et al. 2010) US8318008B1 (Anderson 2009) US9890067 (Monsrreal and Rogelio 2012) CN111302478A (Zhang et al. 2020) CN106966467A (Yang and Li 2017) CN105776442A (Yu et al. 2016)	Flotation system  Treatment using activated charcoal  Ultraviolet sterilizer  Nanofiltration reverse osmosis unit  Electrolysis and electrochemical separation  Modular electrochemical WWT  Ozone disinfection	(WIPO, 2023c)
C02F101	Nature of contaminants in water and wastewater	Covers organic and heavy metal compounds	CN108083579A (Yu et al. 2017), CN113880305A (Luo 2021), CN108911301A (Ren et al. 2018) CN113415959 (Liu et al. 2021), CN113880305A (Luo 2021),	Removal of organic compounds  Removal of heavy metals	(WIPO, 2023c)

IPC	IPC Title	IPC Description	Sample Patents (Source)	Modular Design Features of the Patents	Source
			CN215365331U (Hui et al. 2021)  CN111573999A (Tong et al. 2010), CN111646648A (Yan et al. 2020), CN110127963A (Xu et al. 2019)	Removal of inorganic, organic, and nitrogen compounds	
C02F103	Nature and origin of wastewater	Covers wastewater from different sources such as mining activities, oilfields, and food industries	CN113415959 (Liu et al. 2021)  CN111792764A (Zhang et al. 2020)  CN212504336U (Zhang et al. 2020)  CN108083579A (Yu et al. 2017)	Treatment of sludge and wastewater  Use of trough and subtroughs for connections  Treatment of acid mine wastewater  Treatment of wastewater from animal husbandry	(WIPO, 2023c)
C02F11	Devices for treating sludge	Covers various treatment methods such as biological, anaerobic, and addition of chemical agents	US6379545B1 (Perslow et al. 2000)  US2015014245A1 (Rogelio and Monsrreal 2012)  CN111646648A (Yan et al. 2020)  RU2741566C1 (Sargin and Vinichenko 2020)	Integrated sludge digester  Anaerobic treatment  Screw filters for mechanical dewatering  Use of chemical agents	(WIPO, 2023c)
C02F3	Biological treatment of water and wastewater	Covers various treatment methods such as aerobic and anaerobic processes using activated sludge	WO2008006175A1 (Wallis 2007)  US5688400A (Baxter 1996)  US9079125B2 (Reilly and Jelderks 2011b)  CN208055010U (Tang and Xu 2018)	Biological filtration  Aeration chamber  Activated sludge process  MBR technology	(WIPO, 2023c)
C02F7	Aeration of stretches of water	Covers aeration of water for WWT	CN108178461A (Hui 2018)	Modular high-efficient film device used to treat and aerate wastewater	(WIPO, 2023c)
C02F9	Multistep treatment of water,	Covers various methods such as chemical, electrochemical,	US20160115062 (Krieger 2015))	Mobile WWT facility with flotation, filtration, chemical, sludge	(WIPO, 2023c)

IPC	IPC Title	IPC Description	Sample Patents (Source)	Modular Design Features of the Patents	Source
	wastewater, and sewage	physical, thermal, and biological processes	US2021246058A1 (McFadden 2021)  US2005274669A1 (Marchessault et al., 2005)	treatment modules, and desalination system  Two-step treatment of landfill condensate using an aerobic reactor and an anaerobic reactor  Modular transportable WWT system with bioreactor, anoxic tank, membrane filter, and UV disinfection unit	
E03B	Installations used for obtaining, collecting, or distributing water	Covers methods related to water collection	WO2008006175A1 (Wallis 2007)	Water treatment tank design, collection facility	(WIPO, 2023d)
E03F	Fixed constructions of sewers and cesspools	Covers devices for separating liquid or solid substances from sewage	RU2727420C1 (Ermachenko and Bukhantsov 2019)	Modular block unit for primary treatment of wastewater, equipped with device for collecting oils, fats and grease, semi-permissible baffle for retention of floating substances	(WIPO, 2023d)
G05B	Control systems or control elements used in regulating specific variables	Covers control systems for regulating specific variables	US2012055883A1 (Reilly and Jelderks 2011a)	Program control system for monitoring WWT system	(WIPO, 2023e)
G05D	Control systems for regulating non-electrical variables	Covers control systems for regulating non-electrical variables	US9079125B2 (Reilly and Jelderks 2011b)	WWT design equipped with a sensor for detecting wastewater input flow	(WIPO, 2023e)

2. There are already treatment modules that can be integrated into current WWT facilities to upgrade the treatment capabilities and accommodate a higher influent volume. This integration enables the upgrade of treatment capabilities, allowing these facilities to accommodate higher influent volumes without the need for extensive overhauls or expansions.

3. Current trends in modular wastewater systems emphasize multistep treatment processes. This approach results in robust systems capable of producing effluents with wastewater characteristics that meet or even exceed regulatory standards. It ensures a higher level of treatment effectiveness.

4. Patented systems focus on flexibility and adaptability. By adding or regulating treatment



modules, these systems can easily adapt to changing conditions, influent variations, and capacity demands. This adaptability minimizes the need for overdesign, reducing operational costs.

5. Integration of process control systems into current modular WWT design is crucial to achieving an easy-to-operate WWT system. This integration streamlines system operation, making it more user-friendly. Skilled operators become less of a necessity, thus, lowering operational costs.

The current trend observed in this patent review does reflect the existing WWT challenges and problems to some extent. Modular WWT systems, as seen from the IPC of the patents, offer a promising solution to these challenges. The systems offer energy and space efficiency, on-site treatment, scalability, and regulatory compliance. These systems are gaining attention in patents and industry discussions due to several advantages they bring, which align with the ongoing issues in WWT.

## CONCLUSION

In recent years, there has been a rise in patent applications for modular WWT systems. This is attributed to the need for many communities and industries to comply with stricter effluent standards. Developed countries such as China and the USA are leading the patent applications for modular WWT systems. Most of the patents have companies as their current assignees. The current patent situation for modular WWT systems shows that there is rapid development in the system approach and design.

The relevant patents are analyzed based on their IPCs. The classifications showed that many different treatment methods have been modularized, such as electrochemical methods, separation methods, and sorption. The modular designs of these systems focus on removing contaminants such as nitrogen compounds, organic compounds, inorganic compounds, and heavy metals. Current designs have fully integrated biological treatments such as aerobic processes, activated sludge processes, and anaerobic digestion. Some of them incorporated sludge treatment via dewatering, chemical agents, biological treatment, and oxidation.

The current trend of modular design offers diverse ways to have an adaptable and flexible design. The current design trend also involves multistep treatment of wastewater, which offers robustness to the system. The integration of control systems as well as the further development of treatment modules will be factors in how the modular WWT system is going to be used in the future.

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## ETHICAL CONSIDERATIONS

The conduct of this review did not use/involve any animal or human as subject of the study.

## DECLARATION OF COMPETING INTEREST

The authors declare that there are no competing interests for any authors.

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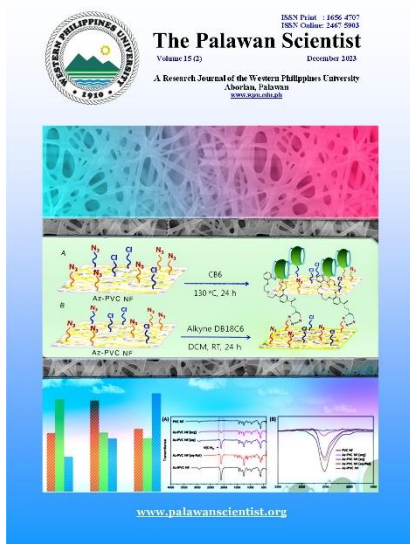
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# Facile preparation of azido functionalized polyvinyl chloride (PVC) nanofiber (NF) membrane for various applications

Erwin C. Escobar<sup>1\*</sup>, Grace M. Nisola<sup>2</sup> and Wook-Jin Chung<sup>2</sup>

<sup>1</sup>Department of Engineering Science, College of Engineering and Agro-Industrial Technology, University of the Philippines Los Baños, College Laguna, Philippines

<sup>2</sup>Environmental Waste Recycle Institute (EWRI), Department of Energy Science and Technology (DEST), Myongji University, Myongji-ro 116, Cheoin-gu, Yongin-si, Gyeonggi-do, South Korea 17058

\*Correspondence: [ecescobar1@up.edu.ph](mailto:ecescobar1@up.edu.ph)

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## ABSTRACT

Polyvinyl chloride (PVC) nanofiber (NF) membrane functionalized with azido groups (Az-PVC NF) was prepared using various techniques with the aim of enhancing azide substitution of chlorine in the PVC backbone. Azidation strategies that were investigated include: (1) catalyst-assisted azidation in organic solvent, (2) phase transfer catalyzed azidation in aqueous media, (3) phase transfer catalyzed azidation in aqueous media using pre-iodinated membrane, and (4) phase transfer catalyzed azidation in aqueous media of electrospun pre-iodinated PVC granules prepared via Finkelstein reaction. Results showed that azidation was greatly enhanced when PVC NF membrane or PVC powder underwent pre-iodination prior to azidation. As far as it is known, this is the first time that preparation techniques for azidized PVC NF membrane in aqueous media that involve halogen exchange have been studied. Hence, the optimized conditions for the preparation of azidized PVC NF membranes and the resulting membrane properties are hereby reported. Furthermore, the utility of Az-PVC NF in the development of water treatment technologies was demonstrated with successful grafting of cucurbiturils via nitrene insertion.

**Keywords:** azide, covalent attachment, halogen exchange

## INTRODUCTION

Polyvinyl chloride (PVC) is among the most widely produced and among the cheapest plastic materials (Pham et al. 2021). It has been used in a great variety of industrial, construction, and household products, and its applications continue to grow as new material processing technologies emerge. For example, with the development of electrospinning techniques for the production of nanofibers (NF) and nanofiber mats, PVC is now being used as membrane in air and water filtration systems, as protective layer in textiles, as anti-corrosion material, and many others (Ouerghui et al. 2016; Pham et al. 2021). These nanofiber-based applications of PVC take advantage of new or

improved properties such as increased porosity and increased surface area resulting from the fabrication process.

Meanwhile, the use of PVC NF membrane in the development of air or water treatment technologies may be expanded by allowing attachment of molecules with desired functionalities. A convenient way to carry out this functionalization is to introduce azido groups on the PVC chain by substitution of chlorine. Thereafter, other molecules can be covalently tethered to PVC NF via nitrene or click chemistry. As such, azidized PVC NF (Az-PVC NF) membrane can be customized to possess features useful for functions other than filtration. In other words, azidation opens



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PVC NF membrane to a multitude of modifications and applications.

Successful azidation of PVC NF membrane, however, is difficult to accomplish through conventional methods such as the use of organic solvents because these cause the fiber to deteriorate or they entail an exceedingly long time for substantial replacement of chlorine to take place. Process economics and occupational safety considerations require that azidation must be as pervasive and as swift as possible. Thus, strategies for the effective and efficient preparation of Az-PVC NF membrane need to be developed.

In this study, an assortment of techniques for the preparation of Az-PVC NF membrane were employed with the aim of achieving as much replacement of chlorine by azido groups as possible. It was found that remarkable improvements in azidation can be achieved using aqueous systems in the presence of a phase transfer catalyst, and that further improvements occur when prior replacement of chlorine by iodine is performed. This is the first report on the efficient and effective preparation of azidized PVC NF membrane in aqueous systems employing halogenation pretreatment. To demonstrate the utility of Az-PVC NF membrane in the development of water treatment technologies, it was grafted via nitrene insertion with cucurbit[6]uril or CB6—a pumpkin-shaped molecule capable of host-guest interactions with organic molecules and selective sequestration of large metal ions such as Cs<sup>+</sup> on two of its carbonyl rims.

## METHODS

### Materials

Polyvinyl chloride (PVC, average Mw~ 62000, average Mn~ 35 000) and cucurbit[6]uril hydrate were purchased from Sigma-Aldrich. Sodium azide (NaN<sub>3</sub>, 99%, extra pure) and tetrabutylammonium bromide (TBAB, 99+%) were purchased from ACROS Organics, South Korea. Sodium iodide (NaI, 99+%) was manufactured by Showa Chemical Co. Ltd. (Japan). Solvents used include dimethylformamide (DMF, anhydrous, Fisher Scientific U.K. Ltd.), acetone (99.8%, DaeJung Chemicals and Metals, Korea), ethanol (99.9%, Samchun Chemicals, South Korea) and tetrahydrofuran (THF, 99.9%, Samchun Chemicals, South Korea). Deionized water (18.2 mΩ·cm<sup>-1</sup> at 25°C) was processed through a Millipore Milli-Q system. Reagents and solvents were used as received.

### Preparation of PVC NF Membrane from pristine PVC Granules

An electrospinning solution was prepared by dissolving PVC granules (15 wt. %) in a 1:2 volume ratio of THF to DMF. Polyvinyl chloride nanofiber

membrane was produced by electrospinning this solution at 22 kV at an ejection rate of 4 ml h<sup>-1</sup> through a 21G needle and a rotating drum collector speed of 230 rpm at a distance of 100 mm.

### Preparation of Iodinated PVC Granules

Iodinated PVC granules were prepared via Finkelstein reaction using sodium iodide as source of halogen nucleophile. Polyvinyl chloride granules (30 g) and NaI (71 g) were mixed in 215 ml acetone for 23 h. The iodinated granules were collected by filtration and dried overnight in an air-circulating oven at 65°C. Iodinated granules were vigorously washed multiple times with deionized water and then dried in a vacuum oven at 80°C for 12 h.

### Preparation of PVC NF Membrane from Iodinated PVC Granules

Electrospinning solution was prepared by dissolving iodinated PVC granules (15 wt. %) in a 1:1.5 volume ratio of THF to DMF. Iodinated PVC NF membrane was produced by electrospinning this solution at 22 kV at an ejection rate of 2 ml h<sup>-1</sup> through a 21G needle and a rotating drum collector speed of 230 rpm at a distance of 110 mm.

### Preparation of Az-PVC NF Membranes

Four preparation techniques were employed to produce Az-PVC NF membrane. The first three involved use of PVC NF membrane prepared from pristine PVC granules (Figure 1) while the fourth involved PVC NF membrane prepared from iodinated PVC granules (Figure 2).

The first membrane underwent azidation by soaking for 24 h in saturated NaN<sub>3</sub> in DMF with 2% NH<sub>4</sub>Cl as catalyst to produce Az-PVC NF (org) (Figure 1A). The second was soaked for 4 h in 6.15 mol L<sup>-1</sup> NaN<sub>3</sub> and 0.037 mol L<sup>-1</sup> TBAB in H<sub>2</sub>O to produce Az-PVC NF (aq) (Figure 1B). The third was subjected to halogen exchange by soaking for 5 h in 6.15 mol L<sup>-1</sup> NaI and 0.15 mol L<sup>-1</sup> TBAB in H<sub>2</sub>O, followed by soaking for 4 h in 6.15 mol L<sup>-1</sup> NaN<sub>3</sub> and 0.037 mol L<sup>-1</sup> TBAB in H<sub>2</sub>O to produce Az-PVC NF (aq-NaI) (Figure 1C). The fourth membrane which was prepared from iodinated PVC granules was soaked for 4 h in 6.15 mol L<sup>-1</sup> NaN<sub>3</sub> and 0.037 mol L<sup>-1</sup> TBAB in H<sub>2</sub>O to produce Az-iPVC NF (Figure 2). After azidation, the membranes were washed with copious amounts of deionized water followed by 70% ethanol, and then dried in a vacuum oven at 80°C for 12 h.

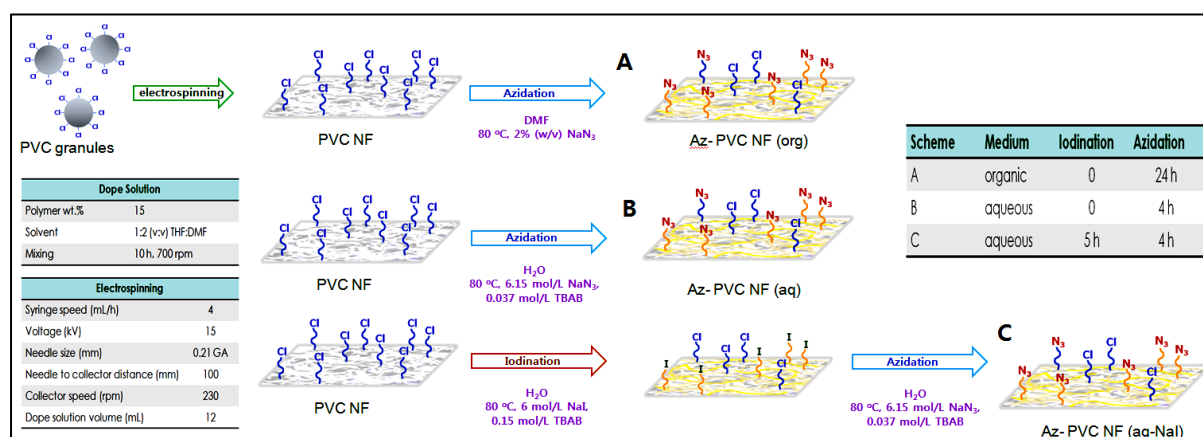
### Preparation of CB6-loaded PVC NF Membrane

Azidized PVC NF membrane was soaked in an aqueous CB6 solution in HCl at 100°C. After reaction, the membrane was washed vigorously with deionized water followed by 70% ethanol, and then dried in a vacuum oven at 80°C for 12 h.

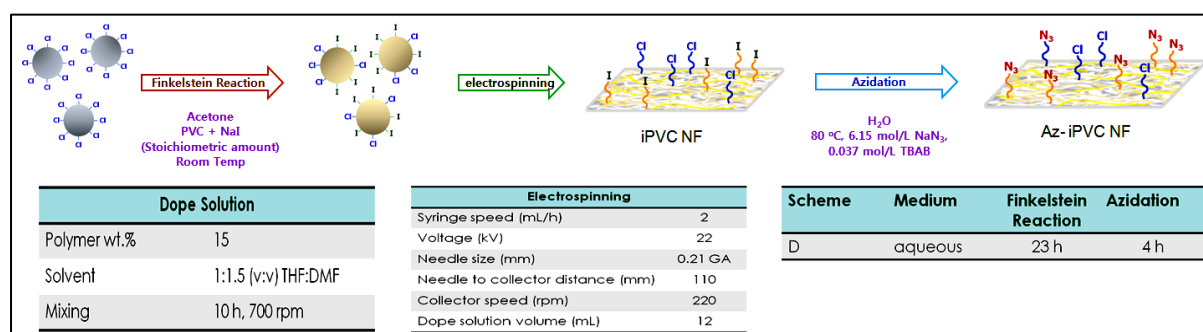
## Instrumentation

Functional groups in membrane materials were identified using a Fourier transform infrared (FTIR) -attenuated total reflectance (ATR) (Varian Scimitar 2000, USA) spectrometer. Morphology and approximate surface elemental compositions were determined using a scanning electron microscope equipped with energy dispersive X-ray spectrometer (SEM-EDX, Hitachi S-3500 N, Japan) while fiber diameters ( $\varnothing$ ) were measured from SEM images using

ImageJ (Schneider et al. 2012). Mechanical properties were determined using a Universal Testing Machine (UTM LFPlus, Lloyd Instruments, UK) equipped with a 1 kN load cell. Azide loadings were quantified by elemental analysis using an Elemental Analyzer (Elementar Analysensysteme GmbH, Germany). Water droplets for contact angle ( $\Theta$ ) measurements were visualized with a video microscope (SCALAR VL-11S, Japan).



**Figure 1.** Preparation of membranes (A) Az-PVC NF (org), (B) Az-PVC NF (aq), and (C) Az-PVC NF (aq-NaI) using electrospun pristine PVC granules.



**Figure 2.** Preparation of Az-iPVC NF using electrospun iodinated PVC granules.

## RESULTS

### Characterization of azidized PVC NF Membranes

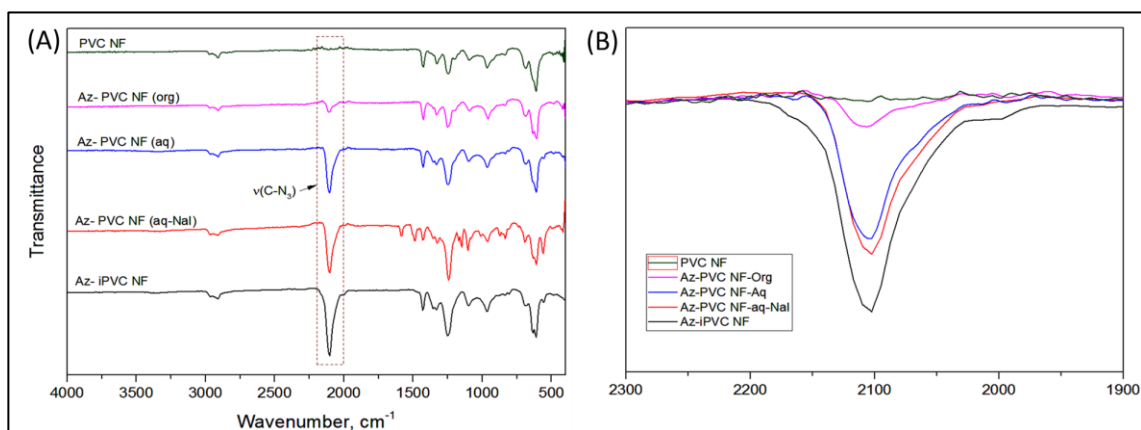
The FTIR spectra of pristine and azidized PVC NF membranes are provided in Figure 3A. The emergence of the azide peak at  $2100\text{ cm}^{-1}$  suggests the presence of azido ( $-\text{N}_3$ ) groups in the membrane which can be attributed to successful substitution of chlorine (Ouerghui et al. 2016). Results revealed a remarkable increase in the intensity of the azide peak in azidized PVC NF membranes prepared using aqueous azidation media. The superposition of the azide peaks (Figure 3B) provides a more vivid comparison of the peak intensities, with Az-iPVC NF exhibiting the strongest.

The azide peak intensity follows the trend Az-iPVC NF > Az-PVC NF (aq-NaI) > Az-PVC NF (aq) > Az-PVC NF (org).

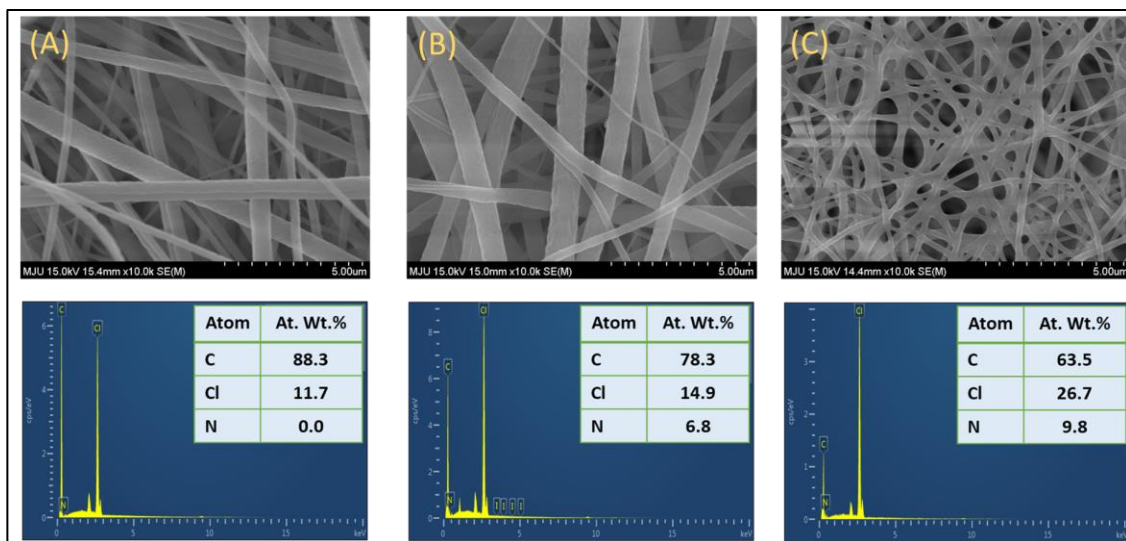
The actual azide load of each membrane was quantified via elemental analysis and results are summarized in Table 1. Results showed that the nitrogen content and consequently the azide load follow the trend Az-iPVC NF > Az-PVC NF (aq-NaI) > Az-PVC NF (aq) > Az-PVC NF (org) which is the same trend observed in the FTIR azide peak intensities. Az-iPVC NF membrane which registered the highest nitrogen content has an azide load of approximately  $2.36\text{ mmol g}^{-1}$ .

Scanning electron microscopy images of pristine and azidized membranes show that electrospun PVC features thin, long fibers with nanosized diameters (Figure 4). Polyvinyl chloride nanofiber (PVC NF,  $\text{Ø} \sim 413 \pm 195$  nm) and Az-PVC NF (aq-NaI,  $\text{Ø} \sim 426 \pm 201$  nm) membranes are essentially morphologically similar which means that iodination and subsequent azidation of PVC NF membrane cause no change in the physical structure of the membrane (Figures 4A and 4B). On the other hand,

having been prepared using a different set of electrospinning conditions, Az-iPVC NF ( $\text{Ø} \sim 187 \pm 53$  nm) features thinner, densely meshed nanofibers (Figure 4C). Results from surface elemental analyses via EDX further confirmed the presence of azide groups in Az-PVC NF and Az-iPVC NF, which contain approximately 6.8% and 9.8% nitrogen, respectively (Figures 4B and 4C).



**Figure 3.** Fourier transform infrared spectra of (A) pristine PVC NF membrane and azidized PVC NF membranes prepared using different techniques, and (B) a superposition of their azide peaks.



**Figure 4.** Scanning electron microscopy image and EDX spectra of (A) PVC NF, (B) Az-PVC NF (aq-NaI) and (C) Az-iPVC NF.

**Table 1.** Elemental composition and azide loading of PVC NF membranes.

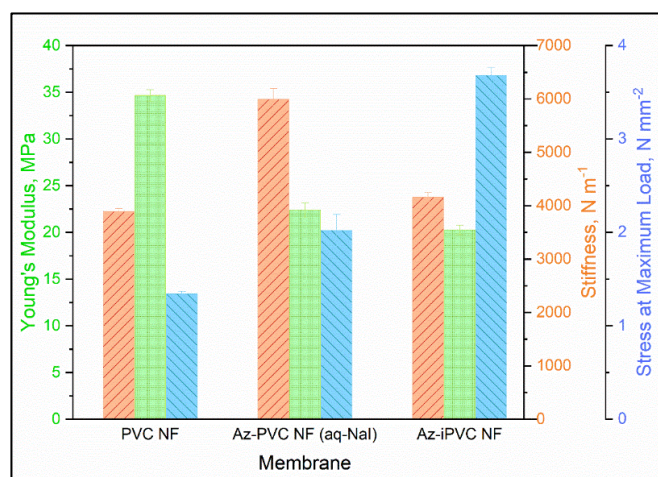
Membrane	% C	% H	% N	Azide Load ( $\text{mmol g}^{-1}$ )
PVC NF	38.64	4.992	0.35	0.08
Az-PVC NF (org)	38.64	4.974	2.80	0.67
Az-PVC NF (aq)	38.66	4.968	6.92	1.65
Az-PVC NF (aq-NaI)	39.23	4.905	8.51	2.03
Az-iPVC NF	39.00	4.868	9.91	2.36

The mechanical properties of the membranes were found to be at variance with one another. Young's moduli are around 20-35 MPa, stiffnesses are at 3900-6000 N m<sup>-1</sup> and stresses at maximum load are at 1.34-2.68 N mm<sup>-2</sup> (Figure 5). Despite the variability, the mechanical properties of all membranes prepared in this study are comparable with published data for which Young's moduli in the range 3.75-308 MPa and stresses at maximum load of 0.67-9.1 N mm<sup>-2</sup> have been reported (Pham et al. 2021).

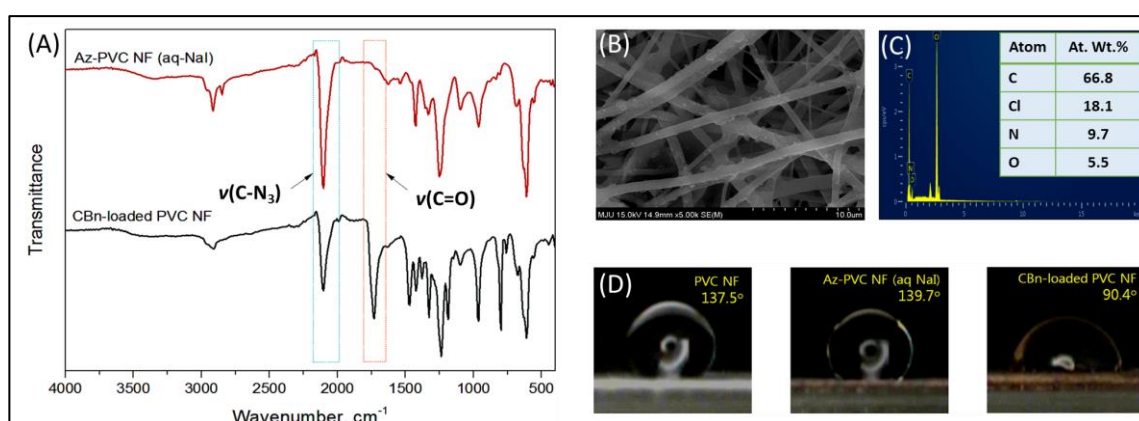
#### Covalent Immobilization of CB6 in Az-PVC NF

Thermally induced generation of nitrenes led to successful immobilization of CB6, as confirmed by the subsequent weakening of the azide peak (2100 cm<sup>-1</sup>) and the emergence of the carbonyl stretch (1740 cm<sup>-1</sup>) in the FTIR spectra of CB6-loaded PVC NF membrane (Figure 6A). The SEM image of CB6-loaded PVC NF ( $\varnothing$ ~ 418±173 nm) membrane is

shown in Figure 6B while the corresponding surface elemental scan which confirms the existence of oxygen, and hence surface-immobilized CB6, is shown in Figure 6C. Significant changes in surface properties were observed from static water droplet contact angle measurements whereby the previously hydrophobic surfaces of pristine ( $\Theta = 137.5^\circ$ ) and azidized ( $\Theta = 139.7^\circ$ ) PVC NF membrane became more hydrophilic after CB6 immobilization (Figure 6D), causing the water droplet to be flatter ( $\Theta = 90.4^\circ$ ). This improved hydrophilicity is likely due to increased attractive electronic interactions (e.g. hydrogen bonding) between water molecules at the droplet interface and the carbonyl groups of immobilized CB6 on the PVC surface which effectively disrupts the droplet surface tension. As with most surfaces, the static water droplet contact angle approaches 90° or less as a consequence of hydrophilicity (Law 2014).



**Figure 5.** Mechanical properties of PVC NF, Az-PVC NF (aq-Nal) and Az-iPVC NF.



**Figure 6.** (A) Fourier transform infrared spectra of pristine and CB6-functionalized Az-PVC NF; (B) SEM image and (C) EDX spectrum of CB6-functionalized Az-PVC NF; and (D) water contact angles of PVC NF, Az-PVC NF and CB6-functionalized Az-PVC NF.



## DISCUSSION

### Azidation in Aqueous Systems and the Importance of Pre-iodination

Azidation of PVC is usually carried out in a polar aprotic solvent (e.g. DMF) since this favors  $S_N2$  reactions by keeping the nucleophile prone to reactive collisions. This approach, however, cannot be applied to electrospun PVC NF membrane because the membrane deteriorates in DMF. Moreover, poor solubility of  $\text{NaN}_3$  in DMF limits the amount of azide nucleophiles available. An azidation solvent system that keeps the PVC NF membrane intact and dissolves the azide salt easily had to be resorted to, and this was accomplished by performing the reaction in aqueous media in the presence of TBAB as phase transfer catalyst (Lakshmi and Jayakrishnan 2002). In this method, the soluble organic cation  $\text{N}^+(\text{Bu})_4^+$  transports  $\text{N}_3^-$  from the aqueous phase to the organic solid phase of the PVC NF membrane where the nucleophilic substitution occurs. These explain the remarkable improvement in azide load of Az-PVC NF (aq) relative to Az-PVC NF (org).

Because a high loading of azide groups is desired, chlorine groups in the PVC NF membrane were pre-displaced with iodine, again by phase transfer catalysis using TBAB. Although a weaker base, iodine is a stronger nucleophile than chlorine if the reaction occurs in a polar protic solvent like water, which allows the replacement of chlorine to proceed spontaneously. Meanwhile, iodine is better than the azide ion at displacing chloride groups in the PVC backbone via nucleophilic substitution, which means that under the same conditions, more chlorine groups are replaced in PVC if it were reacted with iodine anions than when reacted with azide anions. In short, iodine ions are more effective than azide ions at sweeping off chlorine atoms from the PVC chain. Once attached to PVC, iodine atoms are not as firmly bonded as the chlorine groups that they displaced due to the larger atomic size of iodine. This causes iodine to have a greater tendency than chlorine to leave the alkyl backbone of PVC. As such, iodine atoms are more vulnerable to displacement by smaller nucleophiles like azide ions. This makes azidation of PVC more pervasive if preceded by iodination because iodination serves to facilitate the displacement of chlorine atoms and is easily substituted by azide ions thereafter. This explains the higher azide load of Az-PVC NF (Aq-NaI) than Az-PVC NF (Aq).

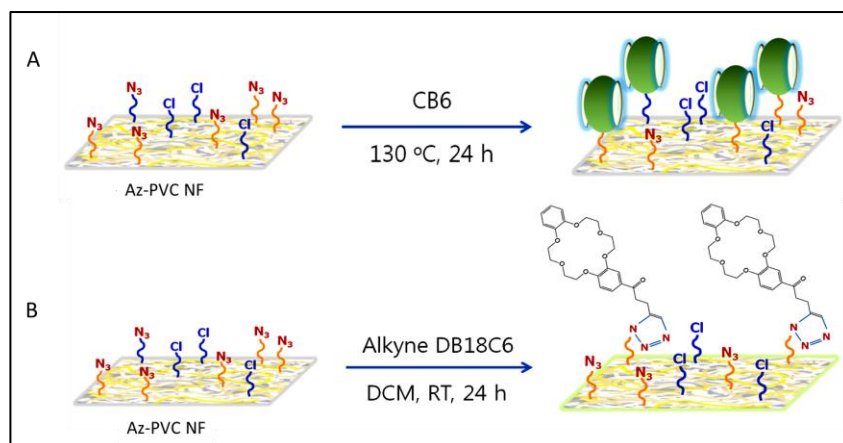
To further exploit the efficacious azidation of pre-iodinated PVC NF, a more vigorous iodination reaction through halogen exchange in acetone was performed. The procedure, however, cannot be applied to PVC NF membrane because the membrane dissolves completely in acetone. Fortunately, PVC granules are not so easily dissolved by acetone, and the halogen exchange reaction—also called Finkelstein reaction—may be performed successfully. Iodinated

PVC NF membrane prepared from such granules registered the highest azide loading among all the different azidation strategies employed.

### Applications of Az-PVC NF Membrane

Azido groups on PVC NF membrane can participate in nitrene insertion or click reactions resulting in covalently bonded molecules (Figure 7). Covalent immobilization of molecules on solid supports, especially the custom-made expensive functional molecules such as cucurbiturils and crown ethers, allows the possibility of multiple reuse which helps offset costs (Escobar et al. 2021; Escobar et al. 2022).

The conversion of azido groups to nitrenes that insert non-selectively into C-H bonds is particularly useful in covalently immobilizing molecules devoid of functional attachments or are difficult to chemically modify such as the cucurbit[*n*]urils or  $\text{CBn}$ 's where *n* represents the number of glycoluril units comprising the molecule.  $\text{CBn}$ 's are a family of macrocyclic molecules that possesses highly symmetric structures with two identical negatively charged portal carbonyl groups and a hydrophobic cavity (Assaf and Nau 2015). Supramolecular assemblies built from  $\text{CB6}$  with 4,4',4''-benzene-1,3,5-triyl tribenzoic acid as directing agent had been shown to exhibit high selectivity for  $\text{Cs}^+$ —an abundant component of radioactive wastes—against common alkali ions in basic medium (Chen et al. 2014), while  $\text{CB8}$  in free form had been reported to have a high adsorption capacity for  $\text{Pb}^{2+}$  (Sun et al. 2016). Meanwhile,  $\text{CB6}$  by itself has been found to form  $\text{CB6-Cs}^+$  complexes, registering stability constants that are several orders of magnitude higher than those formed with either 18-crown-6 ether or cryptand 222, an outcome attributed mainly to the strong negative charge of the carbonyl rim and a close cavity-ion size match (Buschmann et al. 1992). Because of their ability to interact selectively with solutes, the covalent immobilization of  $\text{CBn}$ 's onto solid supports such as PVC NF is a huge step into their utilization for practical solute sequestration technologies. On the other hand, azido groups can participate in click reactions which permit fast and highly selective attachment of molecules that possess alkyne functionalities. Click reactions have been used extensively in the development of water decontamination technologies involving various specialized alkyne-bearing molecules on different azido-functionalized supports (Liang and Astruc 2011; Nisola et al. 2020; Ouerghui et al. 2016). With the azido-functionalized PVC NF membranes developed in this study, applications that take advantage of nitrene or click reactions might be extended over to membranes, thereby allowing the development of membrane materials with highly specialized functionalities.



**Figure 7.** Examples of molecule attachment schemes in Az-PVC NF: (A) nitrene insertion in alkyl chains of cucurbiturils, or (B) click reaction with alkyne-functionalized crown ethers.

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## ETHICAL CONSIDERATIONS

This research did not require human or animal subjects.

## DECLARATION OF COMPETING INTEREST

The authors declare that they have no known competing financial interests or personal relationships that could have influenced the work reported in this paper.

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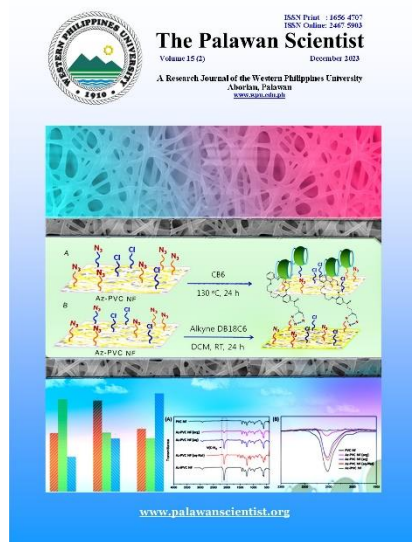
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**ROLE OF AUTHORS:** ECE- methodology, investigation, writing of original draft; GMN- conceptualization, project administration, investigation; WJC- resources, supervision, project administration.



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# Design of a compact wastewater treatment and isolation of microbial consortia for nutrient reduction of plastic recycling wastewater

Jaren U. Tulipan and Jey-R S. Ventura\*

*Biomaterials and Environmental Engineering Laboratory, Department of Engineering Science, College of Engineering and Agro-Industrial Technology, University of the Philippines Los Baños, College, Laguna, Philippines*

\*Correspondence: [jsventura@up.edu.ph](mailto:jsventura@up.edu.ph)

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## ABSTRACT

The wastewater generated during the plastic recycling is an aspect that is often overlooked in the process. This wastewater contains pollutants that can affect the environment and human health. To address this problem, a study was conducted to design and evaluate the performance of an anoxic-oxic (AO) system and an anoxic-oxic-oxic (AO<sub>2</sub>) system in treating plastic recycling wastewater. Additionally, the study collected wastewater from a plastic recycling company and activated sludge from a sewage treatment plant to test the viability of biological treatment for treating plastic recycling wastewater. The results showed that both systems were effective at reducing chemical oxygen demand (COD), with the AO system having a higher percent COD removal (98.13%) than the AO<sub>2</sub> system (85.33%). Also, there were 18 unique colonies isolated from the wastewater and activated sludge. Overall, the study concludes that biological treatment can be used to treat plastic recycling wastewater, and the design of upscaled system is crucial to fully solve the wastewater problem.

**Keywords:** laboratory scale reactor, recycling industry wastewater, wastewater treatment design

## INTRODUCTION

Plastic waste has become a major environmental issue in recent years, with much attention focused on the impact of plastic pollution on the environment and human health. One of the most effective ways to reduce the environmental impact of plastic waste is to recycle it. Plastic recycling is the process of transforming discarded plastic materials into new products that can be used again (Hopewell et al. 2009). However, one aspect of the problem that is often overlooked is the wastewater generated during the plastic recycling process (Gunarathna et al. 2010).

The washing stage plays a pivotal role in removing residues such as wood, pulp fibers, food, and adhesives, which otherwise would have detrimental effects on the final regenerated product. Therefore, these washing stages produce vast amounts of wastewater. The current washing technology used by plastic recyclers has significantly reduced water consumption to 2-3 m<sup>3</sup> of water per ton of plastic material (Altieri et al. 2021). This wastewater can contain a variety of hazardous chemicals and contaminants, and if not properly treated, it can have serious negative impacts on the environment and human health (Setiawan et al. 2021).



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The composition of wastewater varies depending on the source. Altieri et al. (2021) investigated plastic solid waste washing wastewater and found that it contains high amounts of chemical oxygen demand (COD), biochemical oxygen demand (BOD), and total nitrogen (TN). The high concentration of these pollutants may be attributed to the source of the plastic solid waste, which was mainly from the agro-industrial sector. Setiawan et al. (2021) characterized polyester synthetic fiber industry wastewater and found that it contains high amounts of COD, BOD, and total suspended solids (TSS). Santos et al. (2005) characterized effluents from the washing steps of polyethylene terephthalate (PET) and polyolefins cleaning process, while Santhmayor et al. (2020) characterized wastewater from waste plastic recycling machines. Both studies indicated high amounts of COD and TSS in the wastewater. The high COD values may be attributed to the surfactant used in the washing step of the plastics (Santos et al. 2005). Meanwhile, Kolbl (2016) characterized plastic processing plant wastewater and found a high amount of total phosphorous (TP). The above findings show that plastic recycling wastewater can have high amounts of COD, BOD, TP, TN, and TSS. High levels of these pollutants may cause nitrate contamination, eutrophication, and ammonia toxicity (Curtin et al. 2011). Additionally, high levels of TSS may cause problems for aquatic life as TSS may include solids from the plastic recycling wastewater that can harm the environment (Santhmayor et al. 2020). These results show proof that plastic recycling wastewater can be harmful. Therefore, proper treatment of plastic recycling wastewater is essential to prevent these negative impacts.

Plastic recycling wastewater poses harmful effects on the environment, and there is still a lack of available and cost-effective technology for the plastic recycling industry to implement. In the case of Valenzuela City, Philippines, the researchers found that most plastic recycling facilities do not have the necessary equipment or infrastructure to properly treat wastewater. Most of the facilities reuse their wastewater without treatment to cut operating costs. This is not only hazardous but also decreases the quality of recycled plastic (Altieri et al. 2021). Efforts to mitigate the effects of this in the wastewater were made by applying flocculation methods. However, the lack of knowledge on wastewater composition and appropriate treatment system makes it difficult to assess.

To address this problem, this study aimed to design and evaluate the performance of an anoxic-oxic (AO) system and an anoxic-oxic-oxic (AO<sub>2</sub>) system in treating plastic recycling wastewater. These systems were investigated for their potential to enhance biological nutrient removal from plastic recycling wastewater. Also, developing these systems ensures a

low-cost technology that can be adapted by the plastic recycling industry. The specific objectives were the following: (1) to characterize the wastewater generated by a local plastic recycling facility in Valenzuela City, Philippines; (2) to isolate and assess the potential of microbial consortia found in activated sludge for the purpose of treating plastic recycling wastewater; and (3) to determine the removal efficiencies and COD reduction performance of the AO and AO<sub>2</sub> systems. Through these objectives, the study sought to contribute valuable insights into the treatment of plastic recycling wastewater.

## METHODS

### Wastewater Collection and Characterization

Twenty liters of wastewater from a plastic recycling company in Valenzuela City, Philippines were collected and stored. The wastewater was subjected to analysis using standard methods for the examination of water and wastewater to determine the following parameters: COD, BOD, nitrates, phosphates, TSS, pH, and dissolved oxygen.

### Sludge Collection and Acclimatization

Similarly, 20-L of activated sludge were collected from a sewage treatment plant in Biñan, Laguna, Philippines, and divided into two setups for acclimatization in synthetic wastewater. One setup was anoxic, while the other was aerobic using an aquarium pump for aeration. The synthetic wastewater used is composed of 5 g L<sup>-1</sup> glucose, 1 g L<sup>-1</sup> NH<sub>4</sub>Cl, 1 g L<sup>-1</sup> Na<sub>2</sub>CO<sub>3</sub>, 1 g L<sup>-1</sup> NaHCO<sub>3</sub>, 0.2 g L<sup>-1</sup> K<sub>2</sub>HPO<sub>4</sub>, and 1 mL L<sup>-1</sup> of trace metal solution (Alzate-Gavira et al. 2003). To maintain the culture setups, synthetic wastewater was regularly replaced.

### Isolation of Bacteria from the Activated sludge

**Microbial preservation and isolation.** To have a starting culture for the wastewater treatment facility, the microbial consortia from the activated sludge were isolated and stored. Ten mL samples were taken from the treated wastewater, the mixed liquor from the anoxic setup, and the settled activated sludge solution. The samples were stored in 2-mL centrifuge tubes. Then, serial dilution of each sample was done up to 10<sup>-5</sup> dilutions. Plate count agar (PCA) was prepared to serve as the growth medium for the microorganisms in the activated sludge because it is usually used for bacterial enumeration in water and wastewater (Aryal 2022; Ogodo et al. 2022). It is composed of 5 g L<sup>-1</sup> casein hydrolysate, 2.5 g L<sup>-1</sup> yeast extract, 1 g L<sup>-1</sup> glucose, and 15 g L<sup>-1</sup> agar.

After which, the pour plate method was then used. Four plates of treated wastewater, an anoxic setup of mixed liquor, and an activated sludge solution were made. Then, two plates of each were placed in a

desiccator to create an anaerobic condition. The other two plates of each were placed in storage to facilitate aerobic conditions. Under these conditions, the microorganisms were allowed to grow for two days.

Luria Bertani (LB) broth was prepared to serve as the medium for the growth cultivation of bacteria found in the plates. It is composed of  $100 \text{ g L}^{-1}$  casein hydrolysate,  $50 \text{ g L}^{-1}$  yeast extract, and  $60 \text{ g L}^{-1}$  glucose. The bacteria colonies found in each plate were placed on test tubes filled with 15 mL of LB broth. Two test tubes of colonies found from treated wastewater, anoxic setup mixed liquor, and activated sludge solution were placed in a desiccator to facilitate anaerobic conditions. The other two test tubes were placed in storage to facilitate aerobic conditions. The bacteria were allowed to grow for three days. After which, glycerol stocks were made from each test tube. The isolates were then stored at  $4^\circ\text{C}$  with 1 mL of LB broth and 1 mL of glycerol.

**Regrowing and characterization of the stock inoculum.** Plate count agar with nystatin ( $10,000 \text{ units ml}^{-1}$ ) was prepared and served as a medium for growth for the microbial consortia plates. Nystatin was added to ensure that only bacteria would grow on the PCA plates. A total of 24 plates of PCA were prepared. Four plates were prepared for both anaerobic and aerobic setups for treated wastewater, anoxic setup mixed liquor, and activated sludge solution. The plates were incubated in an anaerobic (using a desiccator) and aerobic condition for approximately 4 days to allow the bacteria to grow.

Colonies were identified based on their observed characteristics on the plates. Gram staining was done to further characterize the unique bacteria colonies by their cell membrane structures. The gram-stained bacteria were observed under a microscope.

## AO and AO<sub>2</sub> Reactor Design and Fabrication

**AO system.** A laboratory-scale AO system was designed to determine the COD reduction of the system using the activated sludge. Design considerations were as follows. The volume of the oxic chamber and the anoxic chamber was determined in the ratio of 1:2. The height of the reactor was set to 120 mm, which is roughly the same height as a 250-mL beaker. The width of the reactor was set to 68 mm to fit a magnetic stirrer inside the reactor. To achieve a 1:2 volume ratio for the oxic and anoxic chambers, the length of each chamber was set to 68 mm and 136 mm, respectively. Alternating hanging and standing baffles were placed inside the reactor to compartmentalize the reactors and regulate the flow of liquid from the anoxic chamber to the oxic chamber. The baffles are 3 mm thick and have vertical clearances of 20 mm. The reactor was fabricated by 3D printing using Creality CR-10 Max. The 3D printing parameters were based

on the study of Nasr Esfahani et al. (2021) using polylactic acid (PLA) filament for the printing process. The wall thickness of the reactor was set to 4 mm, and the base thickness was set to 10 mm. The infill pattern was set to tri-hexagon with an infill density of 75% (Nasr Esfahani et al. 2021). An outlet was placed at the same height as the standing baffle. The 3D model of the fabricated AO reactor is shown in Figure 1.

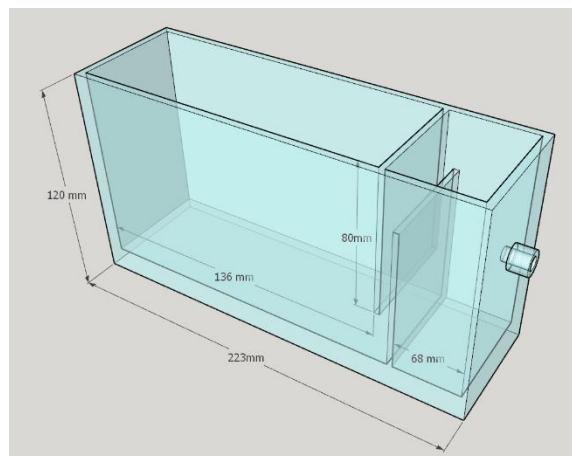


Figure 1. 3D model of the fabricated anoxic-oxic reactor.

The AO system configuration is depicted in Figure 2. The hydraulic retention time (HRT) for the anoxic chamber was set to 16 hr and the HRT for the oxic chamber was set to 8 hr. Meanwhile, the pH and the temperature in both the anoxic and oxic chambers were set at pH 7 and  $25^\circ\text{C}$ . Meanwhile, the influent flowrate and the effluent flowrate were both set at  $1 \text{ m}^3 \text{ d}^{-1}$ .

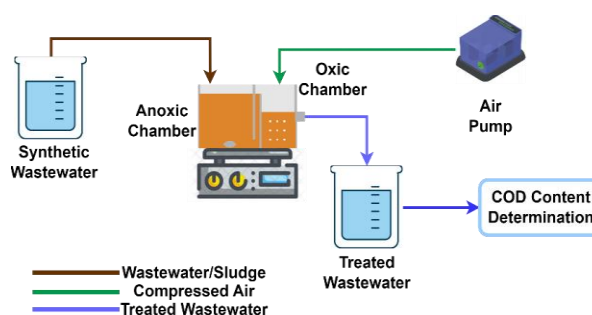
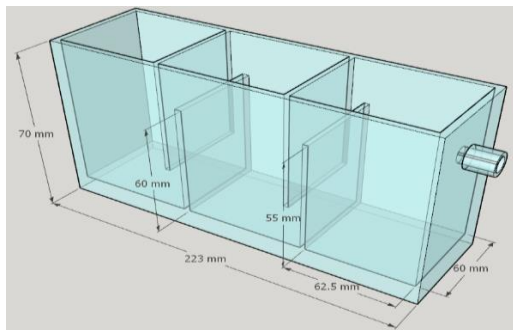


Figure 2. Anoxic-oxic system configuration.

**AO<sub>2</sub> system.** From the results of the AO laboratory scale reactor, it was determined to improve the system configuration by adding another oxic chamber to the design. With this, an AO<sub>2</sub> laboratory scale reactor with cascading baffles was designed. The height of the reactor was set to 70 mm. The wall thickness of the reactor was set to 4 mm, while the base thickness was set to 5 mm. The reactor consists of three chambers – one anoxic chamber and two oxic chambers. The chambers were equal in area, with

length and width measuring 62.5 mm and 60 mm, respectively. Alternating hanging and standing baffles were designed to compartmentalize the chambers and regulate the wastewater flow. The standing baffles were designed to have a 5 mm decrease in height to facilitate cascading flow from one chamber to the next. The standing baffle heights were 60 mm and 55 mm with a vertical clearance of 20 mm. A single outlet was placed at the same height as the last standing baffle and has an inner diameter of 10 mm.

A 3D printed laboratory scale reactor was used in the setup. The infill design was chosen to be tri-hexagon to have better 2D stress management. The infill density was set to be 60%, which was found to be sturdy enough to prevent any leakages based on previous test prints. The 3D model of the fabricated AO<sub>2</sub> reactor is shown in Figure 3. The AO<sub>2</sub> system configuration is shown in Figure 4. The hydraulic retention time (HRT) for the anoxic and two oxic chambers was set to 8 hr. The pH and the temperature in all the chambers were set at pH 7 and 25°C, respectively. Meanwhile, the influent flowrate and the effluent flowrate were both set at 1 m<sup>3</sup> d<sup>-1</sup>.



**Figure 3.** 3D model of the fabricated anoxic-oxic-oxic reactor.

**Chemical Oxygen Demand Reduction Performance**

A10-L container containing synthetic wastewater served as the influent into the system. The wastewater was continuously fed to the reactors at a



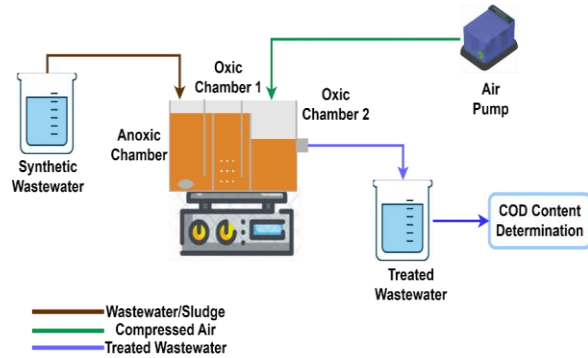
$$\text{Theoretical COD} = \text{mol oxidizable material} \times \frac{\text{mol } O_2}{\text{mol oxidizable material}} \times \text{MM } O_2 \tag{Equation 2}$$

$$\text{Absorbance (nm)} = 0.1396 + 0.5827 * \text{COD (g L}^{-1}\text{)} \tag{Equation 3}$$

$$\text{COD reduction rate} = \frac{\text{Initial COD concentration} - \text{Final COD Concentration}}{\text{Final time} - \text{initial time}} \tag{Equation 4}$$

$$\% \text{ COD reduction} = \frac{\text{Initial COD concentration} - \text{Final COD Concentration}}{\text{Initial COD Concentration}} \times 100 \tag{Equation 5}$$

rate of 1 L day<sup>-1</sup>. The collected activated sludge was placed in the different chambers of the reactor. An air stone was placed in the oxic chamber that was connected to an air pump, which served as the air supply for the oxic chamber, while a magnetic stirrer was placed inside the anoxic chamber.



**Figure 4.** Anoxic-oxic-oxic system configuration.

Meanwhile, the COD of the influent and effluent were measured using the dinitrosalicylic acid (DNS) method, which is based on the National Renewable Energy Laboratory's method for measuring cellulase activity (Adney and Baker 2008). The DNS method was used to determine the COD reduction performance of the reactors by assuming that the glucose content measured by the DNS method is equal to the COD content of the samples. The theoretical COD can be calculated based on the combustion reaction of glucose and oxygen to produce carbon dioxide and water, as seen in Equation 1. Then the theoretical COD can be calculated using Equation 2. Using Equation 2, it can be calculated that for 1000 ppm of glucose, the COD is approximately 1067 ppm. Equation 3 shows the standard curve equation used to determine the COD of the samples for this experiment. The COD reduction rate and percent COD reduction were calculated using Equations 4 and 5.

The reagent for the DNS method was prepared by adding the following chemicals to 1 L of distilled water: 7.5 g dinitrosalicylic acid, 14 g NaOH, 216 g  $\text{KNaC}_4\text{H}_4\text{O}_6 \cdot 4\text{H}_2\text{O}$ , 5.4 mL phenol, and 5.9 g  $\text{Na}_2\text{S}_2\text{O}_5$ . A 1.5 mL of this reagent was then mixed with 0.75 mL of wastewater. The mixture was then heated at 100°C for 5 min and cooled in an ice water bath to develop color. The absorbance of this color was then measured using a spectrophotometer at 540 nm and compared to a standard curve to determine the COD concentration in the wastewater. The standard curve was generated by making standard solutions of glucose with concentrations of 0.2 g L<sup>-1</sup> to 1.2 g L<sup>-1</sup>.

## RESULTS

### Plastic Recycling Wastewater Characteristics

The collected wastewater from the plastic recycling industry was subjected to wastewater characteristic analysis using standard methods. The results are shown in Table 1. From the table, the BOD, nitrate, phosphate and TSS concentration in the wastewater were 651 mg L<sup>-1</sup>, 337 mg L<sup>-1</sup>, 3.16 mg L<sup>-1</sup>, and 338 mg L<sup>-1</sup> respectively. Meanwhile, the COD concentration of the wastewater was 5682 mg L<sup>-1</sup>. Fecal coliform was also found in the wastewater at 1.8 MPN per 100 mL. The color of the wastewater was

measured at 25 TCU. The pH and temperature of the wastewater were 6.47 and 28.2 °C, respectively.

### Bacterial Isolates from Activated Sludge

Tables 2 and 3 summarize the colony and bacterial characteristics observed in the isolates. In total, 18 unique isolates were observed. Seven aerobic isolates were identified from the PCA plates while eleven anaerobic isolates were identified from the PCA plates.

### Performance of the AO and AO<sub>2</sub> System

The percent COD reduction and the COD reduction rate of the AO system are shown in Figure 5. The figure shows that on the first day, there was a COD reduction rate of 60.42 mg L<sup>-1</sup> h<sup>-1</sup>, which translates to a 99.01% reduction. This high reduction rate was maintained throughout the process. Overall, the process was highly effective, with all reduction rates above 92%. Meanwhile, the percent COD reduction and the COD reduction rate of the AO<sub>2</sub> system are shown in Figure 6. The figure shows that the system was stable at the start of the experiments but was unstable towards the end. The AO system showed a more stable treatment efficiency as compared to the AO<sub>2</sub> system in terms of COD reduction performance.

**Table 1.** Wastewater characteristics of plastic recycling industry effluent as compared to DENR general effluent standards. SM – standard methods for the examination of water and wastewater.

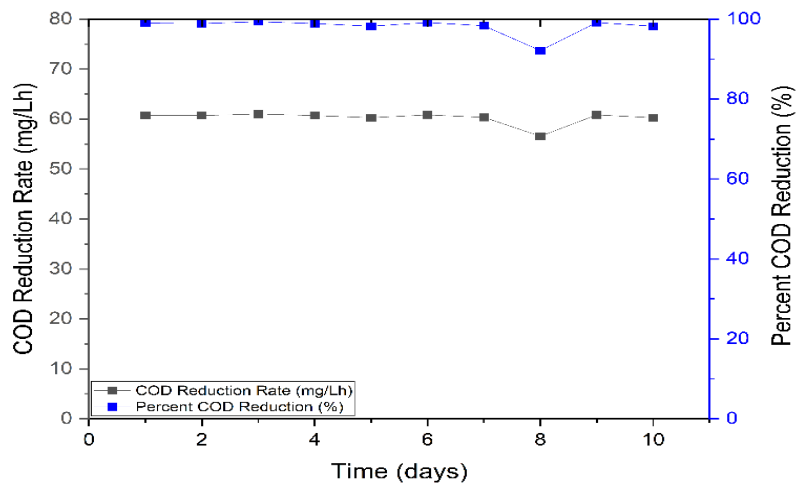
Analysis	Method	Unit	Wastewater	DENR GES Class C
Fecal Coliform	SM 9221 E	MPN per100 mL	1.8	400
Color	SM 2120 B	TCU	25	150
Nitrate	SM 4500 – NO <sub>3</sub> E.	mg L <sup>-1</sup>	337	14
COD	SM 5220 B	mg L <sup>-1</sup>	5682	100
BOD	SM 5210 B.	mg L <sup>-1</sup>	651	50
TSS	SN 2540 D.	mg L <sup>-1</sup>	338	100
Phosphate	SN 4500-P D.	mg L <sup>-1</sup>	3.16	4
pH	-	pH	6.47	6.0-9.5
Temperature	-	°C	28.2	25-31

**Table 2.** Colony and bacterial characteristics of isolates from aerobic conditions.

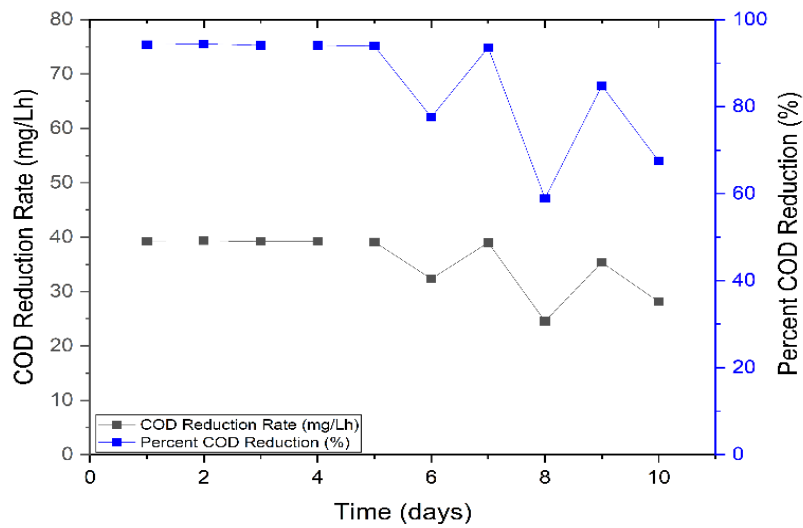
Isolate	Colony Characteristics			Bacterial Characteristics	
	Color	Form	Elevation	Gram Stain	Shape
WW O <sub>2</sub> 1	Creamy	Circular	Raised	Gram-negative	Rod
WW O <sub>2</sub> 2	White	Circular	Flat	Gram-negative	Rod
WW O <sub>2</sub> 3	Bright Orange	Circular	Raised	Gram-negative	Rod
WW O <sub>2</sub> 4	Whitish Yellow	Circular	Flat	Gram-positive	Rod
Sol'n O <sub>2</sub> 1	Creamy	Irregular	Flat	Gram-negative	Rod
Sol'n O <sub>2</sub> 2	Whitish Yellow	Irregular	Flat	Gram-positive	Rod
Iso 1 O <sub>2</sub>	White	Circular	Raised	Gram-positive	Cocci

**Table 3.** Colony and bacterial characteristics of isolates from anaerobic conditions.

Isolate	Colony Characteristics			Bacterial Characteristics	
	Color	Form	Elevation	Gram Stain	Shape
Sol'n An 1	Creamy	Circular	Flat	Gram-negative	Cocci
Sol'n An 2	White	Circular	Flat	Gram-negative Gram-positive	Rod Cocci
Sol'n An 3	Yellow	Circular	Flat	Gram-positive	Rod
WW An 1	White	Circular	Flat	Gram-negative	Rod
WW An 2	Creamy	Circular	Flat	Gram-negative Gram-positive	Rod Cocci
WW An 3-1	Creamy	Circular	Flat	Gram-negative	Rod
WW An 3-2	Yellow	Circular	Raised	Gram-negative	Rod
WW An 4	Yellow	Circular	Flat	Gram-negative	Rod
Sludge An 1	Creamy	Irregular	Flat	Gram-negative	Cocci
Sludge An 2	Creamy	Irregular	Flat	Gram-negative Gram-positive	Rod Rod
Iso 2 An	Faint Orange	Circular	Flat	Gram-negative	Cocci



**Figure 5.** Chemical Oxygen Demand reduction performance of Anoxic-oxic system.



**Figure 6.** Chemical Oxygen Demand reduction performance of the Anoxic-oxic-oxic system.



## DISCUSSION

### Plastic Recycling Wastewater Characteristics

Table 1 shows the wastewater characteristics of the plastic recycling wastewater from local plastic recycling facilities in the Philippines as compared to the Class C general effluent guidelines set by the DENR. The plastic recycling industry falls under the Philippine Standard Industrial Classification Code (PSIC Code) 38210. Based on DENR Administrative Order 2016-08, the significant parameters that should be monitored for the effluent of industries under PSIC Code 38210 are pH, temperature, color, COD, total coliform, ammonia, nitrate, phosphate, and TSS. The BOD in the wastewater ( $651 \text{ mg L}^{-1}$ ) was significantly higher than the effluent standard set by the Department of Environment and Natural Resources (DENR) ( $50 \text{ mg L}^{-1}$ ). This result indicates that the wastewater has a high level of organic matter, which can deplete oxygen levels in water bodies. The COD of the wastewater ( $5682 \text{ mg L}^{-1}$ ) was also found to be significantly higher than the DENR effluent standard ( $100 \text{ mg L}^{-1}$ ). This high COD level suggests a substantial presence of organic pollutants in the wastewater, which can have adverse effects on the receiving water bodies (Tchobanoglous et al. 2003). Also, nitrate concentration in the wastewater ( $337 \text{ mg L}^{-1}$ ) was above the DENR effluent standard ( $14 \text{ mg mg L}^{-1}$ ). High levels of nitrates can pose a risk to aquatic life and can also contaminate groundwater (Tay et al.

2003). The total suspended solids (TSS) of the wastewater ( $338 \text{ mg L}^{-1}$ ) also exceeded the DENR effluent standard ( $100 \text{ mg L}^{-1}$ ). High levels of TSS can reduce the clarity of water, making it unsuitable for various purposes, including industrial and agricultural uses (Tchobanoglous et al. 2003). The phosphate concentration ( $3.16 \text{ mg L}^{-1}$ ) was observed to be barely below the DENR effluent standard ( $4 \text{ mg L}^{-1}$ ). High levels of phosphate can lead to eutrophication, which can cause algae blooms and oxygen depletion in water bodies (Tay et al. 2003). The fecal coliform, color, pH, and temperature of the plastic recycling wastewater were well below the effluent standards. However, it has higher concentrations of nitrates ( $337 \text{ mg L}^{-1}$ ), BOD ( $651 \text{ mg L}^{-1}$ ), TSS ( $338 \text{ mg L}^{-1}$ ), and phosphates ( $3.16 \text{ mg L}^{-1}$ ) than the general effluent standards. This data proves that it is unsafe to discharge wastewater from these industries without proper treatment.

The plastic recycling wastewater from the local industry had higher concentrations of COD than the wastewater from all the characterized wastewaters (Table 4). Also, this wastewater had higher BOD than the wastewater from the studies of Kolbl (2016) and Setiawan et al. (2021). The plastic recycling wastewater also had higher concentrations of nitrates than the study of Altieri et al. (2021). However, it also contains lower concentrations of phosphates than the wastewater from the study of Kolbl (2016) and had a lower pH than all the wastewater (Table 4).

**Table 4.** Wastewater characteristics of plastic recycling industry wastewater from different sources.

Source	Wastewater	COD ( $\text{mg L}^{-1}$ )	TSS ( $\text{mg L}^{-1}$ )	BOD <sub>5</sub> ( $\text{mg L}^{-1}$ )	TN ( $\text{mg L}^{-1}$ )	TP ( $\text{mg L}^{-1}$ )	NO <sub>2</sub> -N ( $\text{mg L}^{-1}$ )	NO <sub>3</sub> -N ( $\text{mg L}^{-1}$ )	NH <sub>3</sub> -N ( $\text{mg L}^{-1}$ )	pH
Altieri et al. (2021)	Plastic solid waste washing wastewater	1371.5	-	398	45.35	8.4	0.1	0.55	22.75	7.6
Setiawan et al. (2021)	Polyester synthetic fiber industry wastewater	579	269	264	-	-	-	-	5.4	-
Santos et al. (2005)	Effluents from washing steps of polyethylene terephthalate (PET) and polyolefins cleaning process	677.5	702.5	-	-	-	-	-	-	9.45
Ramirez-Camperos et al. (2004)	Bottle washings wastewater from soft drinks industry	64.9	66	-	-	-	-	-	-	11.5
Santhmayor et al. (2020)	Wastewater from waste plastic recycling machines	354	400	-	-	-	-	-	-	7.9
Kolbl (2016)	Plastic processing plant wastewater	457	294.3	112.3	-	933.3	248.3	940	456.3	7.6

### Isolated Bacteria in Activated Sludge

There are seven unique bacteria colonies identified in aerobic plates. There are four gram-negative bacterial isolates and three gram-positive bacterial isolates from the aerobic conditions. The color of the colonies of these isolates ranges from white to yellow. Most isolates found in aerobic conditions are rod-shaped with one isolate having a cocci shape. The bacteria colonies isolated from aerobic plates are potential nitrifying bacteria as well as biological phosphorous removal bacteria (Tay et al. 2003). Meanwhile, there are 11 unique bacterial colonies identified in anaerobic plates. Seven are found to be gram-negative while the other four are gram-positive. The color of the colony of these isolates ranges from white to yellow as well. Five isolates have a cocci shape, while the rest are rod-shaped. These bacterial colonies in anaerobic plates can serve as denitrifiers in the anoxic or anaerobic chambers of wastewater treatment (Tay et al. 2003). The presence of both aerobic and anaerobic bacterial colonies gives insight into an effective biological treatment of wastewater.

Out of the 18 isolates, 11 are gram-negative and 7 are gram-positive. The majority of the isolates being gram-negative is an indication that the bacterial community found in the activated sludge is composed of bacteria that are commonly found in wastewater treatment facilities. *Proteobacteria*, *Bacteroidetes*, *Chloroflexi*, and *Acidobacteria* are generally found in wastewater treatment systems and are gram-negative (Hu et al. 2012). Meanwhile, the gram-positive isolates may be due to fecal coliforms such as *Enterococcus faecalis* (He et al. 2021).

### Chemical Oxygen Demand Reduction Performance of the AO and AO<sub>2</sub> Systems

The laboratory-scale reactors were used to treat synthetic wastewater. Chemical oxygen demand was measured every day for 10 days using the DNS method. The AO system had a minimum COD reduction rate of 56.20 mg L<sup>-1</sup> h<sup>-1</sup>, a maximum of 60.24 mg L<sup>-1</sup> h<sup>-1</sup>, and a mean of 59.88 mg L<sup>-1</sup> h<sup>-1</sup>. The system also had a minimum removal of 92.10% COD, a maximum of 99.37%, and a mean of 98.13%. The AO system showed performance stability during the 10-day experiment with a decrease in both the COD reduction rate and percent COD removal on the 8<sup>th</sup> day. However, it stabilized again to a comparable COD reduction rate and percent COD removal on the 9<sup>th</sup> and 10<sup>th</sup> day.

Meanwhile, the AO<sub>2</sub> system had a minimum reduction rate of 24.55 mg l<sup>-1</sup> h<sup>-1</sup> COD, a maximum of 39.35 mg L<sup>-1</sup> h<sup>-1</sup>, and a mean of 35.55 mg L<sup>-1</sup> h<sup>-1</sup>. The system also had a minimum removal of 58.92% COD, a maximum of 94.44%, and a mean of 85.33%. The AO<sub>2</sub> system showed performance stability until the 5<sup>th</sup> day. However, it experienced an erratic performance

from the 6<sup>th</sup> day to the 10<sup>th</sup> day, with a characteristic decrease and then an increase in both the COD reduction rate and percent COD reduction. The erratic performance may be due to the change in pH in the reactor from the 6<sup>th</sup> day until the 10<sup>th</sup> day of the experiment.

Overall, the AO system performed better than the AO<sub>2</sub> system in removing COD from the wastewater. The percent COD removal of the AO system as compared to other systems has a comparable percent COD removal. In Table 5, a comparable percent COD removal was observed in the treatment of raw washing wastewater using a sequencing batch biofilter granular reactor with a removal of 95.24% COD and in the treatment of synthetic wastewater using an intermittently aerated and decanted single-reactor process with a removal of 95% COD (Yoo et al. 1999; Altieri et al. 2021). Meanwhile, the AO<sub>2</sub> system has comparable percent COD removal performance with the coagulation-flocculation process, with 82.58% COD removal (Setiawan et al. 2021). Both the AO system and the AO<sub>2</sub> system have significant COD reduction rates compared to other systems (Balku 2007; Munz et al. 2007; Liu et al. 2013; Kim et al. 2019; Nikmanesh et al. 2018; Altieri et al. 2021).

The above results showed the potential of biological treatment for treating plastic recycling wastewater. To further enhance the removal efficiencies of biological treatment systems, it is recommended to examine and optimize the different operating parameters of the systems, such as the pH, the HRT, the dissolved oxygen content, and the influent flowrate of the system. It is also important to upscale the system and evaluate and validate the laboratory results as compared to pilot-scale treatment efficiencies.

### FUNDING

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### ETHICAL CONSIDERATIONS

No human or animals were involved in the conduct of the study.

### DECLARATION OF COMPETING INTEREST

The authors declare that there are no competing interests for any authors.

**Table 5.** Chemical oxygen demand (COD) and percent COD removal of different systems from various sources.

Wastewater	System	COD Reduction Rate (mg L <sup>-1</sup> h <sup>-1</sup> )	Percent COD Removal	Source
Raw washing wastewater	Sequencing batch biofilter granular reactor (SBBGR)	27.00	95.24	Altieri et al. (2021)
Tannery wastewater	Activated sludge process that was enhanced by coagulation and reverse osmosis	123.97	67.00	De Gisi et al. (2009)
Septic tank effluent	Circulating fluidized bed bioreactor	18.23	78.48	Liu et al. (2013)
Polyester synthetic fiber industry wastewater	Coagulation-flocculation	209.42	82.58	Setiawan et al. (2021)
Tannery wastewater	Powdered activated carbon and membrane bioreactors (MBRPAC)	42.92	79.46	Munz et al (2007)
Synthetic wastewater	Intermittently aerated and decanted single-reactor process	12.44	95.00	Yoo et al. (1999)
Synthetic wastewater	Conventional activated sludge system	32.46	70.11	Balku (2007)
Municipal wastewater	Circulating fluidized bed bioreactor	126.29	92.08	Patel et al. (2006)
Actual WWTP influent	Aerobic-anoxic process	10.54	89.53	Kim et al. (2019)
	Activated sludge process	8.86	75.19	
Municipal wastewater	UASB (Upflow anaerobic sludge blanket) - Activated sludge system	61.42	89.67	Von Sperling et al. (2001)
Municipal wastewater	Extended aeration activated sludge system	2.21	61.40	Nikmanesh et al. (2018)
Synthetic plastic recycling wastewater	Anoxic-oxic system	59.88	98.13	This study
Synthetic plastic recycling wastewater	Anoxic-oxic-oxic system	35.55	85.33	

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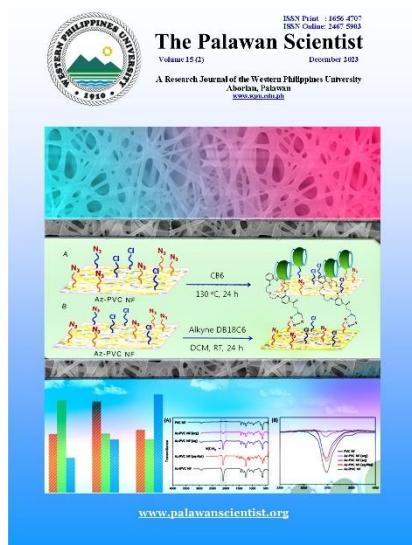
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## Selected physiological requirements of Purple non-sulfur bacteria isolated from Los Baños, Laguna for potential biohydrogen production

Leo Agustin F. Barcelo<sup>1</sup>, Nacita B. Lantican<sup>2</sup>,  
Ruby Lynn G. Ventura<sup>3</sup> and Jey-R S. Ventura<sup>1\*</sup>

<sup>1</sup>Department of Engineering Science, University of the Philippines Los Baños

<sup>2</sup>Institute of Biological Sciences, University of the Philippines Los Baños

<sup>3</sup>UP Rural High School, Los Baños, Laguna

\*Correspondence: [jsventura@up.edu.ph](mailto:jsventura@up.edu.ph)

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### ABSTRACT

Biohydrogen is gaining traction in energy research due to its high energy content and minimal carbon footprint. A typical method of producing biohydrogen is photofermentation using purple non-sulfur bacteria (PNSB). Exploring novel strains of PNSB and studying their versatile metabolism can aid in bioprospecting their potentially valuable by-products and applications, particularly in energy generation. This study investigated the physiological requirements of top biohydrogen-producing PNSB isolated from various Los Baños, Laguna, Philippines sites by measuring their biogas production and growth when subjected to different incubation conditions and macronutrient requirements. Results showed that the three local isolates grown anaerobically in mesophilic conditions without agitation preferred incandescent light. The high biogas yield and growth may be attributed to the light-harvesting pigments in PNSB, which are excited primarily by infrared and near-infrared wavelengths of incandescent light. Furthermore, the isolates can grow from various carbon sources, such as volatile fatty acids (malate, succinate, acetate, butyrate, and propionate) and sugars (glucose and starch). However, two of the isolates (MAY2 and PR2) did not produce biogas when supplied with acetate as a carbon source, which suggests a competing pathway that may have affected the photofermentation of the isolates. Also, the isolates prefer more complex organic sources such as yeast extract and peptone than inorganic sources such as ammonium and less complex organic sources such as urea. Finally, experiments on salt tolerance showed that 0.04% and 0.85% NaCl concentration favors biohydrogen production and growth, as exhibited by high biogas production, yield, and optical density. Results from this study can serve as a basis for future research on optimizing media composition and conditions for biohydrogen production from these isolates.

**Keywords:** biohydrogen, photofermentation, purple non-sulfur bacteria



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## INTRODUCTION

The threat of global warming and the world's energy dilemma have urged experts to find more sustainable energy sources. Along with other alternatives, biobased systems are being explored as replacements for hydrocarbon fuels. Recently, research on biohydrogen has been gaining traction, as it is an efficient energy carrier with around three times more energy than gasoline (EIA 2022) and has negligible greenhouse gas emissions (Kotay and Das 2008). Hydrogen (H<sub>2</sub>) production can be achieved sustainably (Ahmed et al. 2021) through dark fermentation, direct and indirect biophotolysis, and photofermentation. Nevertheless, despite these various biological routes, H<sub>2</sub> technology is limited due to the low efficiency of the production process (Saratale et al. 2019; Tiang et al. 2020).

Photofermentation using purple non-sulfur bacteria (PNSB) is among the widely studied methods for biohydrogen production. Photofermentation can ultimately convert carbon and nitrogen sources to H<sub>2</sub> but is limited by various factors such as light and enzyme activity (Chandrasekhar 2015). In photofermentation, H<sub>2</sub> is generated as a by-product of energy generation via anoxygenic photosynthesis, where electrons from the tricarboxylic acid cycle are subsequently transferred to redox proteins embedded in the intracellular membrane, resulting in a proton gradient that powers the adenosine triphosphate cycle synthase. The nitrogenases and hydrogenases then convert the electrons and the protons to H<sub>2</sub>. Thus, researchers are exploring ways to improve the photofermentation pathway by bio-prospecting novel species, improving strains, and identifying culture conditions that complement the growth requirements of PNSB. Among the growth requirements considered include macronutrients such as carbon and nitrogen sources, which constitute the significant biomolecules in living systems (Bonnet et al. 2020). It is already established that the amount and the type of carbon and nitrogen sources have considerable effects on the H<sub>2</sub> productivity of PNSB, making them primary candidates for optimization studies (Abdullah et al. 2020; Hakobyan et al. 2019). Furthermore, the H<sub>2</sub> photofermentative pathway requires light as an energy source, and light's spectrum and intensity have significant effects on H<sub>2</sub> productivity (Bosman et al. 2023; Rashid et al. 2022). Hence, the light requirement of the isolates is one essential consideration for optimization. On the other hand, there are known salt tolerant PNSBs studied mainly for their ability to utilize waste and wastewater as substrates (Chen et al. 2020; Hülsen et al. 2019). Plans for hybrid setups of wastewater treatment and photofermentation plants could benefit from salt tolerant PNSB isolates. There are limited local studies on the isolation of purple non-sulfur bacteria. One of which is by Montano et al. (2009), which isolated PNSB from rice paddies in

Bulacan. Another local study is by Del Socorro et al. (2013), which isolated PNSB from rice paddies and aquatic sediments in Iligan City. These studies, however, have no emphasis on the biohydrogen production capability of the isolates and are, therefore, currently untapped in the Philippines.

This study characterized selected physiological requirements of the top biogas producers from the pool of PNSB isolates from various areas in Los Baños, Laguna. Specifically, this study investigated the effects of light (blue light-emitting diode (LED), red LED, fluorescent and incandescent), carbon (volatile fatty acids (VFAs), succinate, malate, glucose, and starch), nitrogen (ammonium, urea, glutamate, peptone, yeast extract), and NaCl concentrations (0.04%, 0.85%, 3.0%, 10.0%) on the biohydrogen production potential of the isolates. Studies suggest that the gas generated by PNSB photofermentation is composed primarily of H<sub>2</sub> (Craven 2019; Turon 2018; Ventura et al. 2019). Thus, biogas was measured to approximate biohydrogen production potential

## METHODS

### Purple Non-sulfur Bacteria Isolates

The top three biohydrogen-producing strains (MAY2, IRR1, PR2) out of the 19 PNSB isolates obtained from water sediments in Los Baños, Laguna, Philippines (Ventura et al. 2021) were the focus of this study. MAY2 and PR2 isolates were identified under the genus *Rhodobacter*, and IRR1 was verified to be under the genus *Rhodopseudomonas* using 16s rRNA sequencing.

### Main Culture Preparation

The main cultures of these isolates were revived by multiple subsequent transfers in an acetate-yeast extract medium (AYE) (Montano et al. 2009). This medium has the following components (in g L<sup>-1</sup>): K<sub>2</sub>HPO<sub>4</sub>, 1.0; MgSO<sub>4</sub>·7H<sub>2</sub>O, 0.2; CaCl<sub>2</sub>·2H<sub>2</sub>O, 0.02; Na<sub>2</sub>S<sub>2</sub>O<sub>3</sub>, 0.10; NaCH<sub>3</sub>COO, 2.2; and yeast extract, 4.0. A loopful of each culture from stock was streak-plated in Acetate Yeast Extract (AYE) agar. Cultures were incubated under incandescent light using an anaerobic set-up previously described by Maiti et al. (2013). The isolate grown from this stage was then used for the succeeding experiments.

### Seed Culture Preparation

Modified Biebl and Pfennig (MBP) medium was used for the seed culture preparation. The medium contains the following components (in g·L<sup>-1</sup>): MgSO<sub>4</sub>·7H<sub>2</sub>O, 0.2; NaCl, 0.4; KH<sub>2</sub>PO<sub>4</sub>, 0.5; CaCl<sub>2</sub>·2H<sub>2</sub>O, 0.05; ferric citrate, 0.005; yeast extract, 0.3; vitamin solution (nicotinic acid, 0.0002; nicotinamide, 0.0002; thiamine HCl, 0.0004; and biotin, 0.008); and 1 ml trace element solution. Trace element solution

has the following composition (in mg·L<sup>-1</sup>): ZnCl<sub>2</sub>, 0.07; H<sub>3</sub>BO<sub>3</sub>, 0.06; MnCl<sub>2</sub> · 4 H<sub>2</sub>O, 0.1 mg; CoCl<sub>2</sub> · 2 H<sub>2</sub>O, 0.2; CuCl<sub>2</sub> · 2 H<sub>2</sub>O, 0.02; NiCl<sub>2</sub> · 6H<sub>2</sub>O, 0.02; (NH<sub>4</sub>)<sub>2</sub>MoO<sub>4</sub> · 2 H<sub>2</sub>O, 0.04 and HCl, 0.025% v/v. The pH of MBP was adjusted to pH 6.5, particularly when supplied with volatile fatty acids as the carbon source.

Before the biogas production set-up, a pre-activated seed culture was prepared. A loopful of each isolate from the AYE medium was grown in a 40 ml MBP medium supplied with 7 mm malate and 10 mm glutamate in 150 ml glass vials. The set-ups were then purged with argon gas for 3 min and crimp sealed to simulate anaerobic conditions, vortex-mixed, and incubated under incandescent light (300 nm to 1400 nm) for 3-5 days at room temperature (20-25°C). Growth from the pre-activated cells was then harvested via centrifugation (4000 x g for 5 min) and then transferred to a fresh batch of MBP medium supplied with 7.0 mm malate and 10 mm glutamate. Their optical density (OD) was adjusted to 0.5. Four milliliters of the adjusted culture were then transferred to 36 ml of fresh MBP medium; they were purged with argon and crimp-sealed, vortex-mixed, and then incubated for three days as previously described.

**Biogas Production Set-up**

Four milliliters from the adjusted culture was collected and centrifuged. The pellet was then placed in 150 ml serum bottles with 40 ml fresh MBP. The setup was purged with argon gas for 3 min and then crimp-sealed. The reactor bottles were incubated without agitation under anaerobic conditions at room temperature.

**Effect of Different Light Source**

The biogas production set-up used for the investigation of the effect of light source utilized MBP with mixed acids - 15 mm acetate, 7.5 mm butyrate, and 10 mm propionate (approximately 2.0 g·L<sup>-1</sup> in total), and 2.0 mm glutamate as nitrogen source. The set-ups were incubated under different light types: red LED (660 nm), blue LED (460 nm), incandescent light (300-1400 nm), and fluorescent light (400-700 nm). The light intensity of each light source was also measured using a lux meter (see Table 1).

**Effect of Carbon and Nitrogen Sources**

Using the previously mentioned biogas production setup, the capability of the three isolates to

utilize different carbon sources was also investigated. Instead of mixed acids, the set-ups used individual carbon sources- malate, succinate, acetate, butyrate, propionate, glucose, and soluble starch. The concentration of each carbon source was 2.0 g L<sup>-1</sup> (Basak and Das 2007; Assawamongkholsiri 2019).

Similarly, the nitrogen requirements of the isolates were also investigated using the same biogas assay setup with modification on the nitrogen source. The nitrogen sources investigated were ammonium, urea, glutamate, yeast extract, and peptone. 2.0 mM of ammonium, glutamate, and urea were used for MBP (Ventura et al. 2021), while 2.0 g L<sup>-1</sup> concentration was used for peptone and yeast extract (Hakobyan et al. 2012). The setups were processed and incubated as previously described.

**Effect of Salt (NaCl) Concentration**

The isolates were subjected to different salt (NaCl) concentrations by adjusting the amount present in the MBP medium. The variation in salt concentrations was as follows: 0.04% (recommended amount in standard MBP medium), 0.85% (physiological saline), 3.0% (marine saline concentration), and 10% (hypersaline) (Xiao et al. 2022; Irwin 2020). The MBP medium used in the investigation of the effect of NaCl concentration was similar to the light source investigation setup. The setups were processed and incubated as previously described.

**Measurement of Biogas Production and Optical Density**

The cumulative biogas was measured by recording the displaced gas using a sterile syringe every 24 hours, and the cumulative readings were plotted. Biogas yield was calculated according to the method of Wu et al. (2016; equation 1).

The sampling for biogas assay in the investigation of light source ran for 20 days; while that of carbon and nitrogen sources and salt tolerance ran for ten days.

The optical density in each setup was measured on the last day of observation to approximate cell growth. Two ml of the culture broth was placed in disposable cuvettes, and optical density was measured using a Shimadzu® 1800 UV-Vis spectrophotometer set at 660 nm.

$$\text{Biogas yield ml g}^{-1} = \frac{\text{Total biogas generated per set-up (ml)}}{\text{Volume of media (ml)} \times \text{substrate concentration (g \cdot ml}^{-1})} \quad \text{equation (1)}$$

**Table 1.** Light sources used in the experiment.

Light type	Brand/Model	Wattage (W)	Illuminance (lux)	Wavelength (nm)	Intensity (lux)
Incandescent	Firefly® FINS60F	60	1000	300-1400	~700
Fluorescent	Firefly® 3S32	32	1000	400-700	~3,200
Red LED	Perfin® PFLG02	10	200	660 nm	~200
Blue LED	Perfin® PFLG02	10	300	460	~200

**RESULTS**

**Effect of Different Light Source**

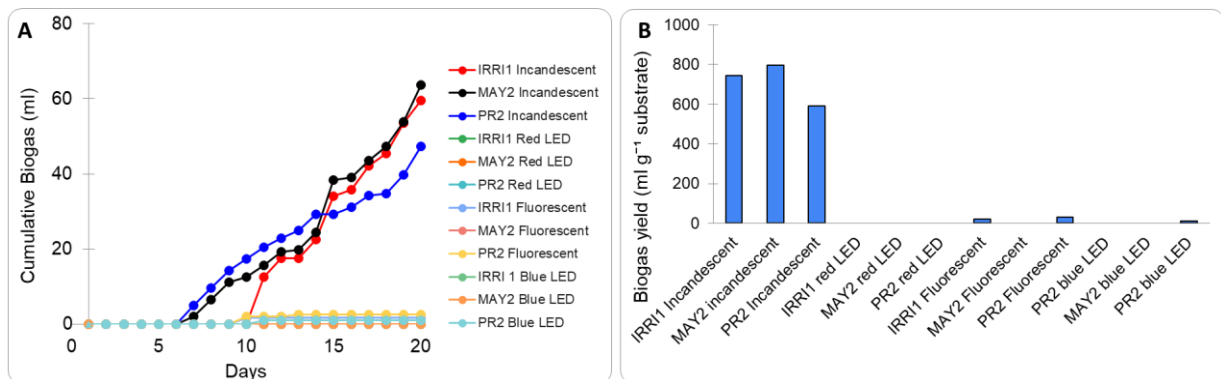
The isolates MAY2, PR2, and IRR11 prefer incandescent light for biogas production. As shown in Figures 1A and 1B, the highest biogas production (in ml) of the three isolates was recorded in setups incubated under incandescent light. MAY2, IRR11, and PR2 produced 63.6 ml (795 ml·g<sup>-1</sup> substrate), 59.6 ml (745 ml·g<sup>-1</sup> substrate), and 47.4 ml (592.5 ml·g<sup>-1</sup> substrate) biogas, respectively. Incubation under different light sources produced minimal to almost negligible amounts of biogas. Setups incubated under red LED light did not produce biogas, while those exposed to blue LED and fluorescent light produced minimal amounts only (< 5.0 ml). Optical density was found to have the highest biomass in set-ups exposed to fluorescent and incandescent light (Table 2). As listed in Table 2, the highest OD recorded for setups under incandescent light were 1.667 and 1.574, respectively, for PR2 and MAY2, while IRR11 gave the highest OD (1.446) for the reactors incubated under fluorescent light. The highest recorded OD for IRR11 was 1.446 in fluorescent light setups, while the lowest OD values (< 0.5) were recorded in setups exposed to red LED.

**Carbon and Nitrogen Requirements**

The isolates were observed to variedly utilize different carbon sources (Figures 2A and 2B). For MAY2 and PR2, the highest biogas production was

recorded in butyrate at 67.6 ml (845 ml·g<sup>-1</sup> substrate) and 55.0 ml (687 ml·g<sup>-1</sup> substrate) of biogas, respectively. On the other hand, IRR11 supplied with acetate produced 62.7 ml of biogas (781.25 ml·g<sup>-1</sup> substrate). Interestingly, PR2 and MAY2, which both belong to the genus *Rhodobacter*, did not produce biogas in acetate. These two isolates were also able to produce gas when supplied with glucose, while IRR11 did not. All the isolates did not produce biogas when supplied with soluble starch. In terms of OD, low OD was observed in setups supplied with succinate, as shown in Table 3. The highest recorded OD was in IRR11 supplied with butyrate at 2.091. The highest OD for MAY2 and PR2 were also recorded in setups supplied with butyrate at 1.817 and 1.419, respectively.

While there was an observed variation in the utilization of carbon sources for the three isolates, a similarity was observed in the preference for nitrogen sources (Figures 3A and 3B). The three isolates produced biogas when supplied with glutamate, yeast extract, and peptone. However, they did not yield gas in ammonium (< 0.5) and urea setups. The three isolates exhibited the highest biogas production in setups supplied with yeast extract: MAY2, PR2, and IRR11 produced 132 ml (1650 ml·g<sup>-1</sup> substrate), 107.3 ml (1446.25 ml·g<sup>-1</sup> substrate), and 88.5 ml (1106.25 ml·g<sup>-1</sup> substrate) of biogas, respectively. The highest OD values were also observed in yeast extract set-ups of MAY2 and PR2 (Table 4). The highest OD value for IRR11 was recorded in set-ups supplied with peptone.

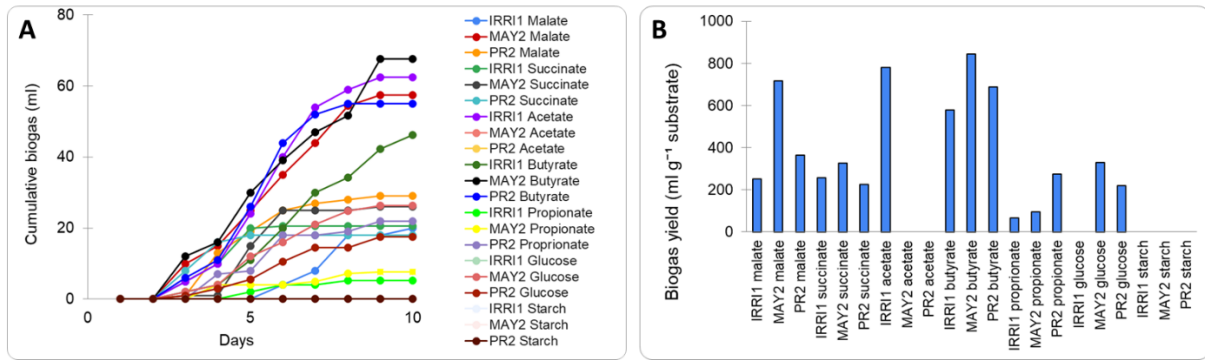


**Figure 1.** (A) Cumulative biogas production and (B) biogas yield of the purple-non-sulfur bacteria isolates incubated under different light conditions.

**Table 2.** Optical density at 660 nm (OD<sub>660</sub>) measurement of purple non-sulfur bacteria (PNSB) isolates set-ups incubated under different light types.

Isolate	Light Type			
	Incandescent	Fluorescent	Red LED	Blue LED
IRR11	1.011	1.446	0.294	1.183
MAY2	1.574	1.002	0.613	0.977
PR2	1.667	1.651	0.363	0.965

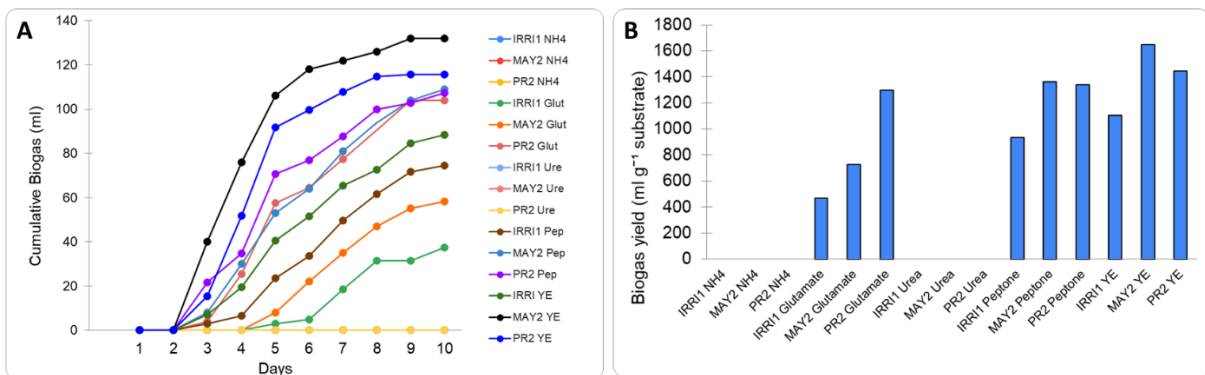




**Figure 2.** (A) Cumulative biogas production and (B) biogas yield of the purple non sulfur bacteria isolates under different carbon sources.

**Table 3.** Optical density at 660 nm (OD<sub>660</sub>) measurement of the purple non-sulfur bacteria (PNSB) isolates set-ups incubated across different carbon sources.

Isolate	Carbon source						
	Malate	Succinate	Acetate	Butyrate	Propionate	Glucose	Starch
IRR11	0.604	0.736	1.329	2.091	1.569	0.459	1.693
MAY2	1.323	0.717	1.417	1.817	0.736	1.645	1.485
PR2	0.814	0.712	1.408	1.419	0.514	1.225	1.114



**Figure 3.** (A) Cumulative biogas production and (B) biogas yield of the purple non sulfur bacteria isolates under different nitrogen sources.

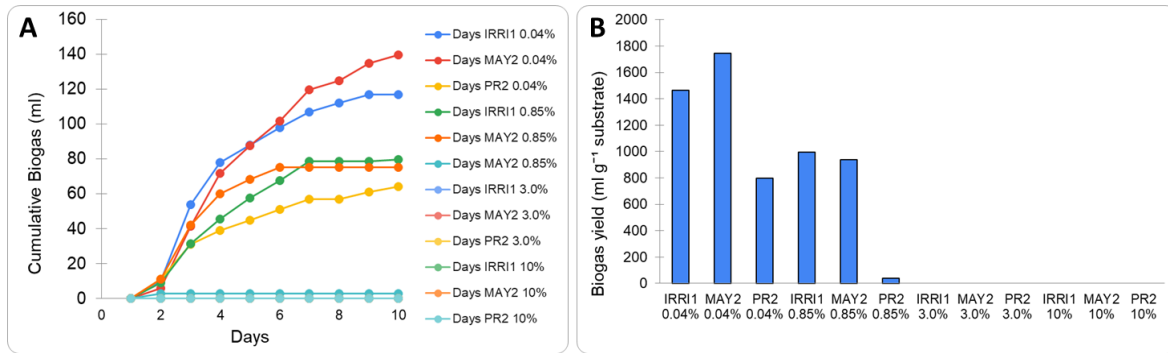
**Table 4.** Optical density at 660 nm (OD<sub>660</sub>) measurement of the purple non-sulfur bacteria (PNSB) isolates set-ups incubated across different nitrogen sources.

Isolate	Nitrogen Source				
	NH <sub>4</sub>	Glutamate	Peptone	Yeast Extract	Urea
IRR11	0.534	0.598	0.765	0.614	0.398
MAY2	0.170	0.561	0.449	1.208	0.973
PR2	0.205	0.940	0.431	1.133	0.823

**Effect of Salt (NaCl) Concentration**

Thus, the salt tolerance of the isolates was also investigated. The results in Figures 4A and 4B showed that the isolates produced the highest amount of biogas in set-ups supplemented only with 0.04% NaCl. This concentration is the standard NaCl concentration of the MBP medium. MAY2, PR2, and IRR11 produced 139.6 ml (1745 ml·g<sup>-1</sup>substrate), 64 ml (800 ml·g<sup>-1</sup> substrate), and 117 ml (1462.5 ml·g<sup>-1</sup> substrate) of biogas, respectively. Adjusting the salt concentration to 0.85% (physiological concentration)

decreased the production of biogas. Moreover, set-ups supplemented with 3.0% and 10% NaCl did not produce biogas. As shown in Table 5, cell growth is also affected by the salt concentration, as OD also decreased with the increase in NaCl. A 0.85 salt concentration exhibited the highest OD for MAY2 (1.129), followed by PR2 (0.936). IRR11, on the other hand, produced the highest OD (0.949) at 3.0% NaCl concentration. The 10% salt concentration barely supported the cell growth of the isolates, as evidenced by an OD below 0.5.



**Figure 4.** (A) Cumulative biogas production and (B) biogas yield of the purple non-sulfur bacteria isolates at varying salt concentrations.

**Table 5.** OD<sub>660</sub> measurement of the purple non-sulfur bacteria isolates set-ups incubated under different salt concentration.

Isolate	Salt Concentration			
	0.04%	0.85%	3.0%	10%
IRRI1	0.798	0.872	0.949	0.223
MAY2	1.020	1.129	0.836	0.373
PR2	0.877	0.936	0.690	0.141

## DISCUSSION

### Biogas Production and Optical Density of the PNSB Isolates under Different Light Sources

Purple non-sulfur bacteria produce light-harvesting complexes (LHC) such as LH-I and LHII, which contain pigments that absorb far-red and near-infrared spectrum (Law et al. 2004). These pigments are mostly carotenoids and bacteriochlorophylls, which typically absorb light at 500 nm and above 800 nm, respectively. The absorbed light provides energy for photofermentation, which is the main pathway that generates H<sub>2</sub> (Deo et al. 2012). Incandescent lamps emitting an infrared spectrum (800 nm – 500 μm) were found to be the best option for H<sub>2</sub> production for *Rhodobacter capsulatus* (Monroy et al. 2013; Fox 2020). Furthermore, H<sub>2</sub> productivity is improved by the presence of both visible (400 nm – 750 nm) and infrared spectra provided by incandescent lamps (Turon et al. 2018; Halabe 2013). Therefore, the broad wavelength spectra of incandescent light provide the wavelength that complements the light absorption requirements of LHCs present in most PNSB. Observations from previous studies support the observed preference of the PNSB isolates in this study. For example, a study on *Rhodobacter capsulatus* found that replacing incandescent lamps with LED that only emit near-infrared light reduced H<sub>2</sub> production by 50% (Turon et al. 2018). In another study by Hu et al. (2018), the cell growth rate and H<sub>2</sub> production of four different species of PNSB were significantly higher when exposed to incandescent light systems compared to fluorescent. In contrast, LEDs and fluorescent are known to emit limited colored spectra (Abdel-Rahman et al., 2017), which may not provide the infrared and

full visible spectra needed for optimized photofermentation, possibly limiting H<sub>2</sub> production. It is important to note, however, that light requirements for biohydrogen production may vary between species and can be affected by factors such as intensity and exposure time (Nath and Das 2009; Androga 2012). Additionally, the biogas production setups in the study took a more prolonged time (6 days) to produce biogas than what was usually observed (3-4 days). So, the 10-day observation period was extended to 20 days.

Based on the optical density results listed in Table 2, all the light source types used in the experiment can support the growth of the three isolates, but the cell growth of each isolate varies according to the type of source. For MAY2, incandescent light is favorable for biomass production, as evidenced by the high OD (1.574). For IRRI1, fluorescent light (1.446), while for PR2, both fluorescent and incandescent lights support high cell growth (1.667 and 1.651, respectively). Overall, many studies suggested that incandescent light best suits PNSB cultivation (Yu et al. 2021; Yu et al. 2022). For example, a study on *Rhodobacter sphaeroides* found that incandescent light generated the most biomass (5.66 g DCW L<sup>-1</sup>) and protein (4.43 g L<sup>-1</sup>) after 7-day cultivation compared to halogen lamp, infrared light, and variously colored LEDs. Full-spectrum incandescent light was also found to produce 3.2 times more biomass than spectral bands incandescent light (Yu et al. 2022). However, some studies focused on using other light types depending on the intended output or product. For example, Hülsen et al. (2019) utilized infrared light to selectively enrich PNSB in a non-sterile-rich medium for COD, NH<sub>4</sub>-N, and PO<sub>4</sub> removal. In a study by Kuo et al. (2012), LED blue

light produced the highest cell and carotenoid concentration in cultures of *Rhodospseudomonas palustris* compared with incandescent, fluorescent, and other colors of LED lights. In a different study by Zhuo et al. (2014), red LED light was found to support higher biomass in production and COD removal of locally isolated *Rhodospseudomonas* compared with other sources of light. Moreover, the highest production of carotenoids was observed under yellow LED light. These studies suggest that blue and red light generally supports cell growth, although the optimum effect on productivity can be species-specific (Rashid et al. 2022).

These results suggest that higher biomass does not automatically mean higher biogas or H<sub>2</sub> production, and different light sources can support biomass growth but may not favor photofermentative biohydrogen production.

### Biogas Production and Optical Density of the PNSB Isolates across Different Carbon and Nitrogen Sources.

The metabolic flexibility of PNSB allows them to assimilate different carbon sources. They are mainly studied for their capability to use volatile fatty acids as carbon sources for H<sub>2</sub> production. Most species of *Rhodobacter* prefer malate, succinate, butyrate, propionate, lactate, and pyruvic acid for H<sub>2</sub> production (Androga et al. 2012). This trend reported in the literature was also observed in the high H<sub>2</sub> productivity of MAY2 and PR2, both members of *Rhodobacter*. This preference can be due to the direct assimilation of the short-chain organic acids to the TCA cycle, which primarily provides electrons and protons to the photofermentation pathway. As the electrons are transferred in the pathway, a proton gradient is created, which drives the ATP synthase to generate energy, after which the protons and electrons are captured by nitrogenase to produce H<sub>2</sub> (Gabrielyan et al. 2015). Nonetheless, these isolates did not produce gas in acetate, which could be attributed to other competing pathways, such as the polyhydroxybutyrate (PHB) synthesis. In some *Rhodobacter* species, acetate is preferentially utilized for PHB synthesis, a pathway that competes with photofermentation for electrons obtained from the oxidation of VFAs. This relationship between the pathways puts a limit on how *Rhodobacter* can utilize acetate. *Rhodospseudomonas* members can utilize various VFAs with efficient acetate utilization (Oh et al. 2004; Touloupakis et al. 2021).

Although gas production was observed in glucose for MAY2 and PR2, the amount of gas produced was relatively lower compared to the VFA setups; starch, likewise, did not fare better. For PNSB to utilize glucose, it must be converted first to VFAs, resulting in inadvertent lowering of pH and a longer time in assimilating the substrate (Jeong et al. 2008). Starch also requires acid or enzymatic hydrolysis for it

to be utilized by photosynthetic bacteria for photofermentation (Vendruscolo 2015).

The observed high biogas productivity of the isolates in setups supplied with yeast extract as a nitrogen source can be attributed to other micronutrients present in yeast extracts, such as amino acids, vitamins, minerals, and growth-stimulating compounds (Gabrielyan et al. 2015; Hay et al. 2013). These micronutrients may serve as co-enzymes and co-factors to the enzymes found in the photosynthetic electron transfer chain of the photofermentation pathway, as well as cofactors to the nitrogenases and hydrogenases. A study by (Hakobyan et al. (2012) compared the H<sub>2</sub> production of *Rhodobacter sphaeroides* supplied with yeast extract and glutamate as a nitrogen source. Yeast extract setups produced 6 times more H<sub>2</sub> than glutamate set-ups, which is attributed to the presence of growth-stimulating compounds, vitamins, and a variety of amino acids present in yeast extract. In another study by Liu et al. (2015), a novel strain of *Rhodospseudomonas* was able to grow and produce H<sub>2</sub> in setups supplied with peptone, beef extract, and glutamate but not in urea and ammonium. These results are similar to the obtained results of this study. The negligible production of biogas in setups supplied with ammonium can potentially be due to the inhibitory effect on ammonium ions to the nitrogenase enzyme, which either repressed the enzyme via feedback inhibition or by repression of nitrogenase-related genes (Androga et al. 2012). Similarly, urea can be broken down into carbamate and ammonium ions, possibly rendering the same effect (Alexandrova and Jorgensen 2007).

### Salt (NaCl) Tolerance of the PNSB Isolates

Some PNSB members were found to withstand high salt concentrations even in seawater and hypersaline environments (Sakarika et al. 2019). Based on the results, high salt concentration halted the H<sub>2</sub> production of the isolates, although they were able to exhibit cell growth at 3.0% salt concentration. Many PNSB members were able to tolerate saline conditions, including species of *Rhodobacter* and *Rhodospseudomonas* (Asif et al. 2021). A study on one marine PNSB, *Rhodovulum sulfidophilum*, observed that biohydrogen production was still attained even at 3.0% salt concentration, and the optimum production is at 2.0% concentration (Cai and Wang 2012). However, for the isolates in this study, low NaCl concentration is still preferred for biogas production. These isolates were not obtained in saline or hypersaline environments and may not have the necessary adaptations for such conditions. Moreover, studies point out that increasing salinity can decrease the activity of the nitrogenase enzyme, which is the primary enzyme for photobiological hydrogen production (Severin et al. 2012).

## Recommendations and Future Directions of the Study

Results from this study can serve as a baseline for future optimization experiments, as these results can narrow down the criterion and parameters of interest. Parameters such as carbon-to-nitrogen ratio and light intensity can also be explored. Furthermore, these can also be used for proper characterization and identification of the isolates.

The study relied only on crude measurement of biogas via displacement, that is why it is recommended to measure biohydrogen content and volatile fatty acid content via chromatography in future studies. Additionally, it is recommended to upscale the reaction and validate if it is comparable to the results from this study or other studies.

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## ETHICAL CONSIDERATIONS

There were no animal or human subjects involved in the study.

## DECLARATION OF COMPETING INTEREST

The authors declare that there are no competing interests of any authors.

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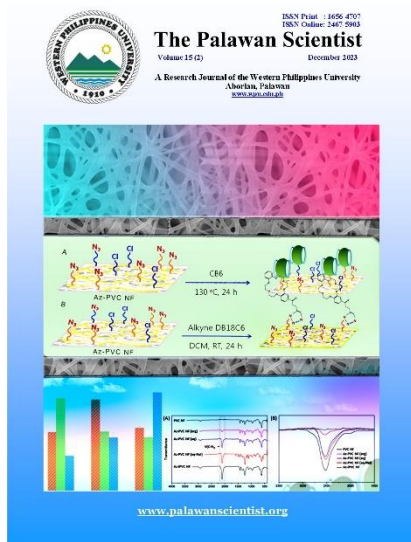
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- ROLE OF AUTHORS:** LAFB – data collection, data gathering, data analysis, manuscript writing, and editing; NBL – supervision, manuscript writing, and editing; RLGV – supervision, manuscript writing, and editing; JSV – conceptualization, project administration, supervision, manuscript writing, and editing.



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# Comparative analysis of unidirectional and bidirectional electric vehicle charging stations (EVCS) optimal configuration in an IEEE 37-bus feeder system using Genetic Algorithm

Aurelio A. Balmeo, Jr. \*, Rodolfo A. Aguirre, Jr.,  
 Ma. Danica G. Castillo, Edward Joseph H. Maguindayao  
 and John Paul P. Manzano

*Department of Electrical Engineering, College of Engineering and Agro-Industrial Technology, University of the Philippines Los Baños, Laguna, Philippines*

\*Correspondence: [aabalmeo@up.edu.ph](mailto:aabalmeo@up.edu.ph)

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## ABSTRACT

Various power system problems and challenges may arise in the future due to the large scale of deployment of electric vehicles (EVs). Hence, the proper placement of EV charging stations (EVCS) effectively mitigates the impact of high EV loads connected to the grid. The research intends to explore and analyze differences between the regulation effectiveness of unidirectional and bidirectional charging technologies by utilizing different comparison evaluation indices. Moreover, considering their penetration level, this study tackles the impact analysis of EV and EVCS integration through time. Specifically, this paper aims to identify the optimal EVCS sites in an IEEE 37-bus test feeder system to minimize power loss brought by EV integration. Through MATLAB R2022b simulation and OpenDSS power flow analysis, the EVCS are optimally located near the supply bus. The findings show a direct relationship between the EV penetration level and system power loss. Due to the EV technology growth, there is an observed voltage profile degradation of up to 1.7094 p.u. The paper also highlights that although EV bidirectional charging technology (BCT) might reduce the load on the grid in the next few years of low penetration compared to unidirectional charging technology (UCT), it will give no significant difference due to the rapid increase of load connected during its high EV penetration.

**Keywords:** electric vehicles, MATLAB, optimization, power loss reduction, voltage profile improvement

## INTRODUCTION

Today's global challenges include reducing carbon footprints and mitigating energy risks. True enough, these will eventually become huge problems if not prevented. Thus, modernizing the transportation mode to electric vehicles (EV) is eyed to cut down the usage of internal combustion engines (ICEs)—the primary contributor to problems in energy security, air

pollution, and global warming (Jacobson 2017). Sofana Reka et al. (2022) proved in their study that EVs have an enormous impact on the environmental aspects of ICEs, in particular with the emissions of CO<sub>2</sub> gas and maintenance costs. Meanwhile, EVs make a way towards sustainability by reducing greenhouse gas (GHG) emissions and fossil fuel consumption (Bayani et al. 2022).



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However, other works disagree on the drive range aspect of EVs since they are unsuitable for long-distance travel and ownership cost is currently less cost-competitive (Sofana Reka et al. 2022; Bayani et al. 2022; Danielis et al. 2018). At present, three main charging methods have been conceptualized and developed: battery exchange, conductive charging, and wireless charging (Arif et al. 2021). EV users find the battery swap station (BSS) technique a convenient option since they will only pay a monthly rental fee for quick battery swapping for their convenience. However, this technique requires users to have high payments for the BSS owner to utilize high-end batteries with a long life span and can be recharged multiple times (Gschwendtner et al. 2021; Brenna et al. 2021). Meanwhile, the wireless power transfer (WPT) differs among the three, for it can recharge the battery conveniently and safely. Even so, due to the weak inductive power and a large amount of eddy current loss, it is not advisable in the long run (Sanguesa et al. 2021). The most widely accepted charging method is the conductive charging (CC), which has a high charging efficiency through direct connection. Moreover, a vehicle-to-grid (V2G) facility can be a way to incentivize compensation, lessen losses, and prevent power grid overload (Arif et al. 2021).

Aside from the charging methods, charging modes are continuously being improved to attain charging speeds comparable to gas refueling for ICEs. Currently and generally, there are three different EV charging levels, as defined by Narasipuram and Mopidevi (2021). In the level I charging, there is an AC to DC converter integrated inside the car and charges within the current range of 15–20 A at 120 V. Research found that the ratio of driving-mile distance per charging hour is 2:1. There have been many innovations over time. Level II charging station has charging characteristics of 80 A at 240 V, which can have a 9–52-mile travel distance after an hour of charging. Meanwhile, the level III charging provides DC power and is already attached to the station. It gives 300 A current flow at 480 V, which can have 170 miles of travel in just thirty minutes of charging. However, type III is only compatible with a few EV models, leaving the rest to choose between types I and II charging. For the purpose of this research, the paper utilizes the level II type of charging. The charging level of this type is the most preferable among all the abovementioned charging types since it enables sufficiently long-distance trips and charges an EV at a reasonable length of time (Lee et al. 2020).

Conversely, the switch to vehicle modernization through EVs also creates another issue requiring plenty of accessible EV charging stations (EVCS) to recharge their batteries, as ICE vehicles need gas stations for refueling purposes. Thus, EV owners expect to have available EVCS to charge their vehicles quickly and hassle-free, maximizing the growth of this particular technology. However, this

innovation might challenge government policies on how to steer the market toward full electrification in transportation despite its growing disadvantages (Sofana Reka et al. 2022).

Even though it gives users convenience, large-scale EVCS deployment negatively impacts the power grid, such as transformer overloading and power quality degradation (Zhou et al. 2017). This impact is especially true given that the distribution systems are reaching their maximum capacities. This anticipated load growth, which is dynamic and highly intermittent, would be a challenging job for the electric power sector (Gupta et al. 2020). Implementing countermeasures, such as appropriate siting of EVCS and developing coordinated unidirectional and bidirectional charging types, plays a crucial role in effectively reducing the load impacts in the distribution network (Zheng et al. 2019).

Identifying suitable EVCS location in the distribution system is fundamental for maintaining the balance between load and generation, reducing power losses, and improving system stability (Rajendran and Kumar 2022). If not arranged properly, this will significantly affect the grid, including voltage fluctuations, flickering, sag, swell, imbalances, harmonics, and notches (Narasipuram and Mopidevi 2021). Researchers have been working on this issue, exploring the optimal location of the EVCS by delving into different perspectives and patterns necessary to real-life situations. Incorporating influencing parameters such as system buses and parking availability plays a significant role in solving the optimal EVCS problem (Zeb et al. 2020). Previous studies have analyzed different power systems such as in the work of Yenchachalit et al. (2018), which utilized the IEEE 30-bus test system to determine the trend in power loss with and without an EVCS installed. Moreover, Clairan et al. (2022), an actual distribution system in Quito, Ecuador, to determine the effect of increasing electric taxi penetration; while Janamala (2022) investigated the optimal siting of EVCS considering EV load growth in IEEE 33-, 69-, and 85-bus systems.

Various emerging optimization methods are consistently used to compare power distribution and transmission system performances. For example, the hybrid technique consisting of moth-flame optimization (MFO) algorithm and particle swarm optimization (PSO) has been done in the paper of Shaikh et al. (2023) to examine the advantages of using different numbers of bundled conductors. On the other hand, the improved MFO has been applied to AC transmission line estimations considering different test case parameters (Shaikh et al. 2022). In calculating the transmission line parameters with load modeling uncertainty, Shaikh et al. (2021) used the whale optimization algorithm (WAO).

To determine the optimal EVCS location in the power network, a balanced mayfly algorithm was



utilized to identify the best proposal for optimal allocation and sizing on a distribution system in India (Chen et al. 2021). Another literature used the Harris hawk optimization with differential evolution to solve the optimization problem in the EVCS allocation in a radial distribution network (Pal et al. 2021). The placement of EVCS in distribution systems, shunt capacitors, and distributed generators in the work of Gampa et al. (2020) used a grasshopper optimization algorithm to improve different electrical power system parameters. Lazari and Chassiakos (2023) used a genetic algorithm (GA) to minimize the overall cost of deploying the charging networks. Meanwhile, a modified GA in another study developed a mathematical program with equilibrium constraints (MPEC) in EVCS location and its discrete transport network designs (Qiao et al. 2023).

Determining possible locations not only by EVCS but also with distributed PV stations through GA was used based on chance-constrained programming (Zhang et al. 2021). Also, GA was utilized for efficient EVCS placements, taking Tunisia's urban area in North Africa as the research locale (Mehouachi et al. 2022). Nonetheless, the above-cited works of literature did not use any statistical test but instead focused on the objective function used in their respective research.

Most of these studies have only considered the unidirectional charging type, where EVs can only draw power from the grid. Bidirectional charging, as the improved type, which enables energy stored in EV batteries to discharge back to the grid, has received little attention. With the expectation that more EVs capable of bidirectional charging in the future, investigating its effect on the present distribution system is crucial in anticipating its benefits and adverse impacts (Isa et al. 2015).

Against this backdrop, this paper is primarily focused on the following aspects: (1) the investigation of the effects of new charging technologies on the power losses and voltage profiles with increasing penetration levels and its compatibility with the grid through EVCS modeling; (2) the optimal siting of EVCS to avoid negative impacts on the grid, particularly in the minimization of the additional system losses in the IEEE 37-bus test system; and (3) the comparison of the effectiveness of unidirectional and bidirectional charging types using four evaluation indices: voltage profile improvement index (VPPI); real power loss reduction index (PLRI); reactive power loss reduction index (QLRI); and apparent power loss reduction index (SLRI). Cover cost-effectiveness in locating the EVCS in the test system is not covered in this paper.

The main contributions of this paper are as follows: (1) the increasing penetration levels will give theoretical trend for the simulation of EVCS optimal configuration between the two charging technologies to differentiate them in terms of power losses and total

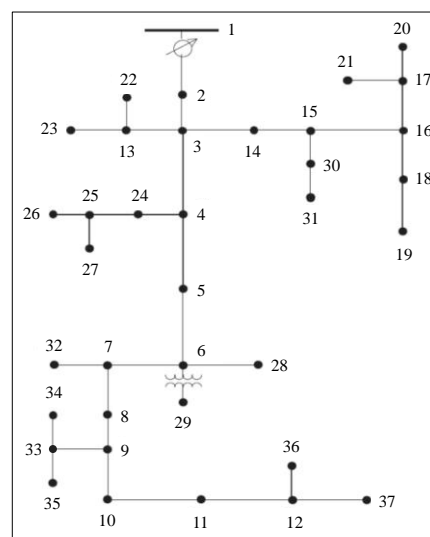
voltage deviation; (2) for the energy department of the country to pinpoint through genetic algorithm where in the bus system should the EVCS be located to obtain the minimized power losses in the grid; (3) for the policymakers to gain insight on the difference between the two charging technologies in terms of the charging behavior, and to decide on the EV charging policies that can help in grid load management.

The rest of the research paper is structured as follows: Section 2, the methods, explains the profile of the feeder, the selection method, EV specifications and modeling, the optimal model and algorithm implementation, the model construction, and indices for evaluation; Section 3, the results, includes reflecting the effect of the total integration of the new technology in increasing penetration levels, the EVCS optimal distribution configuration through genetic algorithm optimization, and charging technology comparison; and Section 4, the discussion, tackles the implications of the total integration of the advancing technology, the causes of the EVCS optimal arrangement, and the prospects, conclusions, and implications of charging implementation of unidirectional and bidirectional charging technologies.

## METHODS

### Institute of Electrical and Electronics Engineers (IEEE) 37-bus Test Feeder Profile and Optimal Location Selection

This research considered the medium voltage IEEE 37-bus test system. It has highly unbalanced load characteristics, differentiating it from other feeder systems. Shown in Figure 1 is the single-line diagram of the renumbered IEEE 37-bus feeder system adopted from the work of Miras et al. (2019).



**Figure 1.** Single line diagram of the renumbered Institute of Electrical and Electronics Engineers (IEEE) 37-bus test system (Miras et al. 2019).

Different approaches can be used to select the optimal locations in a test bus system. In studying bus feeders, the researchers can include a constraint in geographical location employing divisions into different clusters or zones. Considering the whole network is the best option for this paper to achieve the optimal EVCS configuration. Thus, it will exclude the geographical location aspect as a constraint. Moreover, this assessment focuses on taking the feeder as a pure distribution system, minimizing power loss and improving voltage profiles.

### Electric Vehicles (EV) Specifications and Modeling

**Electric vehicles (EV) charging and discharging rates.** The charging and discharging rates, integral to this study, were adopted from Khan et al. (2021) and are expressed in Equations 1 and 3. These mathematical models can be further elaborated expressed as Equations 2 and 4 to determine both the charging and discharging rates and the state-of-charge (SOC) trend per 15-minute interval, respectively.

$$P_{charging}(t) = [V_i(t) - 1] * Chr_{rate} \quad (1)$$

$$SOC_{new} = SOC_{old} + \frac{[V_i(t)-1]*Chr_{rate}}{(TEV)(B_{cap})} * \Delta t \quad (2)$$

$$P_{discharging}(t) = [1 - V_i(t)] * Dischr_{rate} \quad (3)$$

$$SOC_{new} = SOC_{old} - \frac{[1-V_i(t)]*Dischr_{rate}}{(TEV)(B_{cap})} * \Delta t \quad (4)$$

where:

$P_{charging}$	is the charging wattage (kW)
$P_{discharging}$	is the discharging wattage (kW)
$V_i$	is the specific voltage at bus $i$ (kV)
$Chr_{rate}$	is the charging rate (A)
$Dischr_{rate}$	is the discharging rate (A)
$B_{cap}$	is the EV battery capacity (kWh)
$SOC_{new}$	is the updated state-of-charge
$SOC_{old}$	is the previous state-of-charge
$\Delta t$	is the change in time (hr)
TEV	is the total number of EVs
$E_{useable}$	is the useable battery capacity (kWh)
$\Delta SOC_{charging}$	is the SOC difference for charging
$\Delta SOC_{discharging}$	is the SOC difference for discharging

**State-of-charge (SOC) based EV charging and discharging load modeling.** The equation used to identify SOC-based coordinated instantaneous charging and discharging powers, which was lifted from Akil et al. (2022), is employed in this study. This relationship was formulated from the real-time charging profiles of EVs, as shown in Equation 5.

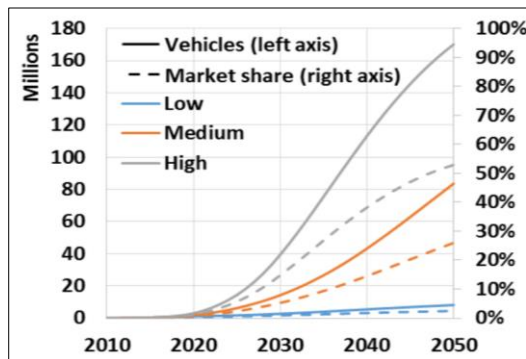
As per the piecewise model dictates, SOC-based instantaneous charging and discharging power reach their maximum when SOC is less than 73%. When SOC is at least 73% but less than 93%, there is a gradual decline in the maximum charging and discharging power. Finally, once SOC reaches 93% and above, the power remains constant.

$$P_i^n(SOC) = \begin{cases} P_i^c = P_{max} & \text{if } SOC < 73\% \\ P_i^c = P_{max} \cdot 3.17 \cdot -\ln(SOC) & \text{if } 73\% \leq SOC < 93\% \\ P_i^c = 0 & \text{if } SOC \geq 93\% \end{cases} \quad (5)$$

where:

$P_i^n(SOC)$	is the SOC-based instantaneous charging/discharging power (kW)
$P_i^c$	is the maximum instantaneous charging/discharging power (kW)
$P_{max}$	is the maximum charging power (kW)
$SOC$	is the state of charge

**Electric vehicle (EV) load growth and penetration model.** The penetration level is used in power engineering to anticipate the future possible impacts and real-case scenarios for power distribution networks. For instance, the Grid Integration Tech Team (GITT) and the Integrated Systems Analysis Tech Team (ISATT) of the United States have developed a real-life diagram that proposes low, medium, and high penetration scenarios from 2010 to 2050 (GITT and ISATT 2019) as shown in Figure 2.



**Figure 2.** Electric vehicles market penetration scenarios (GITT and ISATT 2019).

In examining the differences in the optimal placement between two possible types of EVCS installations, the unidirectional charging and bidirectional charging stations serve as the two main test cases in this study. These are examined through subtest cases according to EV penetration forecast and its corresponding optimal sites: 23% in 2030, 62% in 2040, and 94% in 2050 (GITT and ISATT 2019). The three subtest cases have corresponding optimal site parameters of 11, 30, and 25, respectively (Zambrano-Perilla 2016). The base, observed in 2015, serves as a reference point when no EV penetration rate recorded (GITT and ISATT 2019).

### Genetic Algorithm Implementation

Genetic algorithm (GA) is an iterative method inspired by the Darwinian theory of the survival of the fittest (Kathoch et al. 2021). It is a

common optimization method used in investigating the behavior of the power system. It starts with the random generation of  $n$  chromosomes in a population. After the initialization, the fitness function is computed for each chromosome in the population, followed by ranking for the whole generation. Next, offspring are produced by selecting parents from the existing population using the single-point crossover. The resulting offspring will undergo mutation to produce new offspring. This cycle continues until the desired number of generations is achieved (Liu 2013).

Using MATLAB R2022b, the implementation of the GA in the study involving EVCS integration to the distribution network is shown in Figure 3. Various penetration levels at different periods were introduced, with 100 population and 100 generations as GA parameters. Single-point crossover and a 0.05 mutation rate were utilized to optimize the location of EVCS.

The integration of EVs and EVCS into the distribution network results in an increase in average real and reactive power losses (Khalkali et al. 2015). The parameters were analyzed using OpenDSS as a power flow tool.

In the first subcase, UCT considers only the charging process, while the second subcase, BCT, considers both charging and discharging setup. Moreover, half of the BCT participated in two-way charging, and the other half only utilized one-way charging (Mehrabi et al. 2020). Simulation time varies based on the number of EVCS to be located and the consideration of charging/discharging setups. Overall, more than a day of simulation runtime was spent for six subcases. The simulations were conducted on a computer with an i5-8500 CPU Processor @ 3.00 GHz, 8.00 GB installed RAM and a 64-bit operating system.

### Objective Function

This paper primarily focused on the grid performance with the integration of EVCS into the network, in particular, to minimize the additional losses in the distribution network with the increase in EV penetration. Mathematically, the objective function is expressed in Equation 6:

$$F = \min \sum (S_{loss, EVCS} - S_{loss, base}) \quad (6)$$

where:

$F$	is the minimum additional losses (kVA)
$S_{loss, EVCS}$	is the total system apparent losses with EV integration (kVA)
$S_{loss, base}$	is the total system apparent losses without EV integration (kVA)

For the different cases, the voltage profile, system losses, and optimal sites were compared before and after the addition of EVCS. Moreover, the VPPI, PLRI, QLRI, and SLRI were utilized to differentiate unidirectional and bidirectional charging.

### Constraints

**Maximum EVCS of each bus.** It is highly recommended to have at least one charging station at each optimally selected bus to ensure the proper operation of the power distribution network. In this

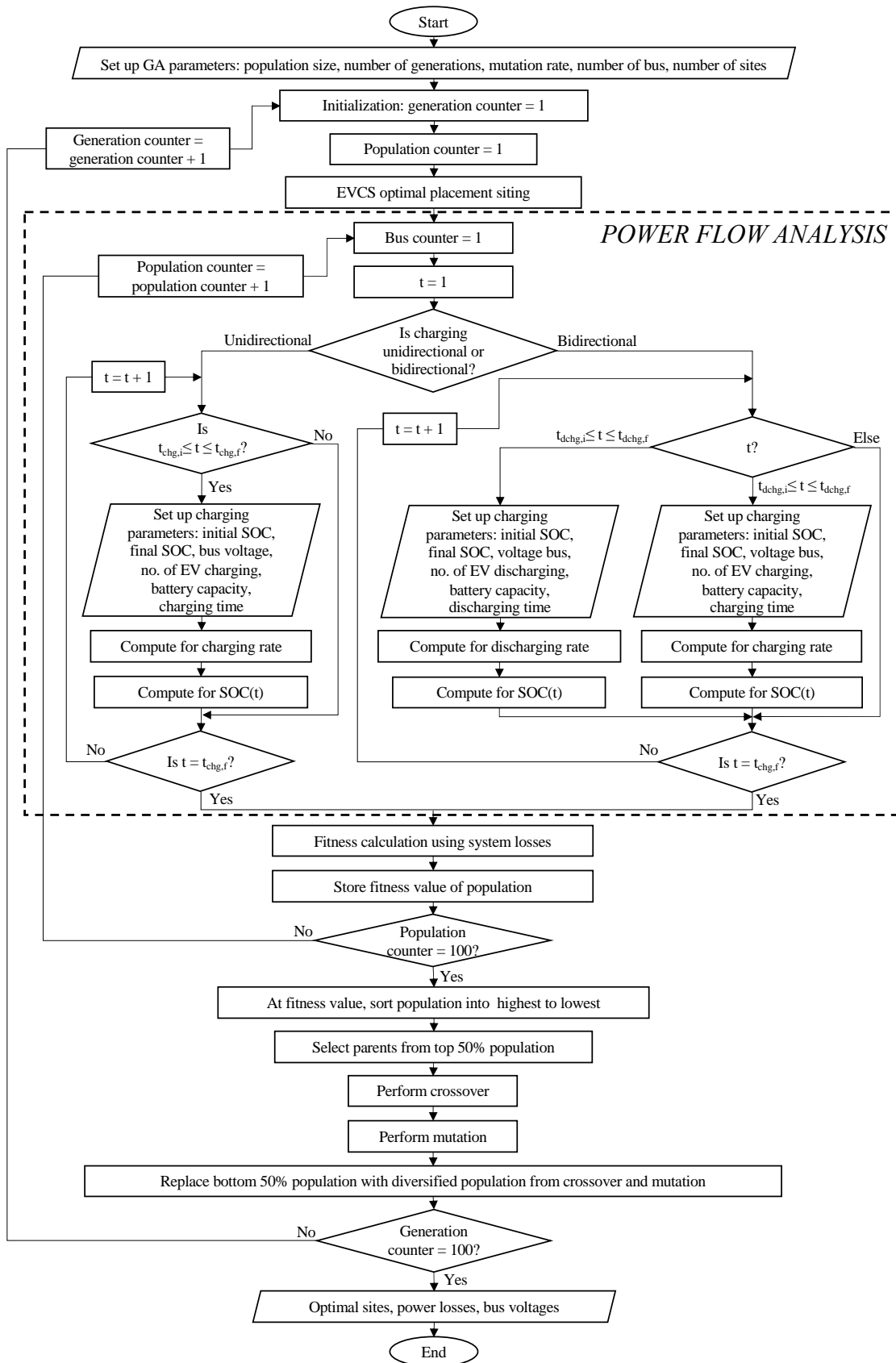


Figure 3. The process of integrating electric vehicle charging station (EVCS) unidirectional and bidirectional charging technologies into the distribution network.

scenario, the algorithm will accommodate a possible transportation route area without sacrificing the objective function (Fredriksson et al. 2019). However, for the purpose of this paper, particular buses suitable for station placements were limited to one to ensure that the EVCS infrastructure planning does not negatively affect the power grid.

**Maximum capacity of EVCS.** Each charging station within the specific area has limited charging ports. This limitation is attributed either to the connection's point inability to serve a large load or lack of budget. According to US DOE EERE (2023), the public EVCS at the Sacramento Parking Garage, located at #939, 10th Street, Sacramento, California, has a total electric vehicle supply equipment of 27 charging ports by which the same number of EVs can charge at the same time. Taking a real-life scenario, the assumed maximum charging capacity of a standard charging station is 27 EVs. However, due to the vast penetration of EVs in the last subcase, 50 charging ports per EVCS was assumed instead, which were used as the capacity in other studies (Kunj and Pal

2020). Equation 7 shows the allowable number of charging ports in each EVCS located at bus  $i$ :

$$CS(i) \leq CS_{max}; \sum_{i=1}^{N_{CS}} = N_{ch,max} \quad (7)$$

where:

- $CS(i)$  is the charging station capacity at bus  $i$
- $CS_{max}$  is the maximum charging station capacity
- $N_{CS}$  is the number of charging ports
- $N_{ch,max}$  is the maximum number of charging ports

**Permissible state of charging and discharging.** EV usage, mainly when it is to be charged, depends on the owner's decision. This study used the SOC range of 10% to 90% for charging, which is typical for an EV battery (Khalkali et al. 2015). Meanwhile, a range of 20% to 80% is used for discharging purposes due to the anticipation that EV owners will aim to maximize the sale of unused EV battery charge during peak electricity prices (Su et al. 2019). Thus, the limitation of the charging or discharging process as expressed in Equation 8 should be:

$$\begin{cases} SOC_n \leq SOC_{n,min}; \text{ for charging only} \\ SOC_{n,min} \leq SOC_n \leq SOC_{n,max}; \text{ for charging/discharging} \\ SOC_n \geq SOC_{n,max}; \text{ for discharging only} \end{cases} \quad (8)$$

where:

- $SOC_n$  is the chosen state-of-charge (%)
- $SOC_{n,max}$  is the maximum state-of-charge (%)
- $SOC_{n,min}$  is the minimum state-of-charge (%)

**Bus voltage tolerance.** Bus voltage can vary provided that it does not exceed the range of allowable voltage values and does not have an adverse effect on

the distribution network operation. To maintain the voltage range, it must be restricted between 0.9 and 1.1 per unit, as shown in Equation 9:

$$0.9 \text{ p.u.} \leq V_i \leq 1.1 \text{ p.u.} \quad (9)$$

where:

- $V_i$  is the voltage at bus  $i$  (p.u.)

## Evaluation Indices

### Voltage profile improvement index (VPPII).

One of the significant parameters that must be observed in the distribution network is its voltage deviation. The indicator assesses the nodal voltages concerning reference nodes and must be within acceptable limits. In evaluating the given cases, the

voltage profile and system losses of different penetration scenarios were compared to the base case. Then, the optimal sites were compared after adding EVCS for charging and charging-discharging scenarios. The mathematical equations are shown in Equations 10 to 11:

$$TVD_x = \frac{1}{n} \sum_{i=1}^n (V_i - V_{ref}^i) \quad (10)$$

$$VPPII = \frac{TVD_A + TVD_B + TVD_C}{3} \quad (11)$$

where:

- $TVD_x$  is the total voltage deviation of the system per unit at phase  $x$
- $V_i$  is the bus voltage at the bus  $i$  per unit
- $V_{ref}^i$  is the reference bus voltage
- $n$  is the total number of buses
- $VPPII$  is the voltage profile improvement index

**System power loss reduction indices.** The system losses and their respective reduction indices were used to compare the two charging technologies

and their corresponding optimal sites. The mathematical equations of the indices are given in Equations 12 to 14:

$$PLRI = \frac{P_{withUCT} - P_{withBCT}}{P_{withUCT}} \times 100 \quad (12)$$

where:

- $PLRI$  is the real power loss reduction index due to EV integration (%)
- $P_{withUCT}$  is the real power losses with EVCS UCT (kW)
- $P_{withBCT}$  is the real power losses with EVCS BCT (kW)

The second evaluation indicator in Equation 12 assesses whether the EVCS charging technology enhances or minimizes power losses. In particular, it

also indicates the loss reduction trend due to the bidirectional participation of the EVCS in the grid relative to the unidirectional charging process.

$$QLRI = \frac{Q_{withUCT} - Q_{withBCT}}{Q_{withUCT}} \times 100 \quad (13)$$

where:

- $QLRI$  is the reactive power loss reduction index due to EV integration (%)
- $Q_{withUCT}$  is the reactive power losses with EVCS UCT (kVAR)
- $Q_{withBCT}$  is the reactive power losses with EVCS BCT (kVAR)

One of the indicators of a stable grid is when the reactive power losses do not exceed a threshold value (Chen et al. 2021). The QLRI in Equation 13 aims to determine if reactive power loss is within the definite threshold value to maintain the power grid stability. Thus, the QLRI has been included in the study.

Shown in Equation 14 is the SLRI, which is one of the important parameters to assess whether the distribution network has improved. When it goes beyond the threshold value, the distribution network is unstable.

$$SLRI = \frac{S_{withUCT} - S_{withBCT}}{S_{withUCT}} \times 100 \quad (14)$$

where:

- $SLRI$  is the apparent power loss reduction index due to EV integration (%)
- $S_{withUCT}$  is the apparent power losses with EVCS UCT (kVA)
- $S_{withBCT}$  is the apparent power losses with EVCS BCT (kVA)

## RESULTS

### Total Integration of EVs and EVCS Considering Increasing Penetration Levels

The graph in Figure 4 illustrates the power losses at each penetration stage for their respective technologies. The anticipated apparent power loss of EV integration in the year 2050 is extremely high at 465.2930 kVA for both charging technologies, as its penetration level is expected to reach 94%. Meanwhile, the year 2040 in the graph has a relatively lower penetration than the year 2050 at 62%, which gives apparent power losses of 364.5683 kVA for both unidirectional and bidirectional modes. In the year 2030 EV integration, forecasted losses are at 96.3131 kVA for unidirectional charging and 92.0304 kVA for

bidirectional charging, given a 23% penetration rate. All these values are relatively higher than the base power loss of 82.5268 kVA in the year 2015 when there was no EV integration.

In Figures 5 and 6, the plots of UCT and BCT per-unit bus voltage for phase A of the system depict a gradual decrease in bus voltages with increasing EV penetration levels, regardless of the charging technology. On the other hand, in terms of VPII, as shown in Table 1, there is an increase in the indices concerning the considered years. Moreover, in the years 2040 and 2050, similar results were observed for both unidirectional and bidirectional charging. Meanwhile, the UCT in the year 2030 has a higher VPII than the BCT.

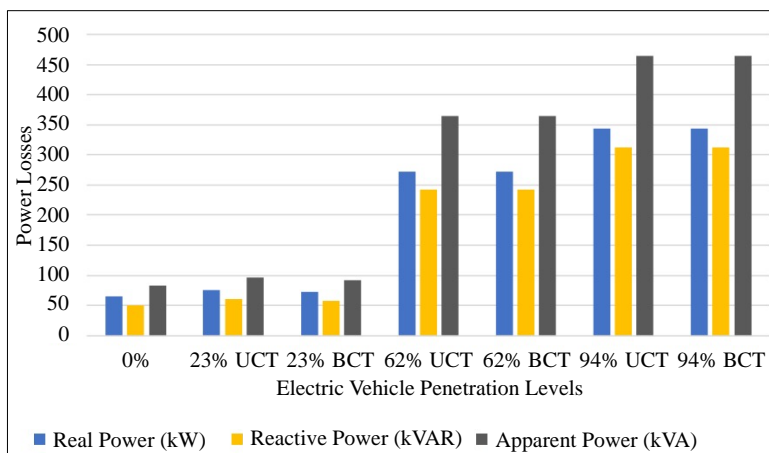


Figure 4. The graph of power losses vs. EV penetration levels. UCT- unidirectional charging technology; BCT- bidirectional charging technology.

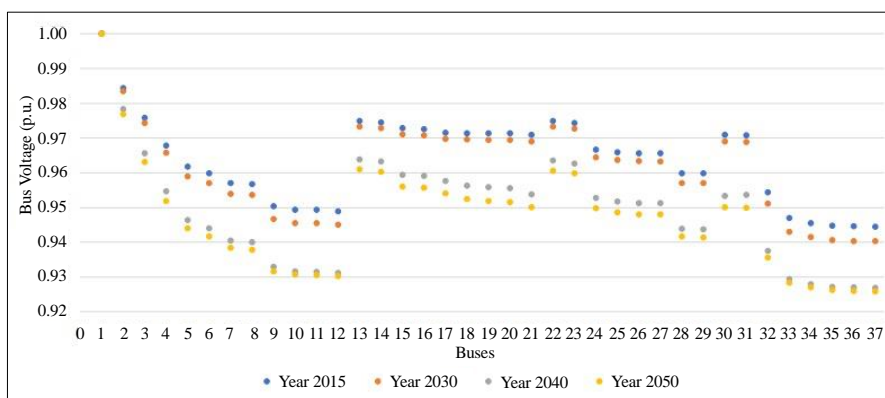


Figure 5. The per-unit bus voltage profile in different test case scenarios in unidirectional charging technology setup for phase A.

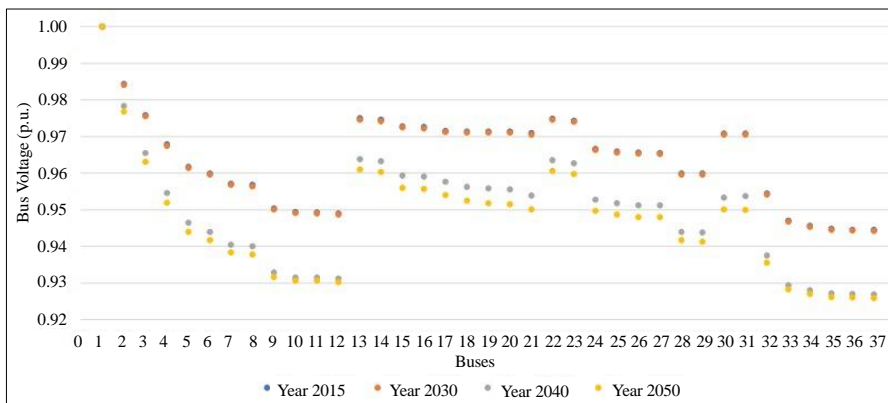


Figure 6. The per-unit bus voltage profile in different test case scenarios in bidirectional charging technology setup for phase A.

Table 1. The voltage profile improvement index (VPII) between unidirectional charging technology (UCT) and bidirectional charging technology (BCT) for different penetration levels.

Year	Subcase (penetration level)	VPII (p.u.)	
		UCT	BCT
2015	0%	1.1139	
2030	23%	1.2020	1.1249
2040	62%	1.6260	1.6259
2050	94%	1.7094	1.7094

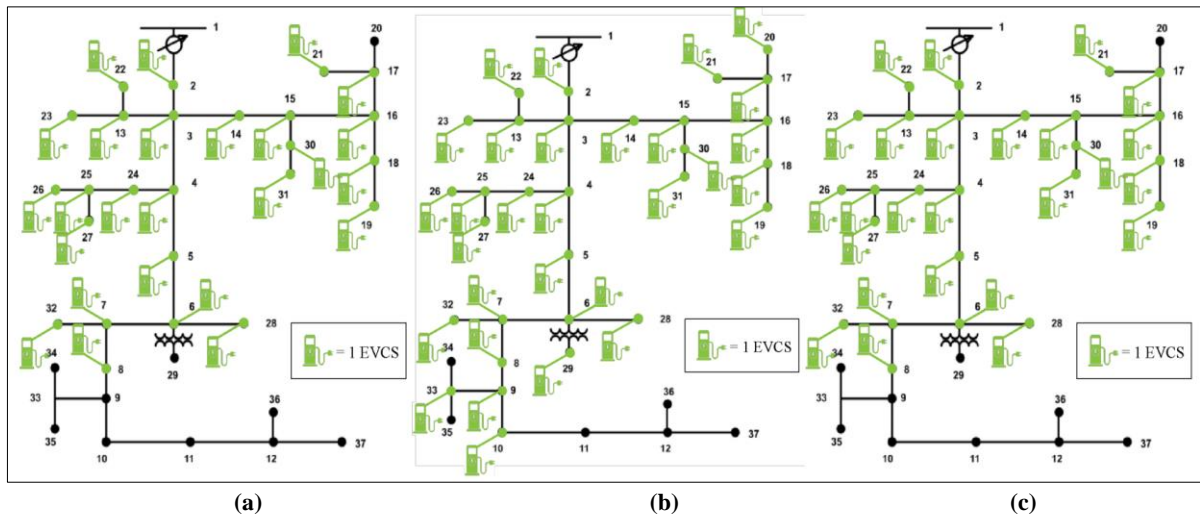
**Electric Vehicle Charging Station (EVCS) Optimal siting**

Optimal placement assesses the possible site combinations to attain the objective function. To facilitate a comparison of optimal locations for each subcase based on the penetration level, Table 2

presents a summary of the optimization results. The same buses were identified in the subcases of both power charging technology for every penetration level. Lastly, Figure 7 illustrates that the resulting optimal sites are concentrated in specific locations.

**Table 2.** Optimal electric vehicle charging station sites at different penetration levels.

Main Cases	Subcase (penetration level)	Year	Optimal Sites
Unidirectional	23%	2030	2, 3, 4, 5, 13, 14, 15, 22, 23, 30
Bidirectional			2, 3, 4, 5, 13, 14, 15, 22, 23, 30
Unidirectional	62%	2040	2, 3, 4, 5, 6, 7, 8, 9, 10, 13, 14, 15, 16, 17, 18, 19, 20, 21, 22, 23, 24, 25, 26, 27, 28, 29, 30, 31, 32, 33
Bidirectional			2, 3, 4, 5, 6, 7, 8, 9, 10, 13, 14, 15, 16, 17, 18, 19, 20, 21, 22, 23, 24, 25, 26, 27, 28, 29, 30, 31, 32, 33
Unidirectional	94%	2050	2, 3, 4, 5, 6, 7, 8, 13, 14, 15, 16, 17, 18, 19, 21, 22, 23, 24, 25, 26, 27, 28, 30, 31, 32
Bidirectional			2, 3, 4, 5, 6, 7, 8, 13, 14, 15, 16, 17, 18, 19, 21, 22, 23, 24, 25, 26, 27, 28, 30, 31, 32



**Figure 7.** The electric vehicle charging stations (EVCS) installation at (a) 23%, (b) 62%, and (c) 94% EV penetration rates.

**Comparison of Index Values Between Charging Technologies**

The VPII of all buses connected to phase A in the test system was calculated with the year 2015 as the reference. Figure 8 reveals that there is almost no difference in VPII for the bidirectional case of the year 2030 compared to the base case. On the other hand, a

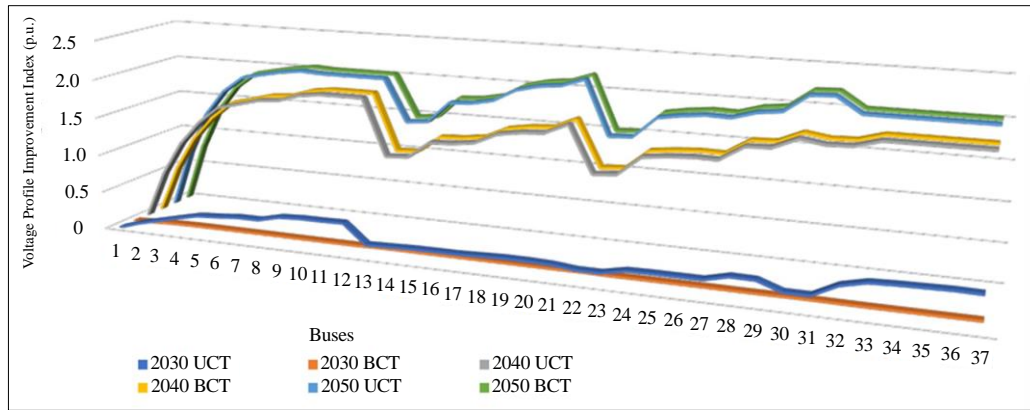
slight distortion is apparent in the graph of unidirectional charging.

Using the results of unidirectional charging as reference values for the bidirectional case, system power losses were analyzed to compare the characteristics of the charging technologies. Table 3 shows that only the year 2030 has non-zero index values.

**Table 3.** Real, reactive, and apparent power loss reduction indices. PLRI - real power loss reduction index; QLRI- reactive power loss reduction index; SLRI-apparent power loss reduction index.

Subcase (penetration level)	Year	Evaluation Indicators of Comparison		
		PLRI	QLRI	SLRI
23%	2030	4.142878856	4.916659909	4.44664329
62%	2040	0	0	0
94%	2050	0	0	0





**Figure 8.** Phase A voltage profile improvement index (VPII) of the entire network. UCT- unidirectional charging technology; BCT- bidirectional charging technology

## DISCUSSION

### Total Integration of EVs and EVCS Considering Increasing Penetration Levels

As observed in the graph presented in Figure 4, there exists a direct relationship between the penetration of EVs and the apparent power loss. Higher integration of EVs results in a higher apparent power loss, regardless of the charging technology employed in the system. The anticipated transition from ICE vehicles to EVs is expected to contribute to higher apparent losses in the power grid. Variations in the penetration level have caused changes in the power loss, the number of optimal sites, and the capacity of the EVCS. It can be concluded that there is a significant increase in real, reactive, and apparent power losses. Thus, the higher anticipation of EV loads due to the increasing penetration level through time is expected to be caused by the advancements in the transportation sector. Without proper intervention, additional EV loads could alter the normal grid operations.

On the other hand, the varying penetration levels in this study assessed the impacts of the EV integration. The dramatic increase in penetration rate has a negative effect on the voltage profile of the power grid. As the injected EV load increases, the bus voltages deviate from the ideal value. With the expected technological advancement in EVs, there is an anticipated voltage profile deterioration by as much as 1.7094 p.u. Based on the trend analysis of Table 1 values, the voltage profile is deteriorating in the subsequent years.

### Electric Vehicle Charging Station (EVCS) optimal siting

Overall, there is a commonality among the respective penetration levels where the resulting optimal sites in the test system cluster in the nodes near the supply bus. Indeed, these results are consistent

with recent works. One of these studies has presented that connecting an EVCS at any bus in the distribution network increases the active power loss due to the resistance of the branches from the slack bus to the considered node (Bilal et al. 2021). Thus, to reduce the resulting power loss, the EVCS should be located closer to the upstream network near the supply (Hadian et al. 2020).

The observation shows similar results when the number of optimal sites increases in the distribution network. Another factor contributing to this trend is the bidirectional participation of EV owners. According to Mehrabi et al. (2020), the willingness of the drivers to participate in the bidirectional charging is 50%. This means the ratio of charging and discharging in the bidirectional technology is 2:1, indicating that for every two charging vehicles, only one can discharge. With this, the research suggests that all EVs still need to be charged before this half undergoes discharging energy into the power system. Hence, UCT and BCT had the same optimal sites since they usually acted as a load.

### Comparison of Index Values Between Charging Technologies

After identifying the optimal sites where minimum power losses are observed, there is a noticeable difference between the two power charging directions. During the years with low penetration, there is less power loss in UCT than in BCT. However, during higher EV penetration beyond the year 2040, the distribution network experiences more apparent power losses, considering that BCT has become the norm in society. Correspondingly, power losses increase as the EV penetration rate increases from the base case year 2015 when there is no EV integration.

The BCT in the year 2030 exhibits a more improved power loss relative to the UCT of the same year since it deviates mainly from the base case. On the other hand, the EV load integrated into the system

contributes to the fluctuations in the UCT in 2030, as observed in Figure 8. Meanwhile, power losses in both charging technologies are the same for the years 2040 and 2050, where EV penetration is higher than in the former cases. The discharging scenario in BCT positively affects the distribution network by providing additional power to the system. Nevertheless, due to the high EV loads, both the UCT and BCT negatively affect the grid, which resulted in high VPPII values.

Table 3 presents the system loss indices for the BCT subcases relative to their corresponding UCT counterparts. For 2030, the power loss index values of 4.14%, 4.92%, and 4.45% are determined for PLRI, QLRI, and SLRI, respectively. Since there is a positive value in PLRI at 4.14%, there is a minimization and loss reduction trend this year brought by utilizing BCT over the UCT case. Moreover, the 4.92% QLRI implies a more stable power grid from BCT application in the said year. Lastly, the 4.45% SLRI shows an overall improvement in the system due to the reduction in apparent power losses from using bidirectional charging.

Meanwhile, the three evaluation indices have zero values at higher penetration rates of 62% and 94%. These results signify that the system losses from using BCT and UCT are the same, and there is no advantage in using the former over the latter.

The results suggest that bidirectional charging can provide opportunities to improve the distribution system if more EV owners utilize it. A 50% EV bidirectional participation is insufficient to make the distribution system better since the result shows no difference compared to unidirectional charging at high EV penetration. Hence, further studies can examine the possibility of incentivizing participation in bidirectional charging through tariffs and tax credits. The distribution system operator can do proper grid interventions, such as replacing the lines with lower resistance and utilizing power loss compensator equipment in anticipation of higher EV loads in the future. Moreover, another path of future research on differentiating the two charging technologies is studying the cost-effectiveness, battery efficiency and deterioration, and grid communication, particularly with traffic demand management of the mixed flow of two charging technologies.

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## ETHICAL CONSIDERATIONS

This study is guided by ethical and legal principles, responsibly ensuring that this is free of research misconduct and presents the results

accurately. Moreover, the authors of the references were duly acknowledged and cited with proper citations.

## DECLARATION OF COMPETING INTEREST

The authors declare that there are no competing interests among any authors.

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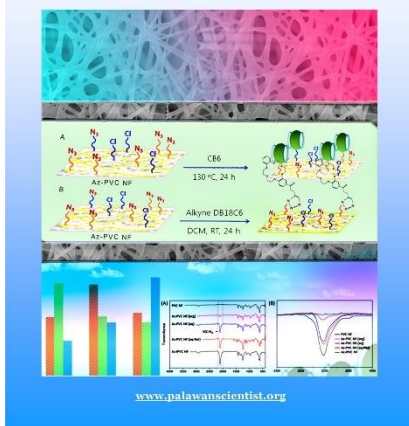
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# Numerical modeling of the drying behavior of Adlai (*Coix lacryma-jobi* L.) grain

Alvin M. Ante<sup>1\*</sup>, Arnold R. Elepaño<sup>2</sup>, Kevin F. Yaptenco<sup>2</sup>  
 and Delfin C. Suministrado<sup>3</sup>

<sup>1</sup>Department of Agricultural and Biosystems Engineering, College of Engineering, Dr. Emilio B. Espinosa Sr. Memorial State College of Agriculture and Technology, Mandaon, Masbate, Philippines

<sup>2</sup>Agricultural, Food and Bio-process Engineering Division, Institute of Agricultural and Biosystems Engineering, College of Engineering and Agro-Industrial Technology, University of the Philippines Los Baños, Laguna, Philippines

<sup>3</sup>Agribiosystems Machinery and Power Engineering Division, Institute of Agricultural and Biosystems Engineering, College of Engineering and Agro-Industrial Technology, University of the Philippines Los Baños, Laguna, Philippines

\*Correspondence: [amante@up.edu.ph](mailto:amante@up.edu.ph) or [antealvin@gmail.com](mailto:antealvin@gmail.com)

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## ABSTRACT

The study aimed to develop computational fluid dynamics (CFD) models for simulating the drying performance of Adlai grain during convective drying with an air temperature range of 30°C to 60°C at around 10% to 80% relative humidity (RH). Before CFD modeling, the calculation of selected thermophysical properties of Adlai through mathematical modeling of its food constituents and thin-layer drying experiments were conducted. The simulation of heat and mass (moisture) transfer and visualization of moisture and temperature gradients in Adlai grain during drying were carried out using Analysis Systems (Ansys) Student 2020 R2 software package, specifically the Fluent solver. Results showed that the CFD models exhibited good agreement with the actual drying performance of Adlai. The models were validated using three statistical parameters: coefficient of determination ( $R^2$ ), standard error (S), and percent mean deviation modulus (P%). The  $R^2$  values ranged from 0.94-0.98; the S values ranged from 0.0018-0.0066; and the P% values ranged from 6.5%-8.68%. Overall, the models were deemed acceptable in estimating the moisture content of Adlai due to high  $R^2$  values, low S values, and P% values of less than 10%. The results validate the use of CFD as a reliable method for predicting the drying performance of Adlai, which contributes to the optimization of the drying process, the improved designing of drying systems, and the enhancement of product quality.

**Keywords:** grain drying, hot air, multiphysics, simulation

## INTRODUCTION

The Bureau of Agricultural Research (BAR) has initiated research and development activities on Adlai (*Coix lacryma-jobi*) since 2011. It saw the highly nutritious Adlai grain or “Job’s tears” as a potential alternative to rice as a staple food. Approximately, the Adlai grain contains the following: 9.9-10.8% water; 13.6-19.1% protein; 5.7-6.1% fat; 58.5-62.7% carbohydrate; 8.4% fiber and 2.2-2.6% ash (Aradilla 2018). Compared to rice and corn, Adlai

has higher food energy, carbohydrates, protein, fat, and dietary fiber. It also has the following minerals: calcium, phosphorus, iron, niacin, thiamine, and riboflavin (Peñaflor et al. 2014). Three of its varieties are endemic in the Philippines – “guilian”, “ginampay”, and “pulot”. Research and other related activities have also been done by other agencies such as the Department of Agriculture (DA) and the Philippine Center for Post-Harvest Mechanization and



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Development (PhilMech) to study and enhance Adlai's production, post-harvest operations, and commercialization.

Currently, Adlai is grown and consumed in areas such as Zamboanga del Sur, Isabela, Batangas, Caraga Region, and the Bicol Region. The production of Adlai follows the same traditional method for rice production. The "ginampay" variety of Adlai could produce a yield of 3.4 tons/ha, which is just slightly lower than the average rice yield of 3.87 tons/ha. The traditional processing of Adlai after harvesting involves drying and milling. Sun-drying of threshed grains occurs two days before milling (Aradilla 2018). Unhulled Adlai grains sell for around PHP 200.00 per kg, while milled grains go for around PHP 300.00 to PHP 500.00 per kg.

There are studies on the processing of Adlai into several products, such as ice cream (Khongjeamsiri 2007), yogurt (Keeratibunharn and Krasaekoopt 2013), and saltine crackers (Andoy et al. 2019). However, some of its post-harvest qualities, such as its drying characteristics and equilibrium moisture content (EMC), remain unknown. Drying is one of the most fundamental and indispensable unit operations in grain processing. It is performed to reduce the moisture content level in food and other biological materials with the aim to prolong its shelf life, reducing the probability of fungi development, and facilitating further processing to obtain a finished product available for consumption, utilization, and other purposes. For cereal grains like Adlai, proper drying must be performed to improve grain quality and storage time. Convective and radiation drying methods are the two general types of drying methods used for grains. These methods are embodied in the commonly used and preferred drying operations in the Philippines: sun-drying and convective drying. The latter is a complex process that involves simultaneous heat and mass (moisture) transfer, change in physical properties, and shrinkage of biological material (Kumar et al. 2012). Convection drying is performed by supplying heated air to a batch of grains via a blower or fan. Moisture removal results from the vaporization of moisture before being taken away by the air stream.

Convection drying of agricultural products, such as grains, is both energy and time-consuming. The food processing industry needs to obtain high-quality dry products utilizing only the lowest possible amount of energy input, including labor, and the shortest drying time. With the powerful computing capabilities of programs like Ansys Fluent, Ansys CFX, COMSOL Multiphysics, etc. numerical modeling of the drying phenomena in grains became achievable. This is done by computational fluid dynamics (CFD) simulation employed by such programs to aid engineers and scientists in solving multiphysics problems involving mass, momentum, and energy transfer. In studying the drying kinetics of

agricultural and biological products, CFD has become a common practice among food engineers and technologists due to its energy, time, and cost-saving benefits in contrast to laborious experimentation. However, actual experiments are still helpful in validating the models. For grains like Adlai, a well-developed CFD model can provide a more accurate and efficient description of its drying behavior.

The study aimed to develop CFD models for simulating the drying performance of Adlai grain by numerical modeling of the heat and mass (moisture) transfer in the kernel during convective drying. Specifically, the study aimed to calculate the selected thermophysical properties of Adlai grain needed for CFD simulation; to perform a thin-layer drying experiment of Adlai grain, to perform CFD modeling using multiphysics software, and to validate the CFD-generated drying curves of Adlai grain with the experimental results. Through numerical simulations, predicting the drying behavior of Adlai under various drying conditions would be possible and would contribute to optimizing further drying systems, equipment, and practices for Adlai.

## METHODS

### Thermophysical Properties of Adlai

Values for selected thermophysical properties of Adlai were calculated using mathematical models developed by Choi and Okos (1986) that used the proximate composition or fraction of food constituents of the product (Ibarz and Barbosa-Canovas 2014). The models or equations used in the study were expressed as a function of temperature in the range of -40°C to 150°C. These equations are presented in Table 1.

### Properties of the Drying Air

The properties of the drying air used in the convection drying of Adlai are necessary in developing the CFD models. They were used in the material properties panel of the fluid cell zone of the models. Table 2 presents the values for the different air properties used in the CFD simulations.

### Thin-Layer Drying Experiment

The initial moisture content of sample Adlai grains in MC d.b. (moisture content dry basis) was calculated using the conventional air oven method, following the recommended procedures from the American Society of Agricultural Engineers (ASAE) Standards for moisture measurement of unground grain and seeds. There is no recommended oven temperature, sample size, and heating period for Adlai yet. However, for the purpose of the study, the maximum allowable oven temperature of 130°C was used along with the standard sample size of 10 g and

heating period of 18 h. This method was also adopted by Kim et al. (2016) in thin-layer drying of sorghum. The study used the “guilian” variety of Adlai due to its availability and cultivation for human consumption.

The grain samples were stored at 10°C in cold storage for seven days and were subjected to room temperature 24 h prior to the commencement of

the thin-layer drying experiment. A tunnel dryer consisting of a blower, ducts, a heating element, and a temperature controller was used to supply heated drying air to the single layer of Adlai grains resting on a perforated stainless-steel drying tray. The mass of the grains used was around 35 g per replication. The dryer’s schematic design is shown in Figure 1.

**Table 1.** Mathematical models used to estimate the thermophysical properties of Adlai (Ibarz and Barbosa-Canovas 2014; Carson et al. 2016).

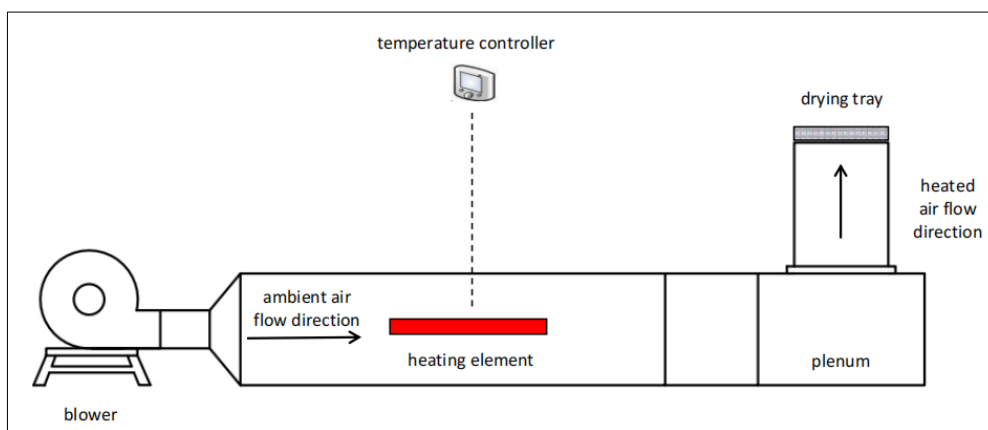
Property	Food Component	Property Model ( <i>t</i> = temperature in °C)	Equation/s
Thermal conductivity (k), W/m-K	Protein	$k = 1.7881E-01 + 1.1958E-03t - 2.7178E-06t^2$	$k_{III} = \frac{\Sigma(k_i v_i)}{\Sigma v_i}$ $k_{III} = \frac{(3v_a - 1)k_a + [3(1 - v_a) - 1]k_l + \sqrt{[(3v_a - 1)k_a + [3(1 - v_a) - 1]k_l]^2 + 8k_l k_a}}{4}$ where: k <sub>l</sub> = thermal conductivity from Parallel Model, W/m-K k <sub>III</sub> = thermal conductivity from Effective Medium Theory model, W/m-K v <sub>a</sub> = volume fraction of air v <sub>i</sub> = volume fraction of food constituent k <sub>a</sub> = thermal conductivity of air, W/m-K
	Fat	$k = 1.8071E-01 - 2.7604E-04t - 1.7749E-07t^2$	
	Carbohydrate	$k = 2.0141E-01 + 1.3784E-03t - 4.3312E-06t^2$	
	Fiber	$k = 1.8331E-01 + 1.2497E-03t - 3.1683E-06t^2$	
	Ash	$k = 3.2962E-01 + 1.4011E-03t - 2.9069E-06t^2$	
Density (ρ), kg/m <sup>3</sup>	Protein	$\rho = 1.3299E-03 - 5.1840E-01t$	$\rho = \frac{(1 - \epsilon)}{\Sigma(x_i / \rho_i)}$ where: ρ = true density, kg/m <sup>3</sup> ε = porosity x <sub>i</sub> = mass fraction of food constituent ρ <sub>i</sub> = density of food constituent, kg/m <sup>3</sup>
	Fat	$\rho = 9.2559E02 - 4.1757E-01t$	
	Carbohydrate	$\rho = 1.5991E03 - 3.1046E-01t$	
	Fiber	$\rho = 1.3115E03 - 3.6589E-01t$	
	Ash	$\rho = 2.42385E03 - 2.8063E-01t$	
Specific heat (C <sub>p</sub> ), kJ/kg-K	Protein	$c_p = 2.0082 + 1.2089 E-03t - 1.3129 E-06t^2$	$c_u = \Sigma(c_i x_i)$ where: c <sub>u</sub> = specific heat, kJ/kg-K c <sub>i</sub> = specific heat of food constituent, kJ/kg-K x <sub>i</sub> = mass fraction of food constituent
	Fat	$c_p = 1.9842 + 1.4733E-03t - 4.8008E-06t^2$	
	Carbohydrate	$c_p = 1.5488 + 1.9625E-03t - 5.9399 E-06t^2$	
	Fiber	$c_p = 1.8459 + 1.8306E-03t - 4.6509E-06t^2$	
	Ash	$c_p = 1.0926 + 1.8896 E-03t - 3.6817 E-06t^2$	
Water	Thermal conductivity	$k_w = 5.7109 E-01 + 1.7625E-03t - 6.7036E-06t^2$	
	Density	$\rho_w = 9.9718E02 + 3.1439E-03t - 3.7574E-03t^2$	
	Specific heat (for temperature range of 0 to 150 °C)	$c_w = 4.1289 - 9.0864 E-05t + 5.4731E-06t^2$	

The Adlai grain samples were subjected to forced heated air drying at three drying air temperatures: 30°C (ambient), 45°C, and 60°C. These temperature values were selected for the study since a drying air temperature of 30°C represents the average ambient air temperature; 45°C drying air was used relative to maintaining seed viability; and 60°C drying

air was used for commercial drying. A constant air velocity of around 1 m/s was used during the experiment. The drying air temperature was maintained during drying via a temperature controller (Elitech STC-1000 Temperature Controller), while the velocity of the drying air was monitored via an anemometer (PEAKMETER® Digital Anemometer PM6252A).

**Table 2.** Air properties used at various drying air temperatures.

Property	Drying Air Temperature (°C)			References
	30	45	60	
Density, kg/m <sup>3</sup>	1.164	1.109	1.059	Cengel and Cimbala 2014
Specific heat, J/kg-K	1007	1007	1007	
Thermal conductivity, W/m-K	0.02588	0.02699	0.02808	
Dynamic viscosity, kg/m-s	1.87 x 10 <sup>-05</sup>	1.94 x 10 <sup>-05</sup>	2.01 x 10 <sup>-05</sup>	
Diffusion coefficient of water into air, kg/m-s	3.118 x 10 <sup>-05</sup>	3.299 x 10 <sup>-05</sup>	3.469 x 10 <sup>-05</sup>	The Engineering ToolBox 2018



**Figure 1.** Schematic of the dryer used for thin-layer drying of Adlai.

The relative humidity (RH) during drying was monitored using a digital thermohygrometer (Senze Instruments Digital Hygro Thermometer TH-D-TH06OH). The change in mass of the grains was recorded manually at intervals recommended by ASAE standards: every 5 seconds during the first 5 min, every 1 min during the next hour, and every 15 min after that. An electronic balance was used for calculating the change in mass of the grain samples (Shenzhen Big Dipper Scale Co., Ltd., BDS-PN PN602A). Subsequently, the drying curve was generated from the experiment. Drying was performed until equilibrium was reached between the Adlai grains and the drying air. The experiment was conducted in triplicate per treatment of drying air temperature.

Fick’s second law of diffusion was also used to describe the drying phenomenon since thin-layer drying occurred in the falling rate period, and diffusion of moisture inside the kernel is the resistance factor governing the process. The simplified diffusion model is based on the solution of Henderson’s moisture

diffusion law (Kim et al. 2016). The model is expressed as:

$$MR = A \exp(-kt) \tag{1}$$

where  $A$  and  $k$  are model constants. The drying constant,  $k$ , is expressed as per unit time. The time duration of drying is represented by  $t$ , usually in hours or minutes.

The average effective diffusivity ( $D_e$ ) values of Adlai grain moisture at different drying air temperatures were calculated from the experimental drying curves and were used in the CFD simulations. It was assumed to be constant during the drying process. The effect of drying air temperature on effective diffusivity was represented through an exponential model (Peleg et al. 2012) expressed as

$$D_e = D_o \exp(cT_a) \tag{2}$$

where  $D_e$  is the effective diffusivity (m<sup>2</sup>/s),  $D_o$  is a diffusivity at infinite high temperature,  $c$  is a constant (1/K), and  $T_a$  is the drying air temperature (K).



### Computational Fluid Dynamics (CFD) Drying Simulation of Single Adlai Kernel

Thin-layer drying can be considered as grain being immersed in fluid (Prakash and Pan 2011); hence, its mechanism can be described by single kernel drying and its governing equations. This was reflected in the study of Chen et al. (2023) on the heat and mass transfer modeling of a single-particle peanut, wherein the numerical model was validated by a thin-layer drying experiment using hot air. The following steps were adopted to create the CFD drying models for a single Adlai kernel: (1) generating the model geometry, (2) mesh/grid generation, (3) setting up the Fluent solver, (4) numerical solution with Fluent, and (5) post-processing. The multiphysics software used for the study was Ansys Student 2020 R2, which is a free academic version of the Ansys software package. Computational fluid dynamics simulations were explicitly conducted using the Fluent solver of Ansys Student.

The 3D teardrop geometry representing the Adlai or “Job’s tears” kernel was made using AutoCAD 2016 software by Autodesk®. The average major and minor axes of the kernel were calculated from the measurement of 20 Adlai grain samples using a Vernier caliper (1/128 in). The 3D Adlai model was further enhanced when imported into DesignModeler, an object-generating component of Fluent, as shown in Figure 2. The flow field for the

drying air was added as an enclosure surrounding the kernel geometry. The mesh of the CFD simulations resulted in 212,436 elements and 43,710 nodes. Since the study used the Ansys Student version only, the mesh was refined to a level where further refinement did not enhance the results of the numerical solution. Unstructured tetrahedral elements were used as recommended for 3D meshing. Inflation was used on the walls of the kernel for improved gradient calculation. Quadrilateral cells were used for inflation along the surface of the kernel. The mesh of the kernel geometry is presented in Figure 3.

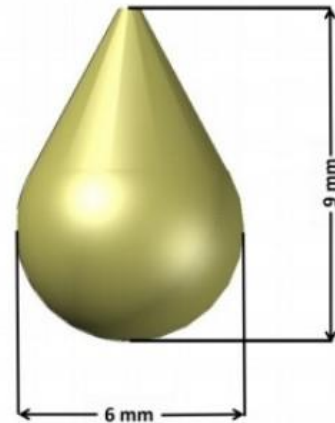


Figure 2. The 3D geometry of single Adlai grain.

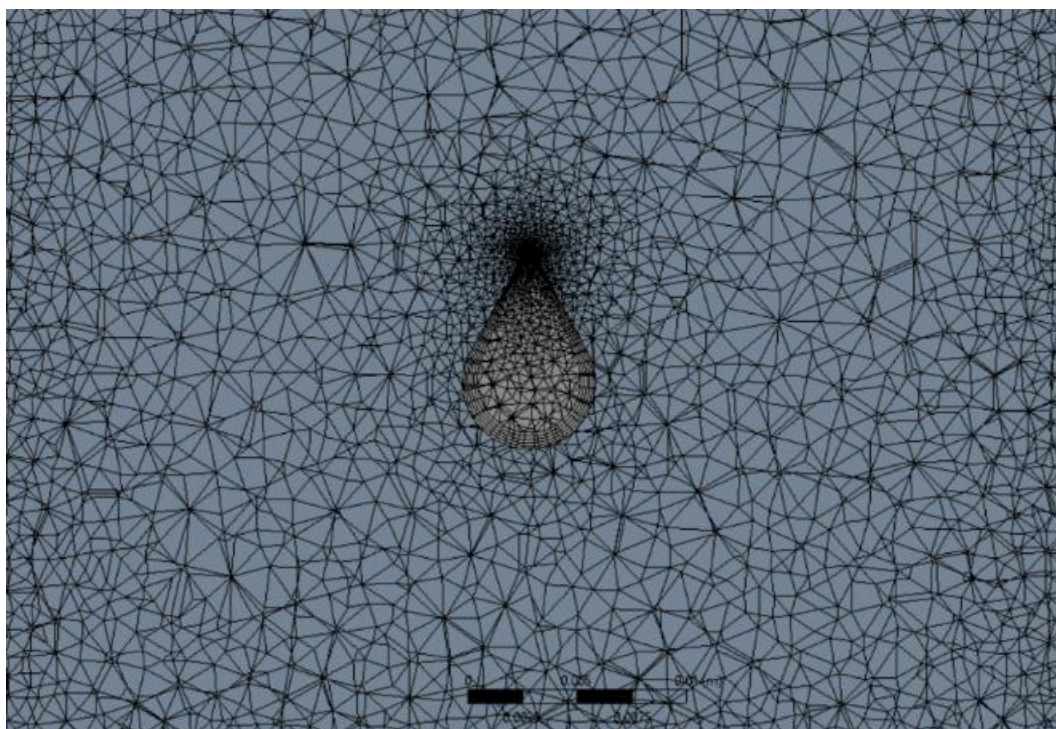


Figure 3. Unstructured mesh (cross-section) of the computational fluid dynamics (CFD) model for convective drying of Adlai grain.

Before proceeding with the development of the CFD simulations using Fluent, specific assumptions about the model were established based on the earlier works of other researchers in the field of CFD and its applications in drying biological products. The following were the governing assumptions adopted for the CFD models in the study: (1) no variation in shape and size of the kernel during drying, (2) the kernel is considered homogeneous, (3) uniform moisture distribution, (4) constant effective diffusivity and convective mass transfer coefficient during diffusion, (5) diffusion is the only transport mechanism of moisture inside the kernel, (6) moisture transfer to the drying air only occurs at the kernel surface, and (7) no volumetric heat generation within the kernel (ElGamal et al. 2014; Chilka and Ranade 2018). The geometry used in the simulations was considered a homogeneous lumped model of the Adlai grain. The thermophysical properties of the grain were applied to the lumped model's entirety because there is no available information from literature on the specific values for the endosperm, bran, and hull of Adlai. Furthermore, grains, in general, are practically dried as a whole with the kernel, bran, and husk layers intact, so a single-component model for Adlai grain is acceptable, at least for the study.

A non-conjugate diffusion-based model was used in this study wherein the heat and mass transfer inside the product are computed without affecting the flow, temperature, and water vapor concentration in the flow field of the drying air. To model the moisture content inside the product, an effective diffusivity is used, which is obtained only through experimentation. The boundary conditions at the interface require values of convective heat transfer coefficient and convective mass transfer coefficient to account for the heat and mass exchange between the product and the drying air (Sun 2019).

In the numerical solution using the Fluent solver, the steady state solution was done before conducting the transient simulation, where the drying air temperature near the kernel surface is set at the initial kernel temperature, and the moisture content of the air is in equilibrium with the kernel's (Sun 2007). Constant air velocity (1 m/s) and temperature (30°C, 45°C, and 60°C) were used at the air inlet. The drying air originated from the bottom of the domain. Zero temperature and moisture gradients were placed on the walls of the flow field. The moisture content of the drying air was not considered in the simulations due to the non-conjugate approach adopted for the study. The SST-k $\omega$  model was used for turbulence modeling. An initial value for the temperature and moisture of the kernel was set. The initial moisture content ( $M_0$ ) was 0.1867 MC d.b. while the kernel temperature ( $T_0$ ) was set to 27°C. The initial conditions for the first stage of the solution were:

$$M(x,y,z,0) = M_0 \quad (3)$$

$$T(x,y,z,0) = T_0 \quad (4)$$

User-Defined Scalar (UDS) equation was used to simulate the moisture content within the kernel. Ansys Fluent solver uses the following UDS equation:

$$\frac{\partial \rho \phi_k}{\partial t} + \frac{\partial}{\partial x_i} \left( \rho u_i \phi_k - \Gamma_k \frac{\partial \phi_k}{\partial x_i} \right) = S_{\phi_k} \quad k=1, \dots, N \quad (5)$$

From left to right, an unsteady term, convective term, diffusion term, and source term comprise the UDS equation as presented in Equation 5. Since it was assumed that the primary driving force for the drying of the kernel is diffusion, only the unsteady and diffusion terms were considered for the model (Chilka and Ranade 2018).

After the steady state solution, the weighted average values of the convective heat transfer coefficient ( $h$ ) were calculated at the concentration profile of the air control volume next to the kernel surface. The  $h$  values were used for the heat flux at the air-kernel interface. The convective mass transfer coefficient ( $h_m$ ) was calculated from its relationship with the effective diffusivity,  $D_e$ , and drying constant,  $k$  (1/s):

$$D_e = \frac{kr^2}{\pi^2} \quad (6)$$

$$k = h_m (A/V) \quad (7)$$

where  $r$  is the characteristic length of the grain (m),  $A$  is the surface area of the kernel (m<sup>2</sup>), and  $V$  is the volume of the kernel (m<sup>3</sup>). The volume and surface area of the kernel were obtained by calculating the cell volume and cell surface area of the 3D object (Adlai kernel) through Fluent, respectively.

The Neumann boundary condition was used at the air-kernel interface of the grain model. This boundary type assumes constant convective heat and mass transfer coefficients (ElGamal et al. 2014). This type of boundary condition is most commonly used due to its simplicity. The  $h_m$  values were assumed to be constant during the drying process and were used for the formulation of the mass flux equation at the interface, while the  $D_e$  values were used as inputs for the material properties of the kernel. The heat and mass flux boundary conditions at the air-kernel interface were as follows:

$$-q = h(T_s - T_\infty) \quad (8)$$

$$-\dot{m} = h_m(M_s - M_e) \quad (9)$$

In Equation 8,  $q$  is the heat flux (W/m<sup>2</sup>),  $h$  is the convective heat transfer coefficient (W/m<sup>2</sup>-K),  $T_s$  is the surface temperature of the kernel (K), and  $T_\infty$  is the drying air temperature (K). In Equation 9,  $\dot{m}$  is the

mass flux,  $h_m$  is the convective mass transfer coefficient (m/s),  $M_s$  is the surface moisture of the grain (kg<sub>water</sub>/kg<sub>dry matter</sub>), and  $M_e$  is the equilibrium moisture content of the grain (kg<sub>water</sub>/kg<sub>dry matter</sub>).

The equations representing the boundary conditions were directly imposed on the air-kernel interface via the named expressions feature of Ansys Fluent. This has become an alternative method to the commonly used CFD technique of writing user-defined functions (UDFs).

The CFD simulations were run on an Acer Nitro AN515-55, Intel Core i7-10750H, 2.60 GHz laptop. For pressure-velocity coupling, the SIMPLE (Semi-Implicit Method for Pressure-Linked Equations) algorithm was used. For pressure, momentum, energy, turbulence, and UDS equations second order discretization scheme was used. Residuals for energy, flow, and UDS equations were set at  $1 \times 10^{-05}$ . Transient simulations were carried out at a time step size of 1s with 10 iterations per time step. This time step size was used in the study because there was observable modeling difficulty when using smaller time steps. For this reason, 1s was considered the minimum time step size for the best possible model accuracy given the study's parameters. The termination time of simulation was 1,265 min, 1,145 min, and 785 min for 30°C, 45°C, and 60°C drying air temperature, respectively. The modeling time went on from 16.5 h to 18 h per simulation.

The volume-weighted averages of the moisture content (UDS Scalar 0) in the kernel domain were calculated at the end of each specified number of time steps to determine the average moisture content of the kernel. These values were recorded to constitute the predicted drying performance by the CFD model.

**Validation of the CFD Models**

To validate the acceptability of the simulated drying curves resulting from CFD modeling, specific

statistical parameters were used as basis. These parameters determine the model accuracy and goodness-of-fit with respect to the experimental data from the thin-layer drying experiment. These statistical parameters were the coefficient of determination ( $R^2$ ), the standard error (S), and the percent mean deviation modulus (P%). The latter must be less than 10% for the model to be acceptable (Mohapatra and Rao 2005).

The  $R^2$  and S values were calculated through the built-in data analysis tools in Microsoft Excel 2016 while the P% values were calculated using the equation:

$$P\% = \frac{100}{n} \sum \frac{abs(M_i - M_{pre})}{M_i} \tag{10}$$

where  $M_i$  is the moisture content from the experiments,  $M_{pre}$  is the predicted moisture content from the CFD simulations, and  $n$  is the number of observations. It should be noted that the predicted and experimental moisture values compared were of the same time step.

**RESULTS**

**Thermophysical Properties of Adlai**

The thermophysical properties of Adlai used in the CFD simulations were estimated using mathematical models involving the food constituents of the product. The estimated values of thermal conductivity, specific heat, and true density of Adlai are presented in Table 3. For drying air temperature range of 30°C to 60°C, the thermal conductivity of Adlai ranged from 0.1491 to 0.1628 W/m<sup>2</sup>-K, the specific heat ranged from 1,986 to 2,024 J/kg-K, and the true density ranged from 885 to 892 kg/m<sup>3</sup>.

**Table 3.** Estimated thermophysical properties, effective diffusivity and convective mass transfer coefficients of Adlai grain at different drying air temperatures.

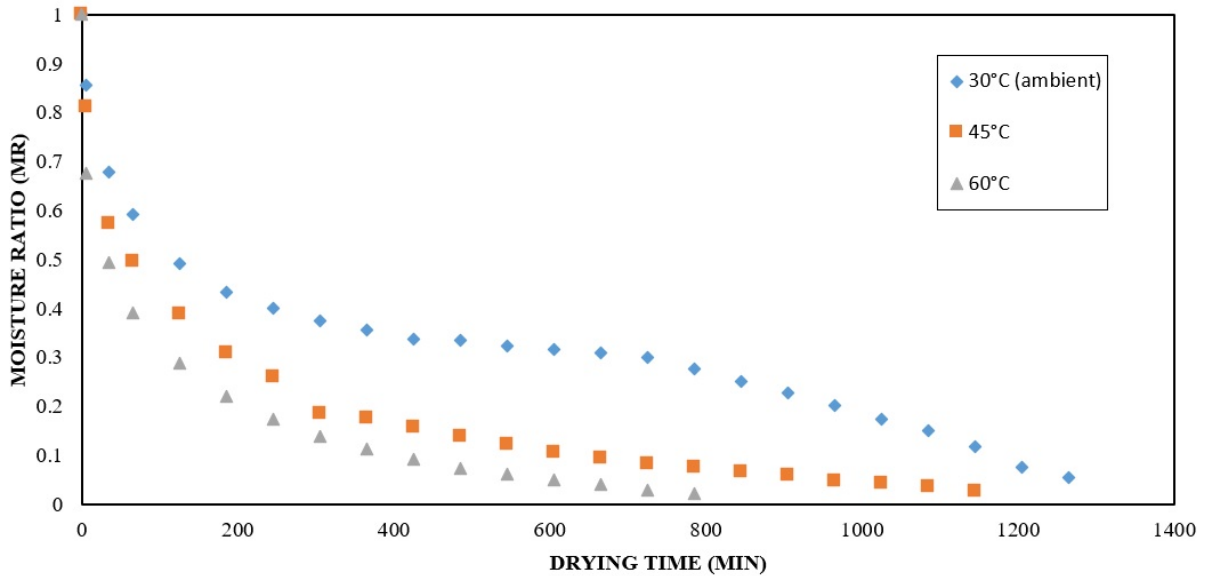
Property	Drying Air Temperature (°C)		
	30	45	60
Thermal conductivity (W/m-K)	0.1491	0.1563	0.1628
Specific heat (J/kg-K)	1986	2006	2024
True density (kg/m <sup>3</sup> )	892	889	885
Effective Diffusivity, $D_e$ (m <sup>2</sup> /s)	$2.772 \times 10^{-12}$	$4.678 \times 10^{-12}$	$7.277 \times 10^{-12}$
Convective mass transfer coefficient, $h_m$ (m/s)	$2.701 \times 10^{-08}$	$4.558 \times 10^{-08}$	$7.090 \times 10^{-08}$

**Thin-layer Drying of Adlai**

A single layer of Adlai grains was subjected to forced heated air drying using 30°C, 45°C, and 60°C at around 10% to 80% RH. The initial moisture content of the sample grains was determined to be 0.1867 MC d.b. The moisture content at specified intervals during drying was measured and recorded manually. The moisture of the grains was expressed as moisture ratio (MR) and was plotted against the drying time to generate the drying curve at different drying air temperatures. Figure 4 presents the drying curves from the thin-layer drying experiments.

Since there are no EMC values for Adlai grain available in literature, the final moisture contents at the end of the thin-layer drying experiments were adopted as approximate EMC values. The EMC values for drying air temperatures of 30°C, 45°C, and 60°C were around 13%, 11%, and 9% MC d.b., respectively.

The average drying time for the grains to reach equilibrium with the drying air was 1,265 min (21 h), 1,145 min (19 h), and 785 min (13 h) for drying air temperatures of 30°C, 45°C, and 60°C, respectively at 10% to 80% RH.



**Figure 4.** Experimental drying curves of Adlai grain at different drying air temperatures.

Table 4 summarizes the model constants and the average drying rates for each drying air temperature. At the drying air temperature range of 30°C to 60°C, the model constants *A* and *k* were

observed to be 0.7386 to 0.8097 and 0.002 to 0.007 per min, respectively. For the same range, the average drying rate was between 0.2479 and 0.6969 % MC d.b. per h of drying.

**Table 4.** Simplified diffusion model constants and average drying rate during thin-layer drying of Adlai.

Drying Air Temperature (°C)	MR = A·exp(-kt)		R <sup>2</sup>	Average Drying Rate (% MC d.b./h)
	A	k (1/min)		
30	0.7386	0.002	0.8843	0.2479
45	0.8097	0.004	0.9451	0.3715
60	0.7878	0.007	0.9260	0.6969

The effective diffusivity ( $D_e$ ) of moisture in the Adlai grain for each drying air temperature was calculated using Equation 6. The  $D_e$  values were calculated by plotting  $\ln MR$  against drying time. The average values of the effective diffusivity are presented in Table 3. In the drying air temperature range of 30°C to 60°C, the effective diffusivity value of Adlai grain moisture ranged from  $2.772 \times 10^{-12}$  to  $7.277 \times 10^{-12}$  m<sup>2</sup>/s.

Temperature has a more significant effect on the drying process of grains compared to the initial moisture content. The effect of temperature on the effective diffusivity was expressed through an exponential model. The values of  $\ln D_e$  were plotted against the absolute drying air temperature. The plot is presented in Figure 5. The exponential model was formulated as follows:

$$D_e = 36.35 \exp(0.0322T_a) \tag{11}$$

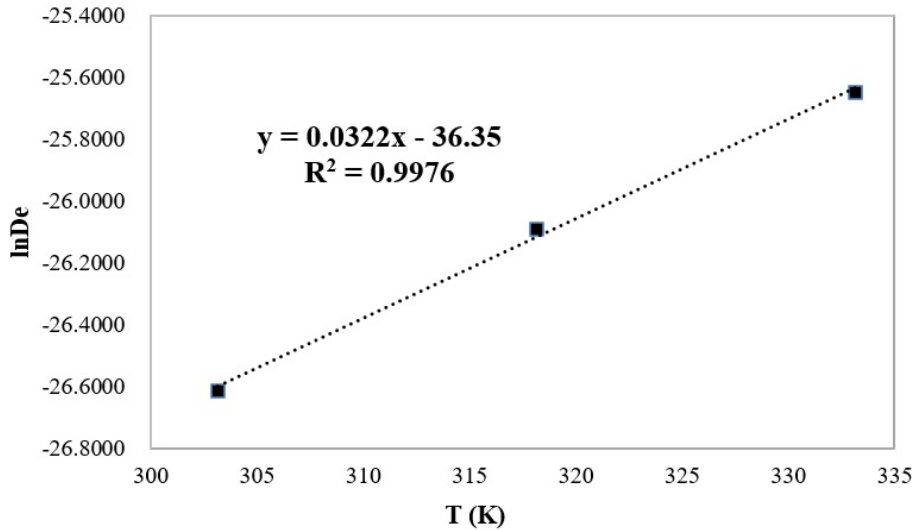


Figure 5. Plot of  $\ln D_e$  vs drying air temperature (K) used in convective drying of Adlai grain.

The value for the diffusivity at infinite high temperature,  $D_o$ , was found to be 36.35, while the value of constant  $c$  was found to be 0.0322. The slope of the curve gives the constant  $c$  while the intercept gives the  $D_o$ .

The average convective mass transfer coefficient ( $h_m$ ) values used to formulate the mass flux equation at the air-kernel interface of the CFD model were derived using Equation 7. The  $h_m$  values used in the study are presented in Table 3. The  $h_m$  values for Adlai kernel varied from  $2.701 \times 10^{-08}$  to  $7.090 \times 10^{-08}$  m/s for 30°C to 60°C drying air temperature. The values presented were necessary inputs for the mass flux equation at the air-kernel interface of the CFD simulations.

**Computational Fluid Dynamics (CFD) Drying Simulation of Single Adlai Kernel**

The average moisture content at each time interval was calculated by determining the volume-weighted average of the UDS Scalar 0 (moisture content) in the kernel domain. Figures 6-8 show the simulated moisture values from the CFD simulation and the experimental moisture values at different drying air temperatures. The drying performance of

Adlai was described by the drying curves obtained by simulating the heat and moisture transfer within the Adlai grain at different drying air temperatures.

The temperature and moisture profiles within the Adlai kernel during drying can also be calculated and shown using the post-processing unit of the Fluent solver. Tables 5 and 6 show these profiles. Both the average temperature and moisture content of the kernel were determined by calculating the volume weighted average of static temperature and UDS Scalar 0 (moisture content), respectively.

**Validation of the CFD Models**

The experimental drying performance of Adlai grain was compared with the predicted drying performance generated by the CFD simulations. The accuracy of the fit was explained through the coefficient of determination ( $R^2$ ) and the standard error (S). The percent mean deviation modulus (P%) was used to define model accuracy. These values are presented in Table 7. At the drying air temperature range of 30°C to 60°C, the  $R^2$  values ranged from 0.94 to 0.98, S values ranged from 0.0018 to 0.0066, and P% values ranged from 6.5% to 8.68%, respectively.

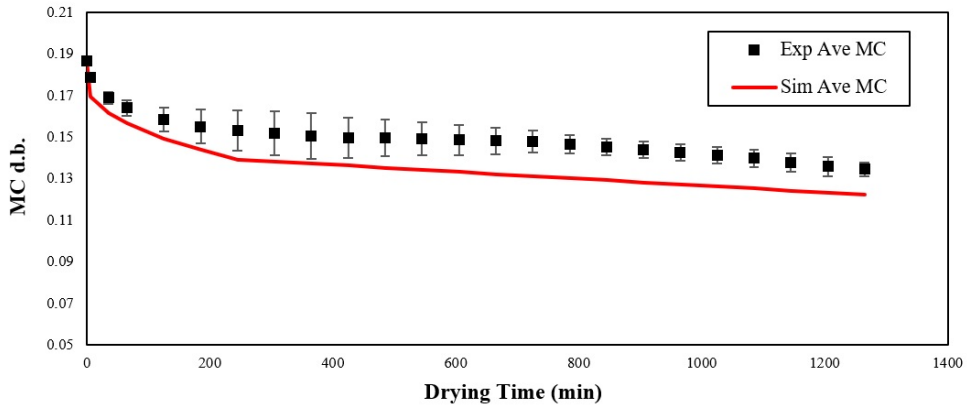


Figure 6. Experimental and predicted drying curves of Adlai at 30°C drying air temperature.

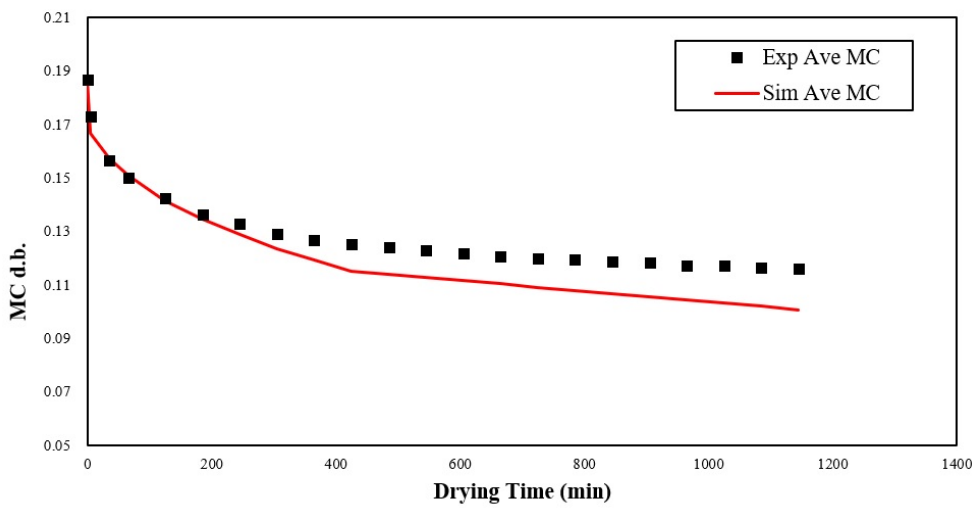


Figure 7. Experimental and predicted drying curves of Adlai at 45°C drying air temperature.

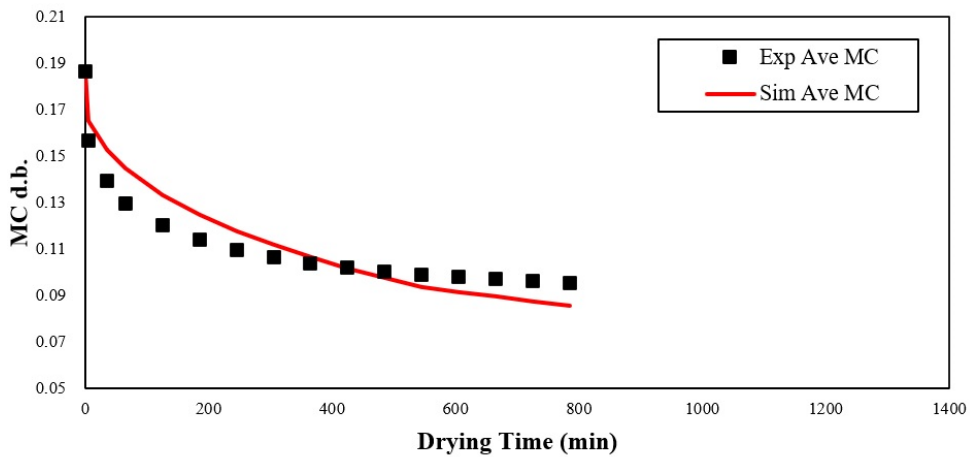
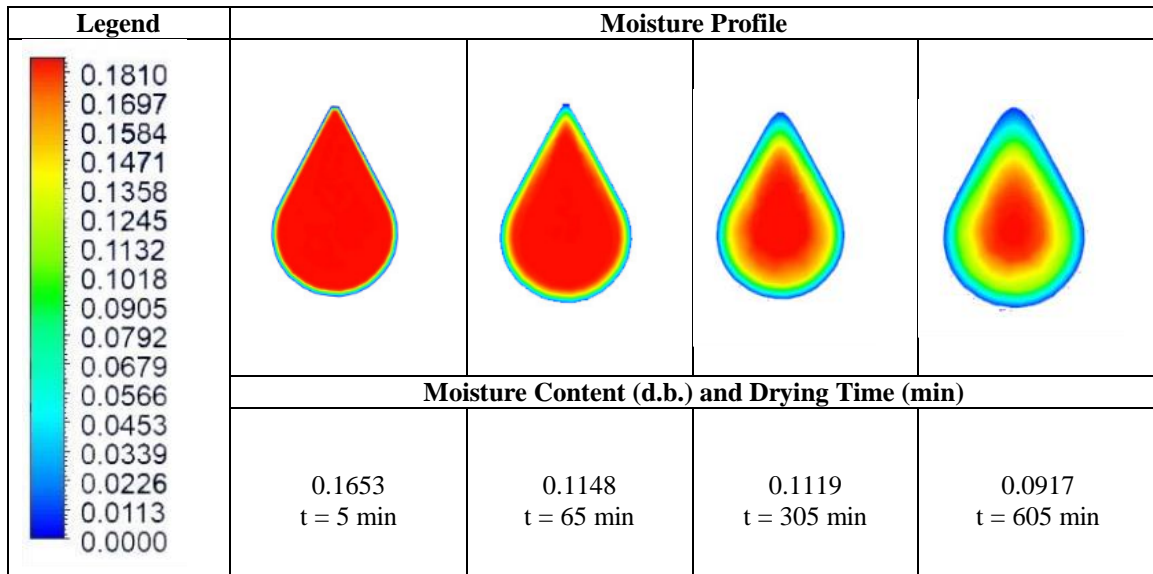
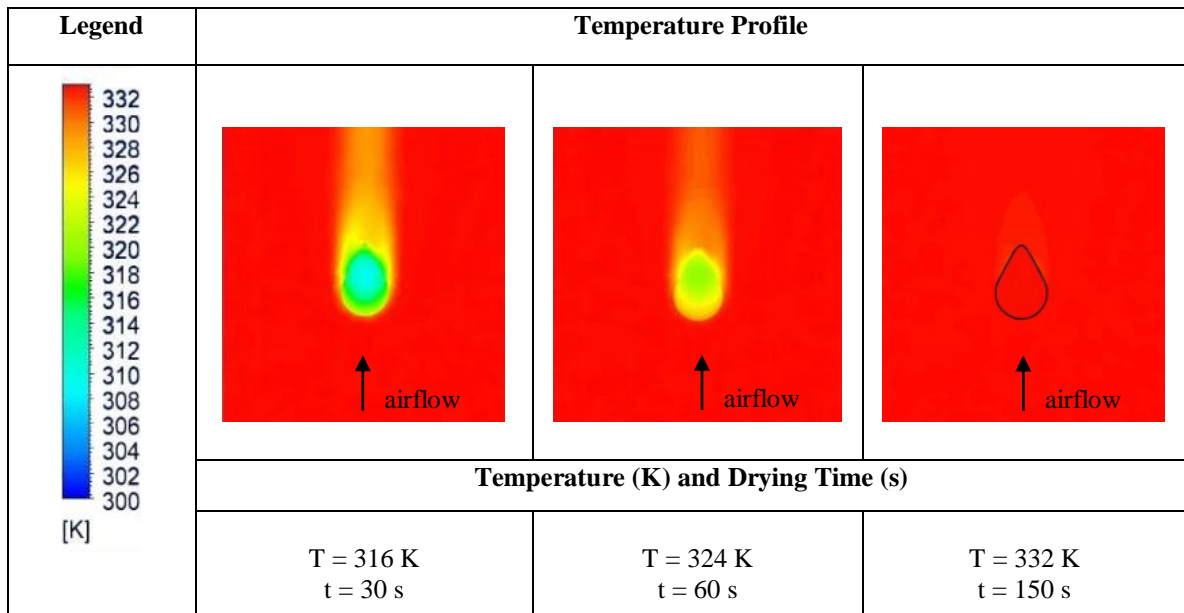


Figure 8. Experimental and predicted drying curves of Adlai at 60°C drying air temperature.

**Table 5.** Moisture profile of Adlai kernel at 60°C drying air temperature.



**Table 6.** Temperature field and profile of Adlai kernel at 60°C (333.15 K) drying air temperature.



**Table 7.** Statistical parameters indicating model accuracy.

Drying Air Temperature (°C)	R <sup>2</sup>	S	P%
30	0.98	0.0018	8.68
45	0.98	0.0027	6.94
60	0.94	0.0066	6.50

## DISCUSSION

### Thermophysical Properties of Adlai

Standard experimental methods for calculating selected thermophysical properties of Adlai were not performed since the study was conducted during the COVID-19 pandemic when there were restrictions to travel, work, and other activities. Estimation of the values through mathematical modeling was the available alternative.

Mathematical modeling of the thermophysical properties of food was used in the study to calculate the values of thermal conductivity, specific heat, and true density of Adlai, which were necessary inputs for the CFD simulations. These values were used in the material properties panel for the solid cell zone of the CFD models, representing the Adlai kernel. It was observed that the estimated thermal conductivity and specific heat of the Adlai grain increased as the assumed kernel temperature increased, while the opposite was observed for the estimation of true density.

### Thin-layer Drying of Adlai

According to the drying curves presented in Figure 3, it was observed that the lower the drying air temperature used, the longer the drying time to reach equilibrium. Higher drying air temperature is usually accompanied by a lower RH, allowing for higher moisture transfer between the kernel and the surrounding air. As shown in Table 4 higher average drying rates (% MC d.b. per h) are achieved at higher air temperatures and lower RH. The average drying rate shows the amount of percent moisture reduction during drying. However, it does not show the high initial moisture removal rate at the beginning of drying since it is averaged over the entire drying time.

The accuracy of the fit of the simplified diffusion model with the experimental drying curves of Adlai grain improved as the drying air temperature increased. This observation was based on the increase in the  $R^2$  value between the models and the experimental drying curves. At higher drying temperatures, the  $R^2$  values were over 0.9 which indicate better fit of the model values with the experimental moisture values from the thin-layer drying experiment.

It was observed that the effective diffusivity of moisture within the grain increased proportionally with the drying air temperature. This was also observed by Jha and Tripathy (2021) in their simulation of heat and mass transfer of solar drying of paddy. For cereal grains, the effective diffusivity found in literature falls within the range of  $10^{-13}$  to  $10^{-08}$   $m^2/s$ . Based on this condition, the estimated  $D_e$  values of Adlai fall within the accepted range. The effective diffusivity of the Adlai grain's moisture is associated with higher drying rates when higher drying

air temperatures are used. This is because more moisture diffuses and leaves the kernel when it is exposed to higher air temperatures at lower RH. The moisture diffusion within the kernel was less affected by the internal resistances at higher drying air temperatures.

The convective mass transfer coefficient ( $h_m$ ) describes the rate of moisture transfer from the surface of the kernel to the drying air, thus affecting the moisture reduction rate within the kernel's volume. It was observed that the higher the temperature of the drying air, the higher the  $h_m$  value. This was also similar with the data of Nguyen et al. (2019) in the study of experimental and numerical investigation of transport phenomena and kinetics of convective shrimp drying. The convective mass transfer coefficient values calculated in this study indicate that the moisture removal at the kernel's surface is more significant and occurs more rapidly at higher drying air temperatures. This is because at higher drying air temperatures, the relative humidity is usually lower, allowing for greater moisture transfer from the grain to the surrounding air.

### Computational Fluid Dynamics (CFD) Drying Simulation of Single Adlai Kernel

Visually, the simulated drying performance of the Adlai grain via CFD modeling fitted closely with the experimental drying performance when plotted together. Agreement between the simulated and the experimental drying curves was better during the initial hours of drying and gradually deteriorated in the latter stages, which was also reported by Chandramohan (2016) in his coupled CFD model for convective drying of a moist object. The numerical models for 30°C, 45°C, and 60°C have the same initial moisture content of 0.1867 and duration of drying time as with the thin-layer drying experiments. However, the final moisture contents were different for simulated and experimental drying curves. The final moisture contents from simulated and experimental drying using 30°C-60°C drying air temperature were 0.1221 to 0.0857 MC d.b. and 0.1344 to 0.0955 MC d.b., respectively. From the data presented, it is clear that the numerical moisture values underestimated the experimental moisture values because the simulations approached the EMC of Adlai at a much faster drying time than the thin-layer drying experiments. For drying temperatures of 30°C, 45°C, and 60°C, the simulation approached Adlai's EMC at around 785 min, 665 min, and 605 min, respectively. The reason behind this could be the constant average values of effective diffusivity and convective mass transfer coefficients. Variable effective diffusivity and convective mass transfer coefficient values would get smaller as the drying process proceeds such that at later stages in drying, the moisture reduction in the model would not be as significant as using constant



values. This would make the agreement between the simulated and experimental moisture values better. Furthermore, the study did not consider physical changes in the grain, which can cause resistance to moisture diffusion and evaporation during the drying phenomenon. Changes such as shrinkage, cracking, and case hardening were not investigated while conducting the thin-layer drying experiments at different drying air temperatures. Nevertheless, these physical changes in the grain have been found out to have an effect on the rate of moisture removal during drying.

Aside from the drying performance over time, the moisture and temperature profiles during the CFD simulations of convective drying of Adlai grain can be visualized through Ansys Fluent. The apex of the Adlai grain model was positioned away from the drying air inlet of the simulation such that the drying air hit the bottom of the grain first. This was the geometric presentation chosen for the study since different orientations of the kernel to airflow direction did not show any effect on the average moisture content of the grain with similar drying conditions. This was also reported by ElGamal et al. (2014) for the convective drying simulation of a single rice kernel. It was assumed that moisture transfer occurs at the kernel's surface, creating areas of varying moisture content within the kernel in the process, as shown in Table 5. It was observed that during the simulation of convective drying of a single kernel of Adlai, moisture reduction occurs much faster near the kernel's surface than the core. As shown in Table 5, as the drying time increased, the moisture content at the center of the grain decreased as well. The moisture diffused from the center of the grain to the surface, where moisture vaporization took place, leaving the regions near the surface with low moisture content. This was also observed in the numerical study of heat and mass transfer of peanut drying (Chen et al. 2023). Over time, the moisture at the core of the grain reduced further, as indicated by the recession of high moisture color bands in the model's geometric profile. The reduction of moisture content in the kernel affected the overall average moisture content of the entire kernel. This phenomenon eventually gave the final moisture content of the grain at the termination of the drying simulation. It was observed that high moisture concentration at the kernel's core would also give a high average kernel moisture value.

Heat transfer during the drying simulation occurred significantly faster than the mass transfer. This was also parallel in the works of Chilka and Ranade (2018) and Nguyen et al. (2019) in drying almond kernel and shrimp, respectively. It took only a short period of time for the kernel temperature to achieve thermal equilibrium with the drying air temperature as shown in Table 6. Over time, the color gradient in the kernel changed as the temperature

increased and became approximately equal to the drying air temperature. At 150 s and further, the kernel temperature is in equilibrium with the drying air temperature. Colored streaks formed around the kernel, indicating its solid-fluid interaction with the moving drying air.

The heat transfer mechanism in Fluent is dependent on the thermophysical properties given to the solid (kernel) cell zone and the fluid (drying air) cell zone. These separate cell zones interact at the air-kernel interface, which has been treated as a coupled boundary between the solid and fluid cell zones.

Moisture and temperature profiles within the kernel via CFD modeling offer essential insights regarding its drying behavior. Moisture variation within the kernel allows the management of potential hygroscopically induced fissuring or cracking. Thus, the decision to incorporate tempering in the drying process could be made if necessary. Also, analyzing the moisture movement and variation within the kernel would help food technologists understand the diffusion mechanism of moisture from the endosperm to the husk to complete the moisture removal process at the surface. However, in the case of Adlai, this would entail further investigations of the properties of its endosperm, bran layer, and husk. When it comes to the temperature profile of grain, understanding the temperature variation within the kernel helps manage grain drying while minimizing the effects of thermal stress that may be caused by rapid drying. All of these are important to the food technologist since the drying rate of grain greatly affects the final quality of the kernels (Bonazzi et al. 1997).

### Validation of the CFD Models

As shown in Table 7, the simulation for 60°C drying air temperature was the lowest in terms of  $R^2$  value since it is the only one lower than 0.98. However, both models for 30°C and 45°C drying air temperature have higher P% values, which is indicative of higher deviation from the experimental data. In general, all the predicted models from the CFD simulations were acceptable due to their high  $R^2$  values, low S values, and P% values of less than 10%. This means that CFD modeling is an acceptable technique for predicting the drying performance of Adlai grain under several convective drying temperatures. This method could be used apart from full-scale experimentation, which is deemed laborious, expensive, and time-consuming. CFD could be used to explore the behavior of Adlai under several other drying conditions.

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## ETHICAL CONSIDERATIONS

No human or animals were involved nor harmed in the conduct of this study.

## DECLARATION OF COMPETING INTEREST

The authors of this study declare that there are no competing interests of any kind with other authors.

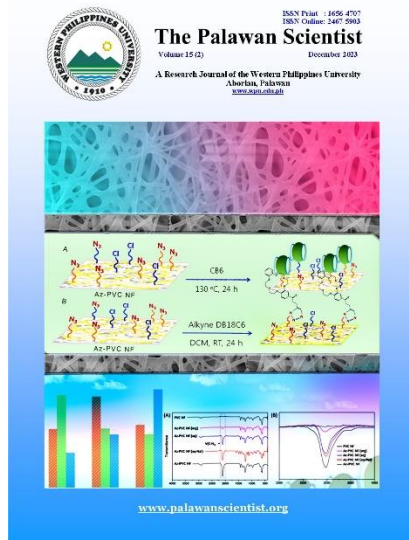
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# Seismic-induced landslide hazard analysis of the recreational area of the Makiling Botanic Gardens, Los Baños, Laguna, Philippines

Angelo Manuel A. Castillo\*, Jedidiah Joel C. Aguirre, Carmina B. Borja and Andre C. Cruz

University of the Philippines Los Baños, Los Baños, Laguna, Philippines

\*Correspondence: [aacastillo3@up.edu.ph](mailto:aacastillo3@up.edu.ph)

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## ABSTRACT

The Makiling Botanic Gardens (MBG) is an educational and eco-tourism site located within the vicinity of Mount Makiling, a dormant volcano susceptible to landslides, and located less than 20 km away from at least two tectonic faults. The study assessed the seismic-induced landslide hazards in the recreational area of MBG mainly based on three parameters: factor of safety (FS), yield acceleration ( $a_y$ ), and slope displacement ( $D_n$ ). Dynamic Cone Penetration Test (DCPT) served as the primary method for obtaining the shear strength of the soil within the area. Meanwhile, other relevant parameters, including material unit weight and internal angle of friction, were determined through laboratory testing, correlations, and available data and models. Moreover, the Peak Ground Acceleration (PGA) and Spectral Acceleration (SA) values were obtained through an interpretation of a probabilistic seismic hazard analysis (PSHA) study of the Philippines. Using the ArcGIS software, an FS map, yield acceleration map, and slope displacement map were generated. Results showed that the slopes within the area exhibit static stability. However, in the event of an earthquake, it is expected that the amplified PGA will exceed the expected yield acceleration values of the slopes in some parts of the recreational area, which could result in dynamic instability and slope displacements as high as 130 mm. Thus, the recreational area of MBG possesses some susceptibility towards a seismic-induced landslide.

**Keywords:** dynamic cone penetration test, factor of safety, Newmark displacement, peak ground acceleration, spectral acceleration

## INTRODUCTION

Landslides are natural hazards generally defined as the downhill movement of soil, rock, and debris on sloping and relatively flat areas such as mountainous regions. They may occur without warning and may be triggered by earthquakes, heavy rainfall, and other activities that can disrupt and

displace a mass of soil and rock which are everyday events in the Philippines, especially in its mountainous regions (Lapitan et al. 2006). Mount Makiling, a dormant but potentially active volcano in the Philippines with an elevation of 1,090 m above mean sea level, is a mountainous region housing several constituents of the Los Baños municipality (KBA 2006). Moreover, according to the ASEAN Centre for

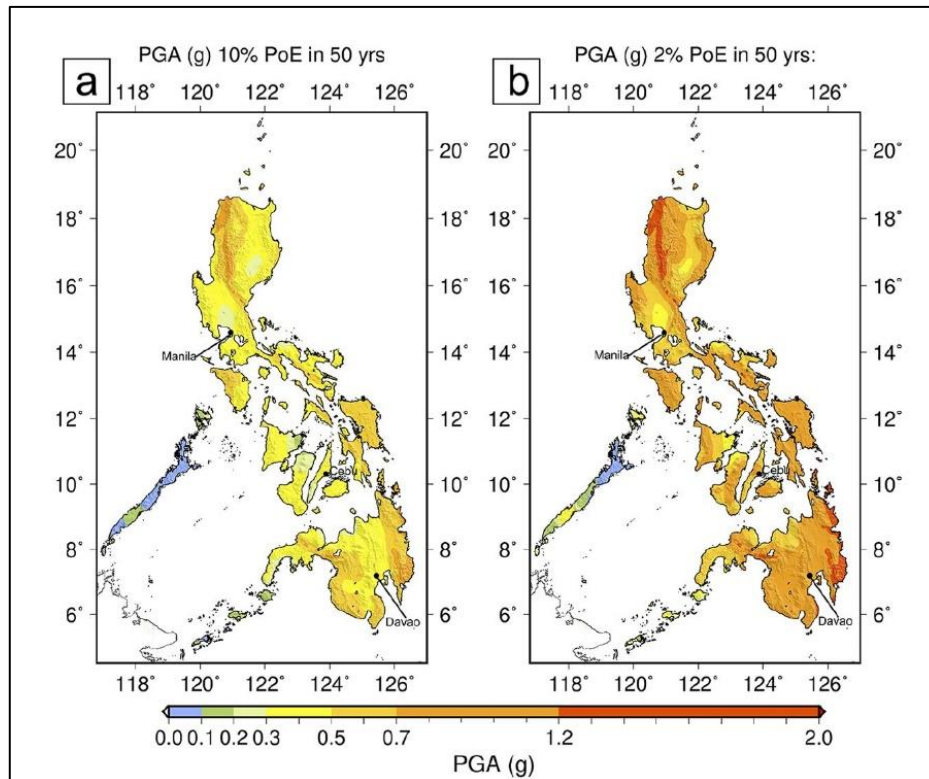


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Biodiversity (2013), within Mount Makiling lies the Mount Makiling Forest Reserve (MMFR), which houses the Makiling Botanic Gardens (MBG), established under Republic Act 3523 and Makiling Center for Mountain Ecosystems (MCME), to serve as an educational and eco-tourism site for the public. However, MBG also houses several sloping areas of soil and rock that may or may not be susceptible to seismic-induced landslides. According to the Philippine Institute of Volcanology and Seismology (PHIVOLCS), as of 2013, the MMFR is located approximately 17 km away from the nearest Valley Fault System (VFS), in which the fault system extends to the municipality of Calamba. Thus, to help establish proactive measures against a possible future destructive event, a seismic hazard analysis and mapping may be deemed necessary.

Seismic hazard is a classification of hazards that refer to the statistical probability of the occurrence of seismic activities, such as earthquakes, within a geographical area (Rojas 2016). In this regard, seismic-induced landslides may occur at any point in time, considering that there is still lack of technology and data for accurate predictions of earthquake occurrence. Concerning seismic activities, peak

ground acceleration (PGA) is the maximum change in speed experienced by a particle in the ground. In contrast, spectral acceleration (SA) is the change in speed experienced by an infrastructure (Bradley 2011). Both PGA and SA may be expressed in  $g$  or the standard acceleration value due to gravity ( $9.81 \text{ m}\cdot\text{s}^{-2}$ ). Additionally, PGA may serve as a hazard index for short buildings with a maximum of seven stories, on average, in which PGA focuses more on peak ground motion. Meanwhile, SA may serve as a hazard index for tall buildings in which SA focuses more on building behavior. Furthermore, a probabilistic seismic hazard model of the Philippines was generated by the study of Peñarubia et al. (2020). In the study, the Philippine archipelago is defined as tectonically complex and seismically hazardous due to its geographical position and its formations. In the study, a PGA map was generated for the Philippines with a 10% probability of exceedance in a period of 50 years, as shown in Figure 1. For the municipality of Los Baños in the province of Laguna, the PGA for the area ranges from 0.3  $g$  to 0.5  $g$ . In the case of MBG, a protected sanctuary, a non-intrusive seismic hazard analysis method was preferred to conform with related laws and official memorandums.



**Figure 1.** Mean peak ground acceleration (PGA)map for the Philippines with 10% and 2% probability of exceedance in 50 years.

The Dynamic Cone Penetration Test (DCPT) is a non-intrusive and non-destructive soil investigation procedure performed to determine the strength of the soil and soil layers present within the study area through the Dynamic Cone Penetration (DCP) – n values (VerTek 2021). In a seismic-induced landslide study in Indonesia by Kiyota (2020), DCPT provides an inexpensive and portable procedure for obtaining in-situ soil data regardless of slope stability conditions. The method is even more crucial in obtaining soil data after the events of a landslide in which slopes are unstable. The DCPT is performed following the standards provided by the ASTM D6951. According to Look (2007), the DCP is equivalent to a third of the energy of the Standard Penetration Test (SPT) in which the DCP – n values can be correlated with the soil strength in terms of cohesion, as shown in Table 1. The DCPT may be used for obtaining soil behavior data without procuring soil samples and disturbing the stratigraphy of a protected site such as MBG.

Since the Makiling Botanic Gardens (MBG), especially the recreational area, is a tourist destination and attracts several visitors annually, it would be beneficial for the area to have established safety measures against seismic-induced landslides, especially considering the unpredictable nature of earthquakes. The main objective of this study is to assess the seismic-induced landslide hazards in the 3.5 ha recreational area of the MBG based solely on projected seismic-induced slope displacements. Specifically, it aimed to identify the geology and soil composition within the recreational area of MBG based on available soil investigation reports and data; calculate the static Factor of Safety (FS) and yield acceleration at different points and areas within the recreational area of MBG using computed and pre-determined relevant soil parameters; and calculate the slope displacements due to seismic-induced landslides at different points and areas within the recreational area of MBG.

**Table 1.** Strength parameters from DCPT Data. Source: Look 2007.

Material	Description	DCP – n (Blows/100 mm)	Strength
Clays	Very Soft	0-1	$C_u = 0-12$ kPa
	Soft	1-2	$C_u = 12-25$ kPa
	Firm	2-3	$C_u = 25-50$ kPa
	Stiff	3-7	$C_u = 50-100$ kPa
	Very Stiff	7-12	$C_u = 100-200$ kPa
	Hard	>12	$C_u > 200$ kPa
Sands	Very Loose	0-1	$\Phi < 30^\circ$
	Loose	1-3	$\Phi = 30-35^\circ$
	Medium Dense	3-8	$\Phi = 35-40^\circ$
	Dense	8-15	$\Phi = 40-45^\circ$
	Very Dense	>15	$\Phi > 45^\circ$
Gravels, Cobbles, Boulders		>10	$\Phi = 35^\circ$
		>20	$\Phi > 40^\circ$
Rock		>10	$C' = 25$ kPa, $\Phi > 30^\circ$
		>20	$C' = 50$ kPa, $\Phi > 30^\circ$

## METHODS

### Descriptive Summary of the Study Area

The study focused only on the 3.5 ha recreational area of MBG, (Figure 2). The area was plotted using WGS84 coordinates and described using maps regarding the local geology and tectonic setting of the MMFR. The MBG has a WGS84 coordinates of 14.1566° North and 121.2342° East. For the procurement of existing pertinent data regarding the

geomorphological characterization of MBG, the following documents were requested and obtained from the Makiling Center for Mountain Ecosystems (MCME): Land cover map of the MMFR ASEAN Heritage Park, Soil cover map of the MMFR ASEAN Heritage Park, and DEM of the MMFR ASEAN Heritage Park. Through ArcGIS, the elevation map at 5 m intervals, slope angle map at 5° intervals, land cover map, and soil cover map for the 3.5 ha study area were extracted.



**Figure 2.** Plot and location of study area.

### Determination of Soil Properties

The Dynamic Cone Penetration Test (DCPT) served as the primary procedure for obtaining primary data to determine needed soil parameters. For the 3.5 ha recreational area, 43 test points were determined and plotted to represent the study area, as shown in Figure 3. For each test point, DCPT was performed once in which an 8 kg hammer was repeatedly raised and dropped at a certain height until it reached either a depth of one meter or refusal.

The penetration depth and number of blows were recorded for each test point. Afterwards, the Dynamic Cone Penetration (DCP) Index and DCP *n*-values for each test point were determined based on the penetration depth at each blow, resulting in generating a DCP profile for each test point. Using Table 1, the DCP – *n* values were correlated with cohesion ( $c'$ ) values. Due to the presence of clayey soil within the study area and as per Table 1, only cohesion values can be correlated with DCP-*n* values. For conservative measures, the lowest cohesion value obtained within a test point served as the governing shear strength for that test point. Then, the shear

strength values were interpolated using the Kriging method of ArcGIS to generate a shear strength map for the study area. Meanwhile, slope angles ( $\alpha$ ) were determined based on the generated slope angle map for the same area.

### Laboratory Determination of Other Soil Properties

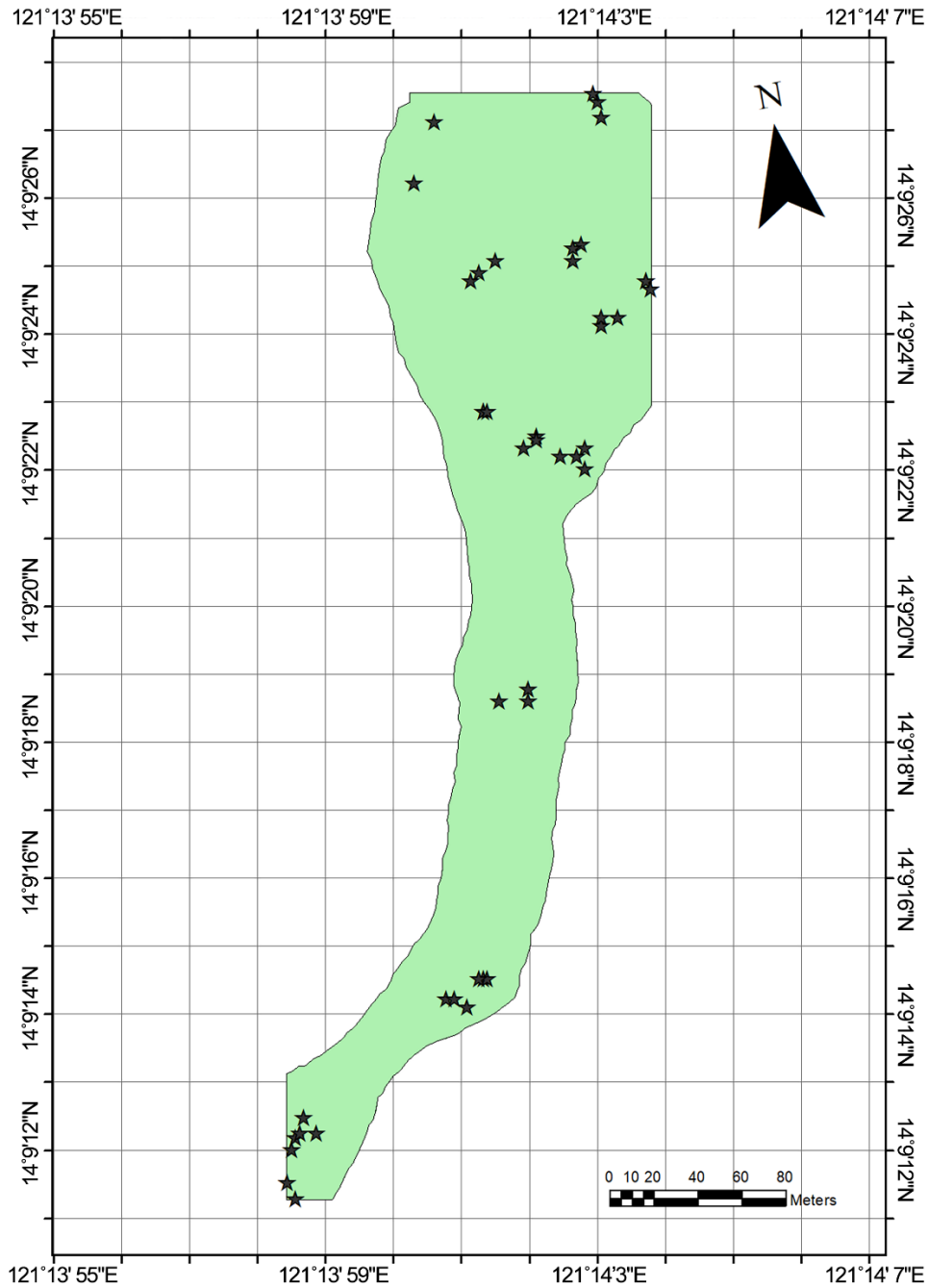
For each of the 43 test points within the 3.5 ha recreational area of MBG, soil samples with respective masses of at least 100 g were obtained for laboratory testing at the Department of Civil Engineering (DCE) in the University of the Philippines Los Baños (UPLB). In the laboratory, the procured soil samples were oven-dried for at least 24 h, pulverized with a hammer, and sieved using Sieve No. 200 and mechanical shaker for 3 min.

After sieving, the mass retained in Sieve No. 200 was measured and recorded. Then, the percentage composition of the recorded mass to the overall mass of the sieved soil was computed and determined. The said parameter served as the coarse percentage of the soil sample, indicating the percentage of the sample with a particle diameter of more than 75  $\mu\text{m}$ . If the

coarse percentage of a sample yielded a value of at least 30%, the internal angle of friction ( $\phi'$ ) was determined through correlation with cohesion values using Table 2.

Meanwhile, for the determination of the material unit weight of the soil ( $\gamma$ ), a small cylindrical

container was prepared and filled with a compacted sample. The material unit weight was then computed and determined based on the contained mass of the container (in kilograms) and the inner volume of the container (in cubic meters) using Table 3.



**Figure 3.** Map of the 3.5-hectare recreational area of Makiling Botanic Gardens showing the sampling sites (★).

**Table 2.** Strength values for cohesive soil. Source: Look 2007.

Type	Description	Effective Cohesion (kPa)	Friction angle (degrees)
Cohesive	Soft- Organic	5-10	10-20
	Soft- Non-organic	10-20	15-25
	Stiff	20-50	20-30
	Hard	50-100	25-30

**Table 3.** Corresponding material unit weight values for DCP – n values. Source: Herath 2005.

$\gamma$ (kN m <sup>-3</sup> )	DCP – n value (mm/blow)
15.3	18.4
17.1	15.0
17.8	17.0
18.9	16.7
17.6	22.5
18.7	13.0
16.8	23.0
15.8	36.0
16.1	18.8
16.7	29.0
16.3	26.1
16.1	27.0
14.1	13.5
14.1	15.2
13.1	15.0
18.2	14.3
17.0	17.2
15.5	6.54
18.4	15.9
18.0	11.7
18.6	23.3
16.8	63.7
17.3	10.6
18.4	26.0

**Determination of Factor of Safety**

The Factor of Safety (FS) serves as a measure of slope stability within an area of interest. It is the ratio between the available shear strength of a slope and the shear forces acting on the slope that may cause landslides. Generally, an FS of greater than unity indicates a stable slope, while an FS of less than unity indicates an unstable slope prone to slope failure (Kramer 1996). The previously generated land cover map, soil cover map, slope angle map, elevation map, and shear strength map for the study area were intersected together with a set of 10 m by 10 m grids. Afterward, the newly generated map divided the

recreational area of MBG into cells in which each cell had a corresponding set of attributes. Using the said attributes, the FS at each cell of the recreational area of MBG was determined using Equation 1, as shown below:

$$FS = \frac{c'}{\gamma t \sin(\alpha)} + \frac{\tan \phi'}{\tan(\alpha)} \tag{1}$$

where:

- FS is the static factor of safety;
- c' is the effective soil cohesion;
- $\gamma$  is the material unit weight of the soil;
- t is the slope-normal thickness of the failure slab (3 m for shallow landslides);
- $\alpha$  is the slope angle; and
- $\phi'$  is the internal angle of friction of the slope.

The resulting values for the FS at each cell of the study area served as the basis for generating an FS map for the said area. The generated map provided a general overview of the static slope stability of the recreational area of MBG regarding the FS.

**Yield Acceleration**

In Newmark’s sliding block analysis, both static and dynamic forces are incorporated to analyze slope stability, which is indicated by the yield acceleration ( $a_y$ ). The yield acceleration is the minimum pseudo-static acceleration that can induce slope instability. Moreover, a mass that reaches its yield acceleration results in an FS of unity, implying that areas with yield acceleration values less than the peak ground acceleration (PGA) are susceptible to seismic-induced landslides. Regarding the FS map of the study area, the yield acceleration values, expressed in g, at each cell were determined using Equation 2, as shown below:

$$a_y = (FS - 1) \sin(\alpha) \tag{2}$$

where:

- $a_y$  is the yield acceleration of the slope;
- FS is the factor of safety of the slope;
- $\alpha$  is the slope angle.

After determining the corresponding yield acceleration values for each cell of the study area, a yield acceleration map for the recreational area of MBG was generated, thus providing an overview of the susceptibility of MBG towards landslides in terms of induced accelerations.

**Determination of Slope Displacement**

As mentioned previously, yield acceleration is the minimum value of acceleration to cause slope



instability. Slope failures, such as landslides, result in permanent displacements. The Newmark method for measuring displacement considers the effects of dynamic forces due to seismic activities. For this study, the pseudo-empirical predictive formula for Newmark displacements derived by Veylon et al. (2017) was used, as shown in Equation 3. For the values of PGA and spectral acceleration (SA) of the study area, the study adopted the maximum average PGA, with a 475 yr return period, for the study area which was from the study of Peñarubia et al. (2020). Furthermore, the

adopted PGA value was amplified based on the specifications of the American Society of Civil Engineers (ASCE) Section 7-05 and based on the site classification of the study area. Meanwhile, the recreational area of MBG was generated based also on ASCE Section 7-05, a spectral acceleration design response spectrum, with a 475 yr return period. Lastly, the corresponding SA values at each cell of the study area was based from the Spectral Acceleration Maps provided by DOST-PHIVOLCS (Figures 4 and 5).

$$\ln(D_n) = -0.56 - 2.23 \frac{a_y}{PGA} + 2.8 \ln\left(\frac{SA(1.5T_1)}{PGA}\right) \tag{3}$$

where:

- D<sub>n</sub> is the Newmark displacement;
- a<sub>y</sub> is the yield acceleration;
- g is the acceleration under the effects of gravitational force;
- PGA is the peak ground acceleration;
- SA is the spectral acceleration; and
- T<sub>1</sub> is the initial fundamental period of the structure.

T<sub>1</sub> measures the vibratory characteristics of the slopes, which can be associated with the height of the slope and its stiffness. Thus, T<sub>1</sub> was determined

$$T_1 = \frac{2\pi}{2.4} \frac{H}{\sqrt{G_{max}/\rho}} \tag{4}$$

where:

- H is the height of the slope;
- G<sub>max</sub> is the maximum shear modulus of the material; and
- ρ is the density of the material of the slope.

using the formula from Veylon et al. (2017), shown in Equation 4 below:

After determining the projected slope displacements at each cell of the recreational area of MBG, a slope displacement map was generated using ArcGIS. The map provided estimates of permanent displacements at different points and areas of the 3.5 ha recreational area of MBG in the event of a seismic-induced landslide occurrence.

## RESULTS

### Geology and Soil Composition of the Recreational Area of the Makiling Botanic Gardens

Makiling Botanic Gardens (MBG) is an eco-tourism site within the Mount Makiling Forest Reserve (MMFR) ASEAN Heritage Park. Figure 6 shows the local geology of the MMFR ASEAN Heritage Park, which defines the lithology and structures present within the area. To further support the implications of the local geology of the MMFR ASEAN Heritage Park, the presence of a particular type of soil suitable for plant growth and development was considered since

the recreational area of MBG contains a diverse range of flora and fauna. For general visualization of the soil composition of the study area, Figure 7 shows the soil cover map for the recreational area of MBG, which was extracted from MMFR ASEAN Heritage Park soil cover map provided by the Makiling Center for Mountain Ecosystems (MCME). Concerning the soil cover map of the study area, Figure 8 shows the land cover map for the same area extracted from the provided MMFR ASEAN Heritage Park land cover map by MCME.

A Digital Elevation Model (DEM) from the National Mapping and Resource Information Authority (NAMRIA 2015) for the MMFR ASEAN Heritage Park was provided by MCME, which served as the basis for mapping the elevations and slope angles of the study area in which Figures 9 and 10 show the varying elevations and slope angles present within the area, respectively. At each of the 43 test points, a Dynamic Cone Penetration Test (DCPT) was performed, in which each test point generated a Dynamic Cone Penetration (DCP) n-value profile

(Figure 3). Then, the obtained DCP n-value profile was correlated to shear strength values (Table 1). The shear strength values at each point were then correlated to generate a shear strength map for the study area (Figure 11).

#### Static Factor of Safety and Yield Acceleration

Through ArcGIS, all the previously generated maps were combined and intersected. Fishnets were established using 10 m by 10 m grids, resulting in a generated map with 1,388 individual cells with their own unique sets of attributes. The static FS of each of the 1,388 cells was determined using Equation 1 and added as an attribute, resulting in an FS map shown in Figure 12. After determining the FS values at each cell and using Equation 2, the yield acceleration values for each of the 1,388 cells of the plotted recreational area of MBG were determined and added as an attribute, resulting in a yield acceleration map (see Figure 13).

#### Seismic-induced Slope Displacements

For the determination of projected slope displacements due to seismic-induced landslides, the peak ground acceleration (PGA) and spectral acceleration (SA), with a 475-year return period, for the recreational area of MBG were determined and considered. Considering the amplified PGA for the recreational area, a conservative value of  $0.85g$  was considered for the PGA of the said area. Meanwhile, using the generated site-specific spectral acceleration design response spectrum for the same area, each of the 1,388 cells of the study area was assigned a spectral acceleration value based on the fundamental period computed for each cell using Equation 4. Finally, using Equation 3, the slope displacement for each cell was determined and assigned as an attribute, resulting in the slope displacement map (see Figure 14).

## DISCUSSION

#### Geology and Soil Composition of the Recreational Area of the Makiling Botanic Gardens

The local geology of the MMFR ASEAN Heritage Park is composed mainly of volcanic materials and residues based on Figure 6. The subsurface material present within the area is part of the Macolod Volcanic Complex, which is composed of volcanoes and volcanic features originating from the Pliocene era (5.3 mya) to the Pleistocene era (11 mya) in which Mount Makiling is part of the said complex (Peña 2008). Moreover, due to the active volcanic processes circulating within the area, tectonic

features are influenced by these processes, resulting in a volcanotectonic fault located around 10 km east of the campus of the University of the Philippines Los Baños (UPLB). Additionally, the soil composition for the recreational area of MBG is mainly Macolod clay, a loam-clay soil (Figure 7) which is a common variety of soil in the Philippines that is suitable for agricultural purposes due to its high nutrient value and water retaining capacity (Montecillo 1983). However, Macolod clay is a relatively soft soil, which may indicate an increased risk for seismic-induced landslides. Additionally, the majority of the recreational area of MBG is built-up, implying the presence of structures such as roads, pavements, and buildings (Figure 8). The presence of these structures may further indicate that slope stabilization measures within the vicinity of the structures have been established and implemented. According to the California State of Water Resources Control Board (2001), slope stabilization measures are established in sloping areas prone to surface runoff and erosion, such as Mount Makiling. The presence of several infrastructures in such an environment may imply that stabilization measures were established to increase the slope's factor of safety and to stabilize the soil. However, based on existing records and observations, some portions of the built-up areas have exhibited signs of degradation due to age and environmental factors. Such a case may further result in outdated structures with unstable frameworks, which may serve as a factor for increased disaster risk in the event of seismic-induced landslides.

Regarding elevations in the study area, the recreational area of MBG exhibits increasing elevation from the entrance (northeast portion of the area) to the *Dipterocarp* area (southern portion of the area) (see Figure 9). The area yielded a maximum elevation of 140 m above mean sea level and a minimum elevation of 85 m above mean sea level, resulting in a maximum slope height of approximately 55 m. Moreover, the immediate changes in elevation in the *Dipterocarp* area and Molawin Creek area (northwest portion of the area) indicate the presence of relatively steep slopes within the said areas. Furthermore, the slope angles within the recreational area range from a minimum angle of  $0^\circ$  to a maximum angle of  $20^\circ$  (Figure 10). This indicates that the highest slope angles of the recreational area are present within the Molawin Creek area (Figure 9). The slope angles were one of the main parameters of interest to determine the static Factor of Safety (FS) values within the recreational area. Lastly, the shear strength of the Macolod clay within the recreational area of MBG ranges from a minimum shear strength of 12 kPa to a maximum shear strength of 26 kPa (Figure 11).

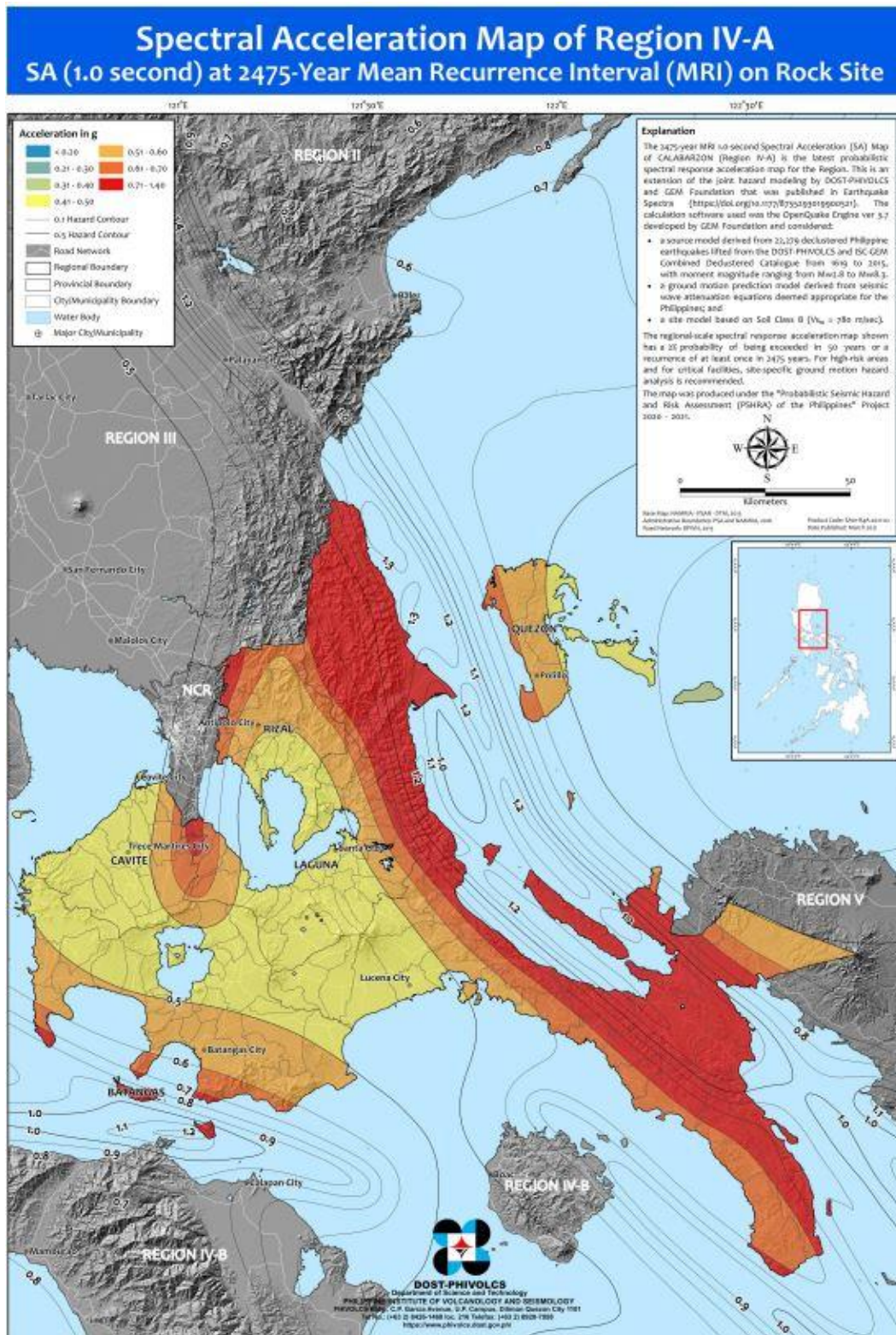


Figure 4. Spectral acceleration map of Region IV-A at 1.0 second.

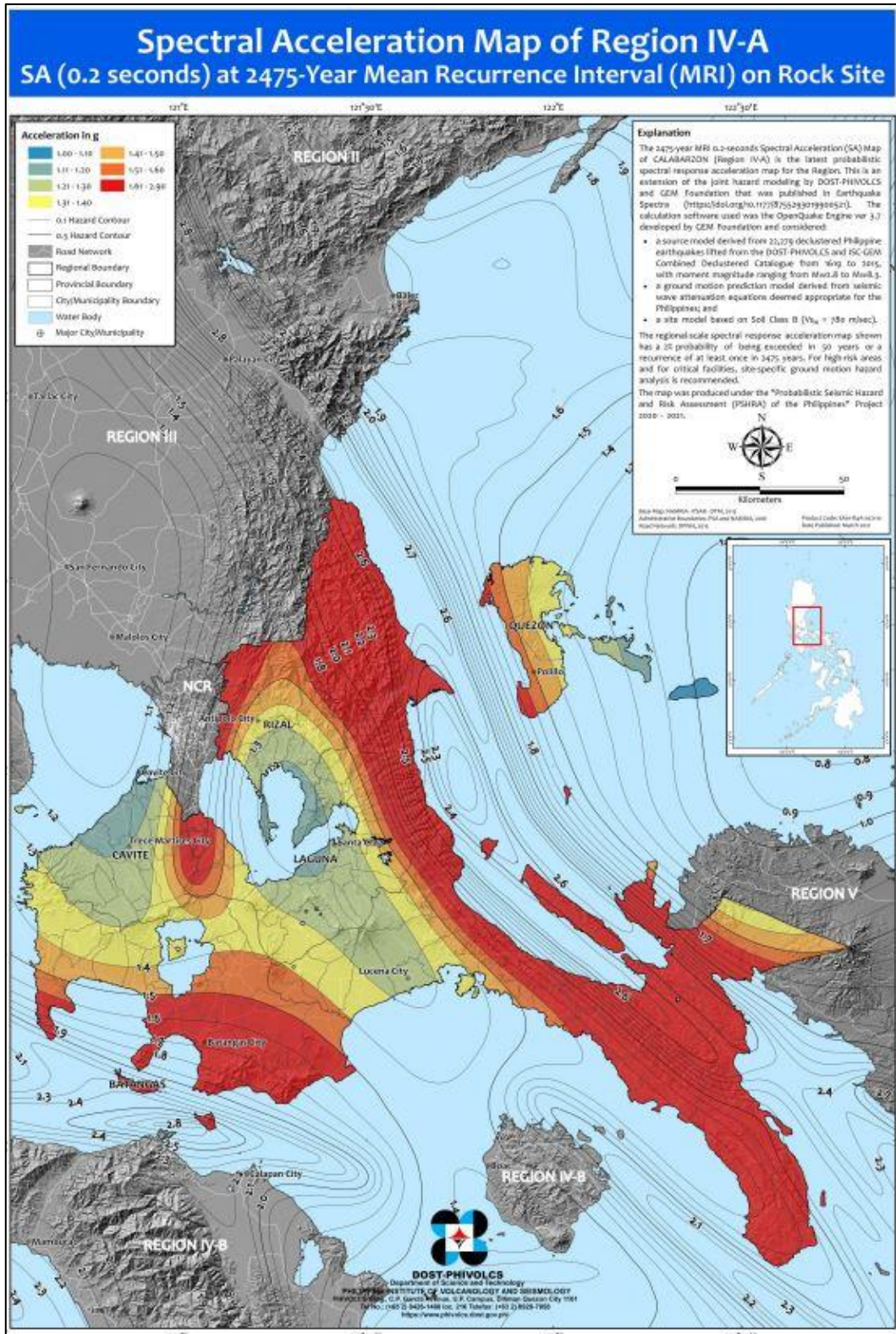


Figure 5. Spectral acceleration map of Region IV-A at 0.2 second.

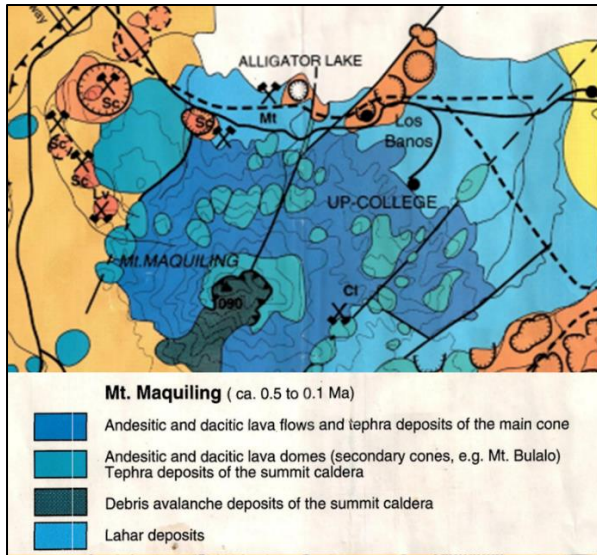


Figure 6. Local geology of Mount Makiling Forest Reserve ASEAN Heritage Park.

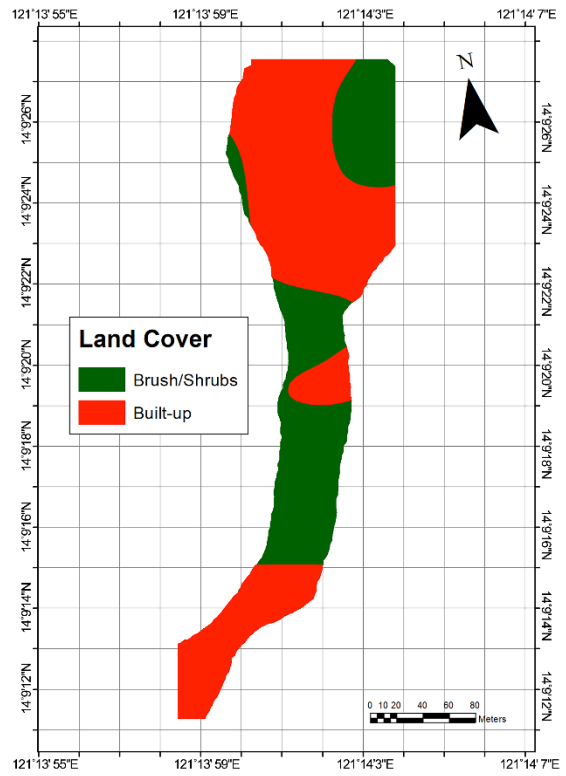


Figure 8. Land cover map of the recreational area of Makiling Botanic Gardens.

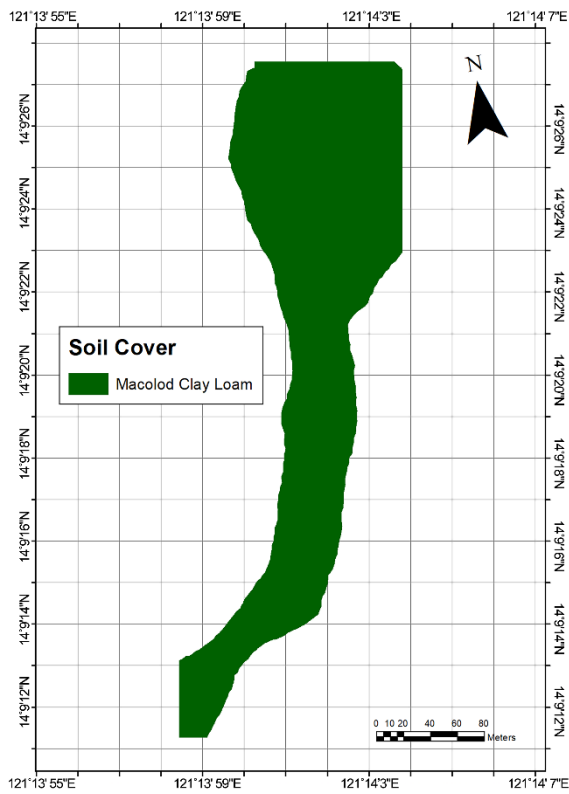


Figure 7. Soil cover map of the recreational area of Makiling Botanic Gardens.

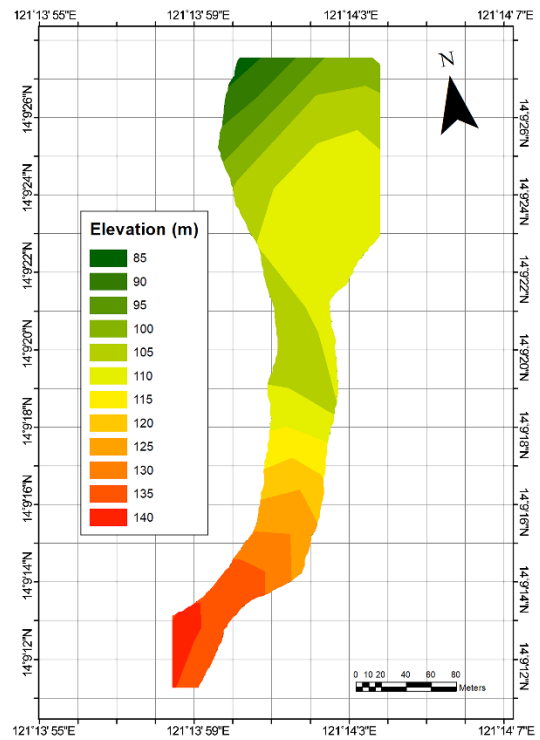


Figure 9. Elevation map of the recreational area of Makiling Botanic Gardens.

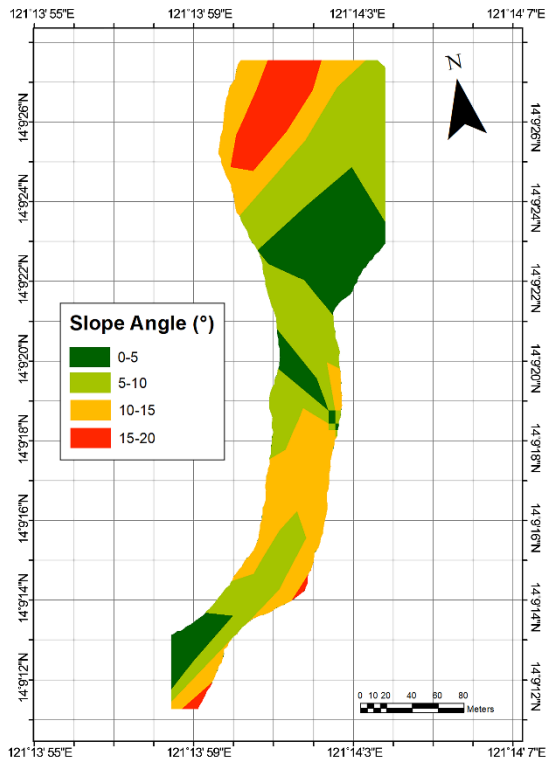


Figure 10. Slope angle map of the recreational area of Makiling Botanic Gardens.

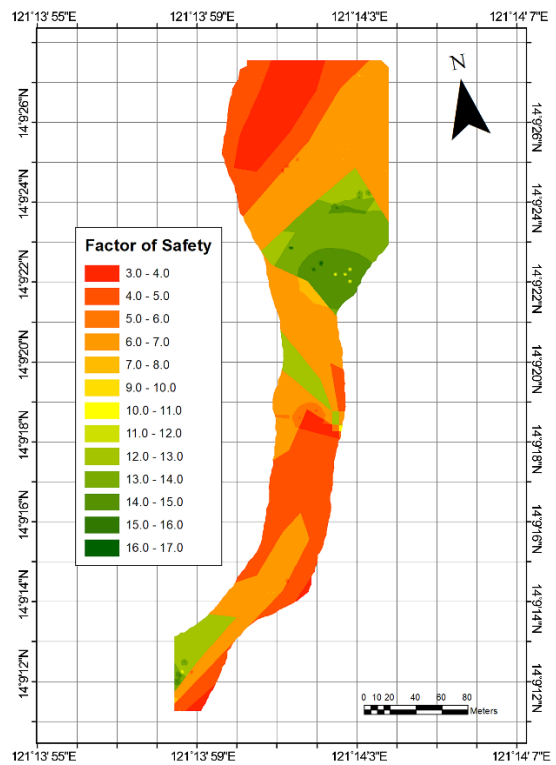


Figure 12. Factor of safety map for the recreational area of Makiling Botanic Gardens.

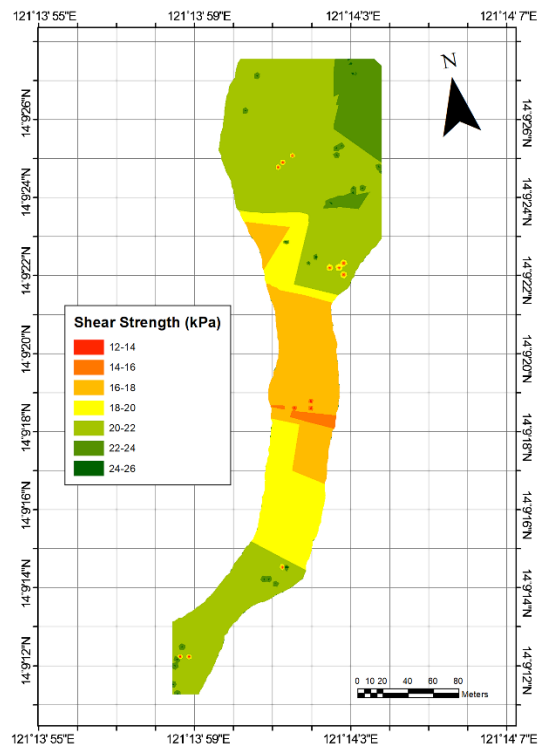


Figure 11. Shear strength map of the recreational area of Makiling Botanic Gardens.

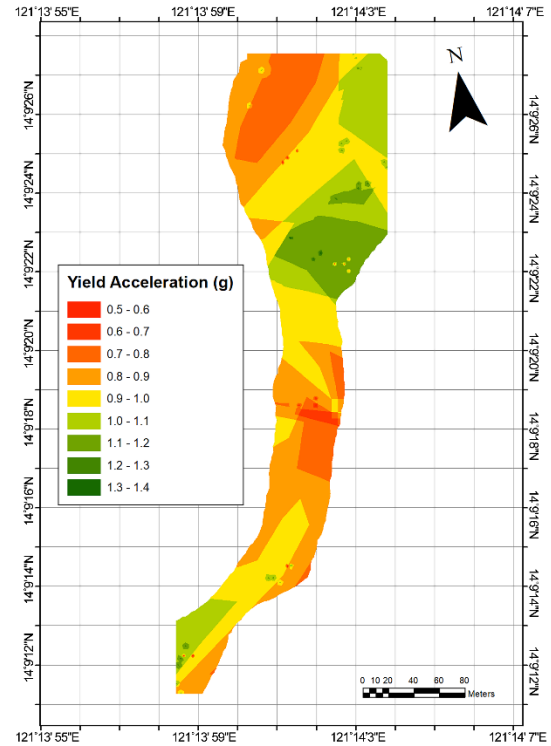


Figure 13. Yield acceleration map of recreational area of Makiling Botanic Gardens.

Relatively, low shear strength is observed in the central portion of the area, while relatively high shear strength is observed in the northern and southern portions of the area. Like the slope angles, the shear strength values served as one of the main parameters of interest to determine the static FS values within the recreational area. Additionally, due to the lack of Standard Penetration Test (SPT) data within the recreational area of MBG, the shear strength map served as the basis for the site classification of the

study area. According to Atkinson (2008), shear strength values of less than 20 kPa and between 20 kPa and 40 kPa indicate the presence of very soft soil and soft soil, respectively. Based on Table 4, site classification E corresponds to an area with a soft soil profile. Thus, the 3.5 ha recreational area of MBG has a site classification of E due to soft soil (Figure 11). The site classification served as a basis for adjusting seismic parameters in the seismic-induced landslide hazard analysis of the said area.

**Table 4.** Soil profile types. Source: National Structural Code of the Philippines 2015.

Soil Profile Type	Generic Description	Average Soil Properties for Top 30 m of Soil Profile		
		Shear Wave Velocity, $V_s$ ( $m \cdot s^{-1}$ )	SPT, N (blows/300 mm)	Undrained Shear Strength, $S_u$ (kPa)
$S_A$	Hard Rock	> 1500		
$S_B$	Rock	760 to 1500		
$S_C$	Very Dense Soil and Soft Rock	360 to 760	> 50	> 100
$S_D$	Stiff Soil Profile	180 to 360	15 to 50	50 to 100
$S_E^1$	Soft Soil Profile	< 180	< 15	< 50
$S_F$	Soil Requiring Site-specific Evaluation			

#### Static Factor of Safety and Yield Acceleration

The entirety of the recreational area of MBG has a static FS of more than 1.5, implying that the area has safe and stable slopes on average since all FS values are higher than unity based on Figure 12. These values further indicate that the resisting forces in the soil and slopes are sufficient to resist the shear forces acting within the soil. Additionally, a conservative value of 1.5 for the FS is generally acceptable as the basis for the design of a stable slope (Das 2010). An FS of less than 1.5 or close to unity may imply that the slope is close to failure with additional loads. Moreover, an FS of precisely unity indicates slope failure when the applied load reaches the design load. At the same time, an FS of less than unity indicates slope failure even without the application of loads. Since the FS of the recreational area of MBG yielded a value of more than 1.5, the area is suitable for some infrastructure developments since loading and environmental conditions are not severe (Maria 2016). Specifically, the minimum FS value yielded by the recreational area is around 3.02, in the Molawin Creek area and some portions of the *Dipterocarp* area. The relatively low FS value yields are consistent with the previously observed trends within the same areas. This implies that these areas have relatively very steep slopes (Figure 9) with large slope angles, which are both factors that affect slope stability. A steeper slope or a higher slope angle would result in a decrease in slope stability, which further increases its susceptibility to landslide occurrence due to the increasing effects of gravity as the plane becomes

more inclined with respect to the horizontal plane. The remaining areas of the recreational area of MBG have relatively flat slopes with small slope angles, resulting in relatively very high FS values compared to the Molawin Creek area and the *Dipterocarp* area. Nevertheless, due to a high FS of 3.02, the Molawin Creek and *Dipterocarp* areas have safe and stable static slopes.

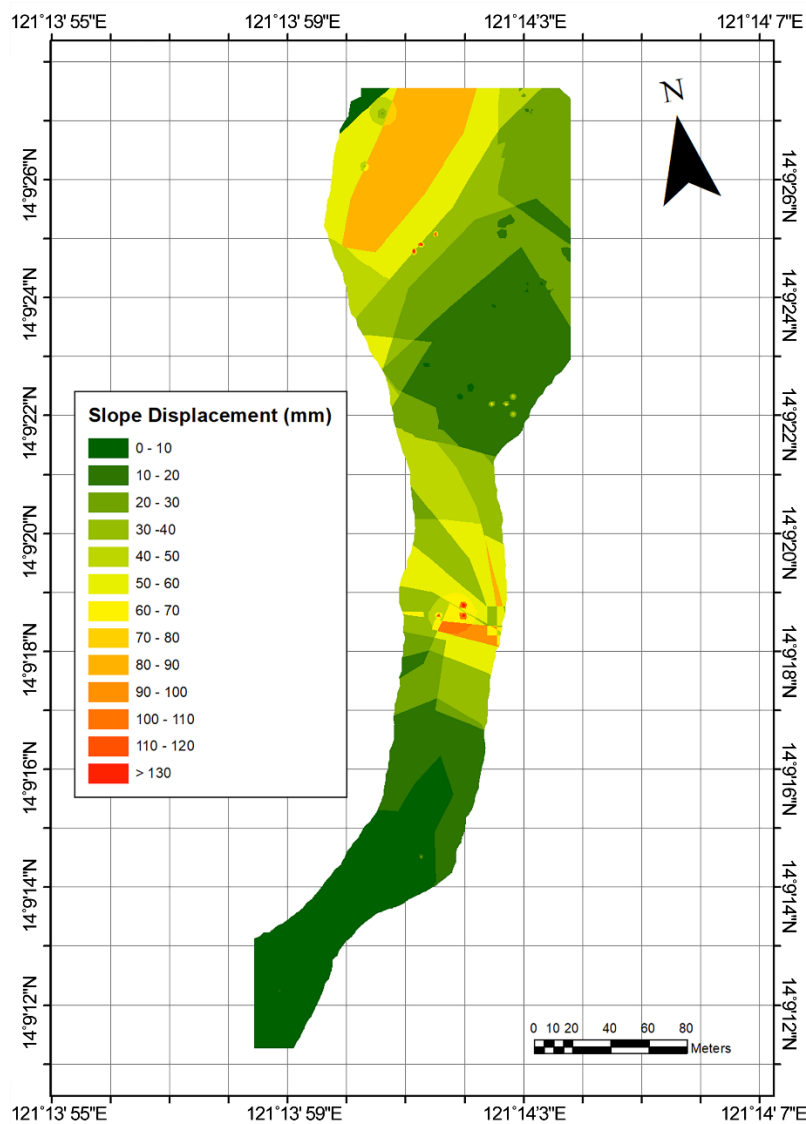
Meanwhile, the entirety of the recreational area has a yield acceleration value ranging from 0.5 g to 1.4 g (Figure 13). The obtained value indicates that a minimum pseudo-static acceleration equal to the acceleration due to gravity multiplied by a factor of 0.5 must be induced to cause slope instability and landslide occurrence in some portions of the recreational area. Furthermore, the Molawin Creek and *Dipterocarp* areas yielded relatively low yield acceleration values, ranging only from 0.5 g to 0.8 g (Figure 12). These areas also yielded the lowest FS values, which served as the main factor for the differences in yield acceleration values at different areas of the recreational area of MBG. Based on the peak ground acceleration (PGA) maps of Peñarubia et al. (2020), the PGA for the area of Los Baños, Laguna, ranges from 0.3 g to 0.5 g considering a return period of 475 years. Following the American Society of Civil Engineers (ASCE) Section 7-05 and consideration of the site classification (E) of the recreational area, an amplification factor was applied to the PGA values, resulting in an amplified PGA for the recreational area of MBG, ranging from 0.70 g to 0.85 g. As a result, in the event of an earthquake, slope instabilities and

movements are highly likely to occur in the Molawin Creek area and *Dipterocarp* area since their respective yield acceleration values are less than or within the range of the amplified PGA for the recreational area of MBG.

**Seismic-Induced Slope Displacements**

Some portions of the recreational area have minimal projected slope displacements in the event of a seismic-induced landslide ranging only from 0 mm to 10 mm (Figure 14). However, some portions of the recreational area have large and significant slope displacements in the event of a seismic-induced landslide (Figures 12 and 13). Specifically, the Molawin Creek area and *Dipterocarp* area yielded the relatively highest projected slope displacements in the

event of a seismic-induced landslide ranging from 80 mm to 130 mm. While the said areas have stable slopes in terms of static factors (Figure 12), the same areas have unstable slopes in terms of dynamic factors (Figure 13) resulting to significant seismic-induced slope displacement values. These dynamic factors generally consider the stability of the slopes against dynamic forces such as seismic-induced forces. Unfortunately, the Molawin Creek and *Dipterocarp* areas yielded relatively low resistance against these forces. As a result, in the event of an earthquake, the slopes within the Molawin Creek and *Dipterocarp* areas are high and likely to be significantly and progressively displaced which could lead to slope failures and seismic-induced landslides.



**Figure 14.** Slope displacement map of the recreational area of Makiling Botanic Gardens.



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## ETHICAL CONSIDERATIONS

Test points used for landslide hazard analysis was safe and practical for researchers to avoid accidents and injuries. Studies conducted in a protected area employed non-intrusive and non-destructive methods to preserve the environment and ecosystem of the said area. Studies conducted during the COVID-19 pandemic observed proper protocol and guidelines imposed by the agencies and institutions involved in the study.

## DECLARATION OF COMPETING INTEREST

The authors declare that there are no competing interests to any authors.

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**ROLE OF AUTHORS:** AMAC – concept, overall method design, data gathering and analysis; writing, revising, and finalizing the manuscript JJCA – advising, method designs, data gathering and analysis, revising the manuscript; ACC – method designs (geotechnical), data gathering and analysis, revising the manuscript; CBB – method designs (seismic), data gathering and analysis, revising the manuscript.

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- j. Citing a chapter in a book – name(s) of author(s), year, full title of the chapter in a book (capitalize each main word), last name of editor and title of book, edition, publisher, place of publication and page range of that chapter:  
Poutiers JM. 1998. Gastropods. In: Carpenter KE and Niem VH (eds). *FAO Species Identification Guide for Fishery Purposes. The Living Marine Resources of the Western Central Pacific Seaweeds, Corals, Bivalves and Gastropods*. Food and Agriculture Organization, Rome, Italy, pp. 364-686.
- k. Citing a Webpage – names of the author (s), year, title of the article, webpage address and date accessed.  
Froese R and Pauly D (eds). 2022. *FishBase*. [www.fishbase.org](http://www.fishbase.org). Accessed on 07 March 2022.  
CITES (Convention on International Trade of Endangered Species). 2014. *The CITES Appendices*. Convention on International Trade in Endangered Species of Wild Flora and Fauna. [www.cites.org](http://www.cites.org). Accessed on 07 March 2022.
- l. Citing a thesis or dissertation – author’s family name, initial names of the author, year, title of the thesis, degree, name of institution, address of the institution, total number of pages (pp).  
Guion SL. 2006. Captive breeding performance of *Crocodylus porosus* (Schneider 1901) breeders at the Palawan Wildlife Rescue and Conservation Center. BS in Fisheries. Western Philippines University-Puerto Princesa Campus, Palawan, Philippines. 28pp. (include weblink, if there’s any, followed by accessed date).  
Lerom RR. 2008. Biosystematics study of Palawan landraces of rice (*Oryza sativa* L.). Doctor of Philosophy, Institute of Biological Sciences, University of the Philippines-Los Baños College, Laguna, Philippines. 197pp. (include weblink, if there’s any, followed by accessed date).
- m. Citing a Report  
Picardal RM and Dolorosa RG. 2014. *Gastropods and bivalves of Tubbataha Reefs Natural Park, Cagayancillo, Palawan, Philippines*. Tubbataha Management Office and Western Philippines University. 25pp. (include weblink, if there’s any, followed by accessed date).
- n. In Press articles when cited must include the name of the journal that has accepted the paper.  
Alcantara LB and Noro T. In press. Growth of the abalone *Haliotis diversicolor* (Reeve) fed with macroalgae in floating net cage and plastic tank. *Aquaculture Research*.
- o. Citing an article from an online newspaper.  
Fabro KA. 2021. Surge in seizures of giant clam shells has Philippine conservationists wary. *Mongabay*. <https://news.mongabay.com/2021/03/surge-in-seizures-of-giant-clam-shells-has-philippine-conservationists-wary/>. Accessed on 07 March 2022.

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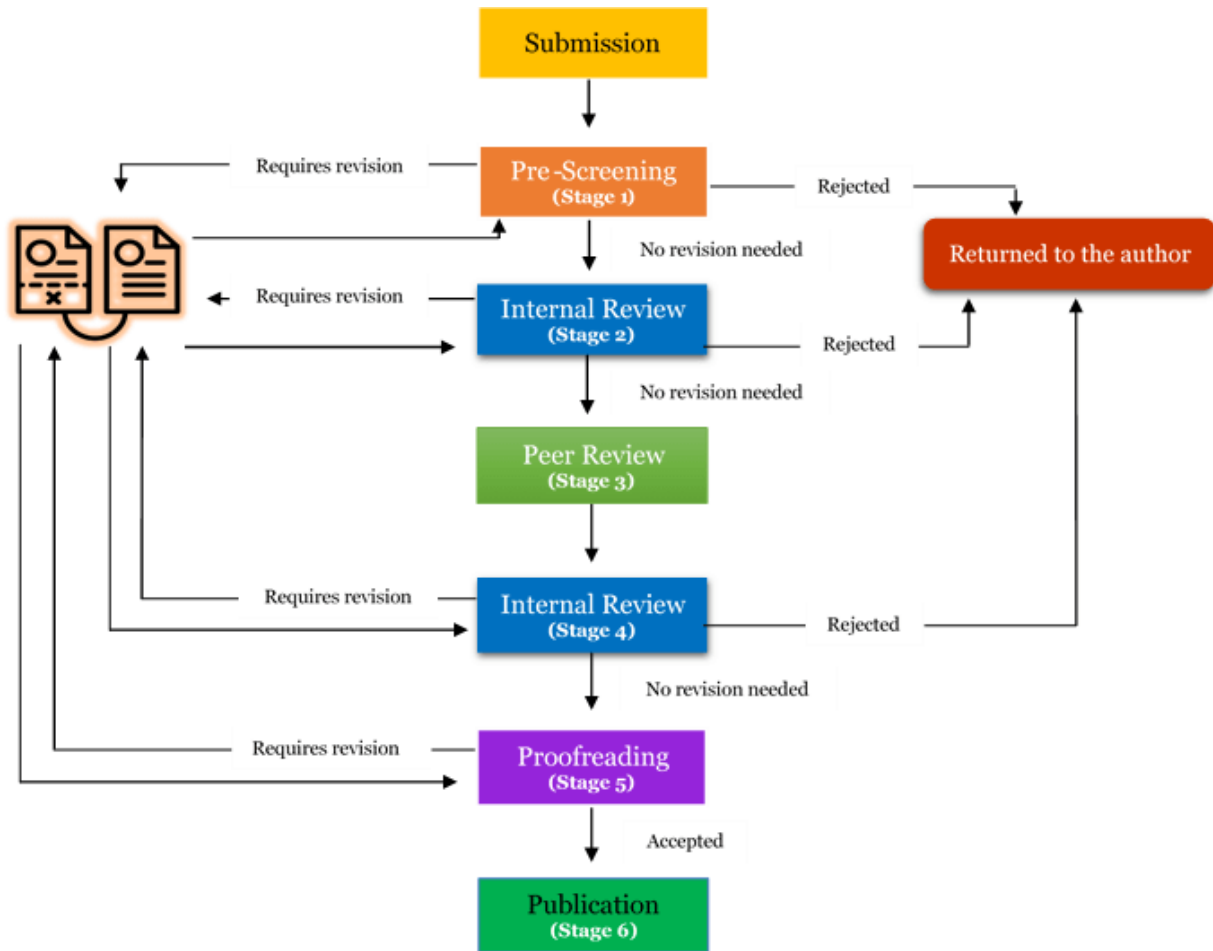
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